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# TASMANIAN SCALEFISH FISHERY ASSESSMENT 2017/18

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This assessment of the Tasmanian Scalefish Fishery is produced by the Institute for Marine and Antarctic Studies (IMAS) using data downloaded from the Department of Primary Industries, Parks, Water and Environment (DPIPWE) Fisheries Integrated Licensing and Management System (FILMS) database. The information presented here includes all logbook returns for 2017/18 season that were entered prior to November 2018.

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# Executive Summary

The Tasmanian Scalefish Fishery is a multi-species fishery operating in State fishing waters and encompassing a wide variety of capture methods and species. The Scalefish Fishery Management Plan (amended in 2015), provides the management framework for the fishery.

## Fishery assessment

Since the early 1990s, annual commercial catches of the major species have generally declined. Total scalefish production (excluding small pelagics and cephalopods) declined from over 1,000 t in the mid-1990s to around 300 t in recent years, due in part to changed targeting practices and market demand, the transfer of the southern shark fishery to the Commonwealth in 2000 and introduction of the Scalefish Fishery Management Plan in 1998.

There has been a continuous decline in the number of vessels participating in the scalefish fishery and in the number of scalefish fishing licences since 2000. Although effort performance indicators were not triggered for any fishing method, for certain species (e.g. Blue Warehou) there are on-going concerns regarding the level of latent capacity within the fishery from licence holders who are currently either participating at low levels or not active (only 20–50% of licences are active depending on the type).

Catch and effort information for the recreational fishery are available periodically and have demonstrated that the recreational catch in recent years represents a significant component of the total harvest for species such as Sand Flathead, Striped Trumpeter, Bastard Trumpeter, Blue Warehou, Flounder, Mullet, Barracouta, Jackass Morwong, Cod, and Leatherjacket.

## Species assessments

For the current report, available fishery and biological information has been assessed against a series of revised reference points that have been applied for several years, and now formally replace out-dated ones proposed in the original (1998) policy document. In assessing the status of each species, the national stock reporting framework (i.e. Sustainable, Recovering, Depleting, Depleted and Undefined) has been adopted. It should be noted that the stock reporting framework adopted here only defines the stock against the limit reference point of whether it is recruitment overfished or not. Target reference points (i.e. those that correspond to levels of biomass and fishing pressure that are considered to provide for optimal harvests) have not been defined at this stage. It should also be noted that Banded Morwong has been considered in a separate assessment report. This change from previous reports reflects differences in the timetable for setting the annual total allowable catch for Banded Morwong (based on quota year) compared with routine assessment reporting for the other scalefish species (based on financial year).

## Species assessments

Species/Species group	Status	Comment	Page
<b>Australian Salmon</b> <i>Arripis trutta</i>	<b>SUSTAINABLE</b>	Species has a long history of exploitation across south-eastern Australian. Low commercial landings in Tasmania in recent years are driven by market demand not abundance. The current level of fishing pressure in Tasmania is well below historical levels and unlikely to cause the biological stock to become recruitment impaired.	38
<b>Australian Sardine</b> <i>Sardinops sagax</i>	<b>SUSTAINABLE</b>	The fishery is in a developmental phase in Tasmania, with low catches reported to date. The species is classified as not overfished nor subject to overfishing by ABARES for 2017 (Patterson et al. 2018), while all Australian stocks are similarly classified as sustainable in the 2016 Status of Australian Fish Stocks. The current level of fishing pressure in Tasmania is unlikely to cause the biological stock to become recruitment impaired.	44
<b>Barracouta</b> <i>Thyrsites atun</i>	<b>UNDEFINED</b>	Catches of Barracouta have declined steadily since the mid-2000s due to a decrease in targeted effort as a result of a lack of market demand. Catches and catch rates are not considered indicative of stock status and there is insufficient information to confidently classify the stock.	52
<b>Bastard Trumpeter</b> <i>Latridopsis forsteri</i>	<b>DEPLETED</b>	Trends in commercial catch suggest record low population levels and that the species is recruitment overfished. The current minimum legal size limit is below the size of maturity. Although commercial catches have remained low for the past decade, fishing pressure may be too high to allow stocks to recover.	58
<b>Blue Warehou</b> <i>Seriola lalandi</i>	<b>DEPLETED</b>	This is predominately a Commonwealth-managed species that is classified as overfished by ABARES in 2017 (Patterson et al. 2018). It is sporadically abundant in Tasmanian waters. Despite a reduction in Total Allowable Catch (TAC) for the Commonwealth fishery to 118 t and the initiation of a rebuilding strategy in 2008, stocks have shown little evidence of recovery.	65
<b>Tiger Flathead</b> <i>Platycephalus richardsoni</i>	<b>SUSTAINABLE</b>	This is a predominately Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES in 2017 (Patterson et al. 2018). In Tasmania, Tiger Flathead are caught predominately by the commercial sector and in recent years and while catches have fluctuated between 20–60 tonnes in recent years they have typically represented under 5% of Commonwealth trawl landings.	72

<p><b>Sand Flathead</b> <i>Platycephalus bassensis</i></p>	<p><b>DEPLETING</b></p>	<p>Recreational catches of Sand Flathead dominate landings of this species in Tasmania. Fishery independent surveys suggest relatively low abundances of legal sized fish, particularly in south-eastern Tasmania, with populations subject to heavy fishing pressure. While a recent increase in minimum size limit and reduction in bag limit should benefit stocks, current levels of fishing pressure are likely to cause the stock to become recruitment impaired.</p>	<p>79</p>
<p><b>Flounder</b> Pleuronectidae family</p>	<p><b>UNDEFINED</b></p>	<p>Greenback Flounder (<i>Rhombosolea tapirina</i>) constitute the majority of the commercial catch, which remains low due to the ban on overnight gillnetting and limited market demand. Catch and catch rates are considered unreliable estimators of abundance and the status of the stock remains uncertain.</p>	<p>91</p>
<p><b>Gould's Squid</b> <i>Nototodarus gouldi</i></p>	<p><b>SUSTAINABLE</b></p>	<p>This is a predominately Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES in 2017 (Patterson et al. 2018). Dual-licensed vessels fish in Tasmanian waters especially in years of peak abundance. The species is characterised by high inter-annual variability in abundance in State waters with low catches in recent years.</p>	<p>97</p>
<p><b>Jack Mackerel</b> <i>Trachurus declivis</i></p>	<p><b>SUSTAINABLE</b></p>	<p>This is a predominately Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES for 2017 (Patterson et al. 2018). Only minor catches of this species have been taken from Tasmanian waters in recent years due to one operator leaving the fishery, and thus patterns of catch and effort do not reflect the stock status. The current level of fishing pressure in Tasmania is unlikely to cause the stock to become recruitment impaired.</p>	<p>103</p>
<p><b>Jackass Morwong</b> <i>Nemadactylus macropterus</i></p>	<p><b>SUSTAINABLE</b></p>	<p>This is a Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES for 2017 (Patterson et al. 2018), with very low commercial catches in Tasmania.</p>	<p>109</p>
<p><b>Leatherjacket</b> <i>Monacanthidae</i> family</p>	<p><b>UNDEFINED</b></p>	<p>A number of Leatherjacket species are found inshore around Tasmania's coastline and are not differentiated in logbooks. Leatherjackets are a by-product species and are not actively targeted due to a lack of market demand. Catch is therefore not a good indicator of abundance and there is little biological information to confidently classify the status of Leatherjacket stocks</p>	<p>115</p>
<p><b>Longsnout Boarfish</b> <i>Pentaceropsis recurvirostris</i></p>	<p><b>UNDEFINED</b></p>	<p>Boarfish are a by-product species of Banded Morwong fishing with low catches due to the large minimum legal size. There</p>	<p>120</p>

		is insufficient information available to confidently classify this stock.	
<b>Yelloweye Mullet</b> <i>Aldrichetta forsteri</i>	<b>SUSTAINABLE</b>	Catches are at low levels but this is unlikely to be indicative of abundance. Yelloweye Mullet are most abundant in estuarine habitats, where netting is prohibited or restricted, thereby providing a high degree of protection throughout most of their range. It is therefore unlikely that the stock is recruitment impaired or that current fishing pressure is too high as to cause the stock to become recruitment impaired.	125
<b>Snook</b> <i>Sphyraena novaehollandiae</i>	<b>SUSTAINABLE</b>	Snook catches declined slightly in 2017/18 relative to 2016/17 but were consistent with long-term averages. Catch rates are an unreliable estimate of abundance due to the species not being actively targeted. Recent biological analyses indicate that the current level of fishing mortality is unlikely to cause the stock to become recruitment impaired.	131
<b>Eastern School Whiting</b> <i>Sillago flindersi</i>	<b>SUSTAINABLE</b>	This is a predominately Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES for 2017 (Patterson et al. 2018). Tasmanian catches fluctuate due to market demand and species targeting (Tiger Flathead or Eastern School Whiting) by the primary operator. Tasmanian commercial catches represent a small fraction of the Commonwealth commercial catch.	136
<b>Southern Calamari</b> <i>Sepioteuthis australis</i>	<b>DEPLETING</b>	State-wide commercial catches in 2017/18 declined by more than 50% from 2016/17 when record high levels were reported, largely due to a decline in catch in the northern areas of the State. Fishing effort similarly declined, particularly on the north coast, though remained high relative to historic levels. Catch rates fell markedly along the north coast and in Great Oyster Bay relative to 2016/17, though were relatively stable for Mercury Passage. Preliminary information from in-water surveys suggest low levels of egg-laying activity during the peak spawning season in late 2017, implying low abundance of adults. Catches in 2015/16 and 2016/17 were above estimates of MSY. While there is uncertainty as to the current status of the biomass, recent high catches and effort, particularly off the north coast, suggest that fishing mortality has been excessive and is likely to cause the stock to become recruitment impaired.	142
<b>Southern Garfish</b> <i>Hyporhamphus melanochir</i>	<b>DEPLETED</b>	After a strong decline in catches in 2006/07 and 2007/08 coupled with changes in population structure, which prompted management actions, this species showed signs of recovery. However, over the last five fishing seasons there have been	150

		significant reductions in total catch and in catch rates, particularly beach seine (the main catching method), with both total catch and beach seine catch rates falling to record low levels. Despite declining catches, catch curve analysis indicates fishing mortality and exploitation rates to be higher in 2017 and 2018 than the late 2000s, when the stock was considered to be in a depleted state. These results suggest current low levels of fishing pressure may be too high to allow stock to recover.	
<b>Striped Trumpeter</b> <i>Latris lineata</i>	<b>RECOVERING</b>	After a lack of recruitment for over a decade, and limited biological information from which to assess stock status in recent years, Striped Trumpeter is showing signs of recovery in the last two years, with evidence for recent recruitment.	160
<b>Wrasse</b> <i>Notolabrus tetricus</i> (Bluethroat Wrasse) <i>Notolabrus fuciola</i> (Purple Wrasse)	<b>SUSTAINABLE</b>	Catches, effort and catch rates have remained relatively stable for almost a decade providing no indication that recent fishing mortality is too high. There is, however, some uncertainty over the size of the catch taken by rock lobster fishers and used for bait.	174

# 1. Introduction

## The Tasmanian Scalefish Fishery

The Tasmanian Scalefish Fishery is a multi-gear and multi-species fishery. The main gear types include gillnet, hooks and seine nets, harvesting a diverse range of scalefish, shark and cephalopod species. Other fishing gears in use include traps, Danish seine, dip nets and spears. A listing of common and scientific names of species reported in catches is presented in Appendix 1.

In many respects the scalefish fishery is dynamic, with fishers readily adapting and changing their operations in response to changes in fish availability and in response to market requirements and opportunities. As a consequence, only a small proportion of the fleet has specialised in a single activity or targeting a primary species (Ziegler 2012). For many operators, scalefish represent an adjunct to other activities, for instance rock lobster fishing.

## Management objectives and strategies

The Scalefish Fishery Management Plan [*Fisheries (Scalefish) Rules 1998*] was first introduced in 1998 (DPIF 1998) and was reviewed in 2001, 2004, 2009 and most recently in 2015. The management plan provides the regulatory framework for the fishery, which covers commercial and recreational components. While the plan contains the overarching legislation under which the fishery operates, the following objectives, strategies and performance indicators are contained in a policy document currently under review.

### **Major objectives**

- To maintain fish stocks at sustainable levels by restricting the level of fishing effort directed at scalefish, including the amount and types of gear that can be used;
- To optimise yield and/or value per recruit;
- To mitigate any adverse interactions that result from competition between different fishing methods or sectors for access to shared fish stocks and/or fishing grounds;
- To maintain or provide reasonable access to fish stocks for non-commercial fishers;
- To minimise the environmental impact of scalefish fishing methods generally, and particularly in areas of special ecological significance;
- To reduce by-catch of juveniles and non-target species; and
- To implement effective and efficient management.

### **Primary strategies**

- Limit total commercial fishing capacity by restricting the number of licences available to operate in the fishery;
- Define allowable fishing methods and amounts of gear that can be used in the scalefish fishery by both commercial and non-commercial fishers;
- Monitor the performance of the fishery over time, including identification and use of biological reference points (or limits) for key scalefish species;
- Protect fish nursery areas in recognised inshore and estuarine habitats by prohibiting or restricting fishing in these areas;
- Employ measures to reduce the catch and mortality of non-target or undersized fish; and
- Manage developing fisheries under permit conditions.

## **This report**

This report covers the assessment of selected scalefish and cephalopod fisheries under Tasmanian jurisdiction, as well as selected species under Commonwealth jurisdiction that are also taken in state waters by state-licensed operators. Formal assessments of species under Commonwealth jurisdiction (e.g. Tiger Flathead, Blue Warehou, Jackass Morwong, Ocean Perch, School Whiting, Blue-eye Trevalla, Blue Grenadier, School and Gummy Shark) are undertaken by the Southern and Eastern Scalefish and Shark Fishery Assessment Group (SESSFAG, e.g. Morison et al. 2012) and are summarised in fishery status reports produced by the Australian Bureau of Agricultural and Resource Economics and Sciences (ABARES; e.g. Patterson et al. 2015, 2016, 2017, 2018).

## **Data sources**

Commercial catch and effort data are collected through compulsory Tasmanian Commercial Catch, Effort and Disposal Returns, and Commonwealth non-trawl (GN01 and GN01A) and Southern Squid-jig Fishery (SSJF) logbook returns. Unless noted otherwise, catch and effort data reported in this assessment relate to the commercial sector. Catch and effort information for the recreational sector are collected from surveys that are conducted periodically.

### **Tasmanian General Fishing Returns**

The catch and effort logbooks have been amended several times (1995, 1999, 2007, 2010 and 2013) in an effort to report at finer spatial scales and provide greater operational detail. While the offshore fishing blocks are still at the 30nm (1/2 degree) spatial resolution, the logbooks introduced in 2010 have redefined the scale of the coastal blocks (Fig. 1.1). In analysing catch and effort information some data quality control has been undertaken, details of which are provided in Appendix 2.

### **Commonwealth catch returns**

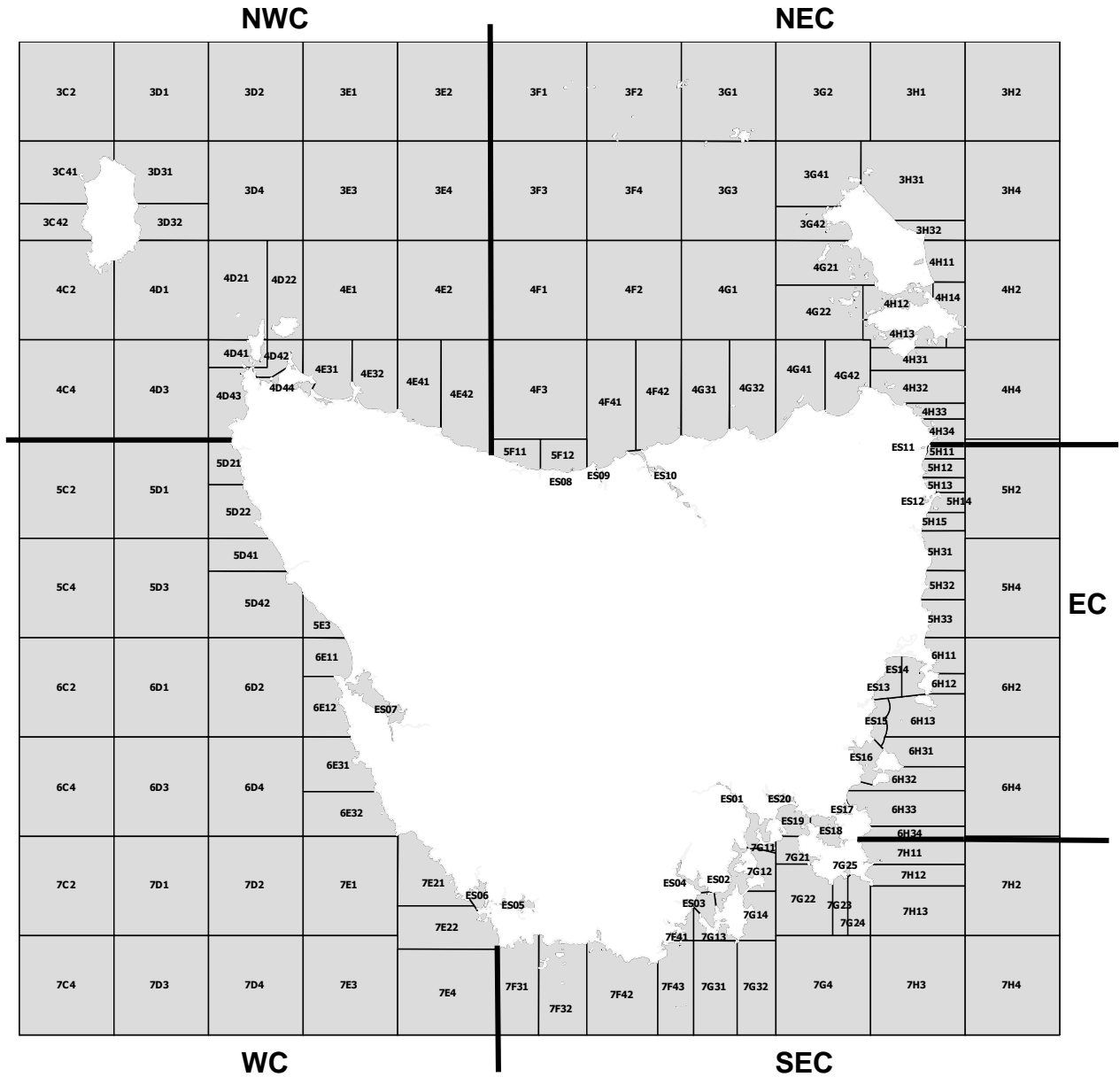
Following the introduction of the Commonwealth non-trawl logbook (GN01 and subsequent versions) in late 1997, dual endorsed Tasmanian and Commonwealth (South East Non-Trawl and Southern Shark) operators generally commenced recording all of their catch and effort data, including fishing in State waters, in the Commonwealth logbooks. In addition, several dual endorsed squid operators reported some or all of their state waters fishing activity in the Southern Squid-jig Fishery (SSJF) logbook. As most of these operators did not explicitly indicate whether fishing occurred in State or Commonwealth waters, it has been necessary to incorporate all activity reported from coastal fishing blocks in the analyses. For details of data restrictions and quality control involving Commonwealth logbook data see Appendix 2.

During 2001, dual endorsed fishers were instructed to report all fishing activities under State jurisdiction in the Tasmanian catch and effort logbooks. This should have removed the necessity to include subsequent Commonwealth catch and effort data into analyses, but it has become apparent that there was some confusion amongst fishers about reporting requirements. For example, catches of species such as Striped Trumpeter taken by Commonwealth operators were not routinely reported in the Tasmanian catch returns. Commonwealth logbook data since 2001 have been available for the current assessment and have been checked for possible double reporting (*i.e.* on both the Tasmanian and Commonwealth catch returns) and where this was not the case, the catch and effort database used in this assessment was updated.

### **Data analysis**

For the purposes of this assessment, effort and catch rate analyses are restricted to commercial data provided for the period 1<sup>st</sup> July 1995 to 30<sup>th</sup> June 2018.

A fishing year from 1<sup>st</sup> July to 30<sup>th</sup> June in the following year has been adopted for annual reporting. This better reflects the seasonality of the fisheries for most species rather than reporting based on calendar year, with catch (and effort) generally concentrated between late spring and early autumn. In addition, it better encompasses the biological processes of recruitment and growth for most species. Unless otherwise stated, data have been analysed at state-wide and regional levels. Five broad assessment regions are used: southeast coast (SEC), east coast (EC), northeast coast including Flinders Island (NEC), northwest coast including King Island (NWC), and west coast (WC) (Fig. 1.1).



**Figure 1.1** Map of Tasmania with the fishing blocks and the assessment regions. SEC = southeast coast, EC = east coast, NEC = northeast coast, NWC = northwest coast, and WC = west coast.

There are 14 main fishing methods used in the Tasmanian Scalefish Fishery. Catch and effort by gear types are presented in Table A2 in Appendix 4. For assessment purposes, effort has been primarily expressed in terms of number of days fished using the specified gear type, irrespective of the amount of gear utilised each day. Since a variety of gear types are represented, in some cases it has been necessary to additionally express effort in gear units appropriate to each specific fishing method (Table 1.1). Although days fished represents a less sensitive measure of effort, it has become apparent that some fishers have misinterpreted reporting requirements for effort. Attempts have been made to reduce this problem by updating the logbook; however, confusion about the new reporting requirements may have biased some effort measures. Examining effort in terms of days fished overcomes any uncertainty about the reporting of effort units and provides consistency assuming there have been no major changes to fishing practices throughout the time series.

**Table 1.1** Table of effort gear units by fishing method.

Method	Effort gear units
Beach seine/purse seine	No. of shots
Graball/small mesh net	100 m net hours
Drop-line	100 hook lifts
Hand-line	Line hours
Fish trap	No. trap or pot lifts
Squid-jig	Jig hours
Spear	Fisher hours
Dip-net	Dip-net hours

Catch returns for which effort information was incomplete or unrealistically high or low (either due to data entry error or misinterpretation of information requirements by fishers), were flagged and excluded when calculating effort based on gear units or catch rates based on catch per unit of gear. Only a small number of fishing records for 2017/18 needed to be excluded in this manner. All records were, however, included for reporting catch, days fished and catch per day.

Since catch rate data are typically log-normally distributed, the geometric rather than arithmetic mean of all valid individual daily catch records has been calculated when generating catch rate statistics. The geometric mean is calculated as the  $n^{\text{th}}$  root of the product of the individual rates ( $y_i$ ):

$$GM_{\bar{y}} = \sqrt[n]{\prod y_i}$$

This is equivalent to computing the arithmetic mean of the natural logarithm of each number, and then taking the exponent:

$$GM_{\bar{y}} = \exp \left[ \frac{1}{n} (\sum \ln(y_n)) \right]$$

It should be noted that catch rates calculated in this manner may differ slightly from the more simplistic approach of dividing total catch by total effort or using the arithmetic mean. The

geometric mean has the advantage of being less affected by the few observations that are skewed very high, as often happens with log-normally distributed data.

### **Recreational fishery**

Information on recreational fisheries in Tasmania is relatively sparse in comparison to commercial data. Detailed analyses of the Tasmanian recreational fishery available are based on the 2000/01 National Survey (Lyle 2005), the 2007/08 state-wide fishing survey (Lyle et al. 2009) and the 2012/13 state-wide survey (Lyle et al. 2014a). Additional data are provided by targeted surveys of the offshore recreational fishery (Tracey et al. 2013), recreational gillnet fishery (Lyle and Tracey 2012) and fishing practices (Lyle et al. 2014b), along with recreational net licence numbers.

### **Assessment categories**

Assessment categories take into account the species' importance and the quantity of data available to inform the appropriate reference points for each species, with the intention to assign a stock status to each species based on the presented criteria.

### **Species' importance**

There are over 90 scalefish species reported in commercial catches in Tasmania. Harvested species have variable social and economic values, meaning that all species cannot be given the same priority for assessment. In assessing the importance of a species, a combination of factors was taken into account, including:

- Whether the species is a target, secondary target or by-product;
- The economic importance of the species;
- The annual landings for the species (i.e. annual catch > 5 t for 50% of the time between 1995 and present);
- The number of operators targeting the species and;
- The "conservation" value of the species.

Species are consequently classified as "Key species" or "Minor species". The remaining species reported in commercial catches are considered of minor significance and not at threat from current fishing practices. As such, these species are not included in the assessment.

### **Reporting levels definitions**

Each species in the assessment is associated with a level of reporting (Full, Medium or Minor), which stems from the available data for the species/species group. The aim is, in time, to have full reporting for all key species and medium reporting for all minor species. The attributes of the reporting categories are defined in Table 1.2. Table 1.3 summarises the species selected for the assessment, their importance and level of reporting.

**Table 1.2.** Summary of the attribute for the reporting categories.

Attribute	Reporting level		
	Full	Medium	Minor
Time series estimate of biomass from dynamic models	•		
Time series estimate of total, natural and fishing mortality from dynamic models	•		
Quantitative risk analysis of future harvesting using dynamic models	•		
Time series of age and/or length composition data	•		
Estimates of total, natural and fishing mortality (from catch curves)	•		
Local (TAS) information for growth, mortality, selectivity and maturity	•	•	
Representative time-series of commercial catch	•	•	•
Single biological species or stock	•	•	•
Sporadic age and/or length composition data		•	
Non-local (non-TAS) information for growth, mortality, selectivity and maturity		•	•
Complex of related species		•	•

**Table 1.3.** Summary of importance and reporting level for all retained species.



Species/Species group	Importance	Reporting level
Banded Morwong <sup>1</sup>	Key	Full
Australian Salmon	Key	Medium
Bastard Trumpeter	Key	Medium
Blue Warehou	Key	Medium
Flathead	Key	Medium
Southern Calamari	Key	Medium
Southern Garfish	Key	Medium
Striped Trumpeter	Key	Medium
Wrasse	Key	Medium
Barracouta	Minor	Minor
Flounder	Minor	Minor
Gould's Squid	Minor	Minor
Jack Mackerel	Minor	Minor
Jackass Morwong	Minor	Minor
Leatherjacket	Minor	Minor
Longsnout Boarfish	Minor	Minor
Mullet	Minor	Minor
Pike	Minor	Minor
School Whiting	Minor	Minor
Australian Sardine	Developmental	Minor

<sup>1</sup> Note Banded Morwong are considered in a separate, stand-alone report.

### Stock status definitions

In order to assess species in a manner consistent with the national approach (and other jurisdictions) we have adopted the national stock status categories used in the 2018 Status of Australian Fish Stocks (SAFS) reporting (Table 1.4). These categories define the assessed state of the stock in terms of recruitment impairment, which is often treated as a limit reference point. Recruitment impairment occurs when the mature adult (spawning biomass) population is depleted to a level where it no longer has the reproductive capacity to replenish itself. Hence, recruitment impaired stocks have not necessarily collapsed but they do have reduced productivity. Fisheries are ideally also managed towards target reference points that aim to maximise benefits from the harvesting, such as economic yield or provision of food. The scheme used here does not attempt to assess the fishery against any target outcomes.

**Table 1.4.** The stock status classifications that were adopted for this assessment.

Stock status	Description	Potential implications for management of the stock
<b>SUSTAINABLE</b>	Biomass (or proxy) is at a level sufficient to ensure that, on average, future levels of recruitment are adequate (recruitment is not impaired) and for which fishing mortality (or proxy) is adequately controlled to avoid the stock becoming recruitment impaired.	Appropriate management is in place.
<b>RECOVERING</b>  	Biomass (or proxy) is depleted and recruitment is impaired, but management measures are in place to promote stock recovery, and recovery is occurring.	Appropriate management is in place, and there is evidence that the biomass is recovering.
<b>DEPLETING</b>  	Biomass (or proxy) is not yet depleted and recruitment is not yet impaired, but fishing mortality (or proxy) is too high (overfishing is occurring) and moving the stock in the direction of becoming recruitment impaired.	Management is needed to reduce fishing mortality and ensure that the biomass does not become depleted.
<b>DEPLETED</b>	Biomass (or proxy) has been reduced through catch and/or non-fishing effects, such that recruitment is impaired. Current management is not adequate to recover the stock, or adequate management measures have been put in place but have not yet resulted in measurable improvements.	Management is needed to recover this stock; if adequate management measures are already in place, more time may be required for them to take effect.
<b>UNDEFINED</b>	Not enough information exists to determine stock status.	Data required to assess stock status are needed.

## Performance indicators and reference points definitions

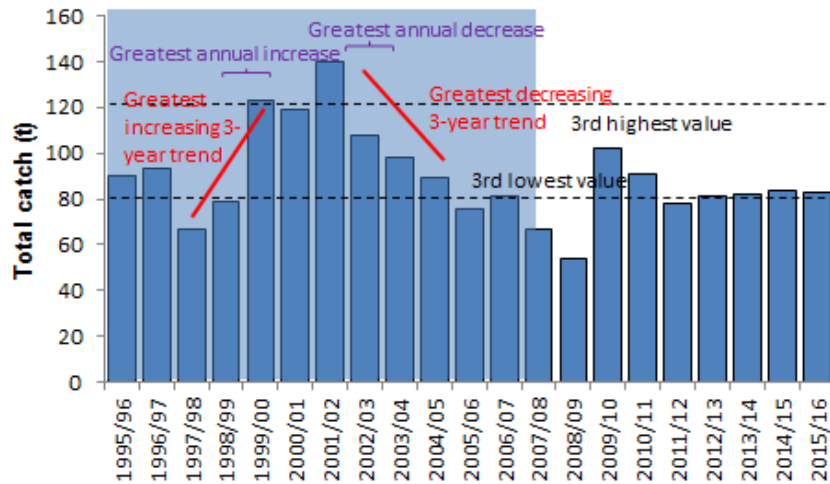
The determination of stock status is based on the consideration of model outputs (for species with full reporting) and commercial catch and effort data, which are assessed by calculating fishery performance indicators and comparing them with reference points. For the current assessment, the decision has been taken to only present performance indicators and reference points recommended in recent assessment reports as alternatives to those originally proposed in the Scalefish Fishery Management Plan policy document (DPIF 1998).

The performance indicators are biomass and fishing mortality. For medium and minor reporting, proxies (commercial catch and CPUE) are used instead as there are insufficient data to calculate biomass or fishing mortality. These are compared to a reference period (1995/96<sup>2</sup> to 2006/07 unless stated otherwise) for each species. The reference points for more generic and full reporting are species-specific while the reference points for medium and minor reporting are applicable for all species. Reference points are illustrated in Fig. 1.2.

**Table 1.5** Summary of the performance indicators and reference points for each reporting standard.

Reporting	Performance indicators	Reference points	
Full	Fishing mortality	<ul style="list-style-type: none"> <li>Maintain an appropriate spatial distribution of catch</li> </ul>	
	Biomass	<ul style="list-style-type: none"> <li>High probability of staying above a certain level of mature biomass</li> <li>High probability of staying above a certain CPUE</li> </ul>	
Medium	Fishing mortality	<ul style="list-style-type: none"> <li>Catch &gt; 3<sup>rd</sup> highest catch value from the reference period</li> <li>Catch &lt; 3<sup>rd</sup> lowest catch value from the reference period</li> <li>Catch variation from the previous year above the greatest inter-annual increase from the reference period</li> <li>Catch variation from the previous year above the greatest inter-annual decrease from the reference period</li> <li>Latest recreational catch estimate &gt; recreational catch estimate from the reference period</li> <li>Proportion of recreational catch to total catch &gt; previous proportion estimate</li> </ul>	
		Biomass	<ul style="list-style-type: none"> <li>CPUE &lt; 3<sup>rd</sup> lowest CPUE value from the reference period</li> <li>Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period</li> </ul>
Minor	Fishing mortality	<ul style="list-style-type: none"> <li>Catch &gt; 3<sup>rd</sup> highest catch value from the reference period</li> <li>Catch &lt; 3<sup>rd</sup> lowest catch value from the reference period</li> <li>Latest recreational catch estimate &gt; recreational catch estimate from the reference period</li> <li>Proportion of recreational catch to total catch &gt; previous proportion estimate</li> </ul>	
		Biomass	<ul style="list-style-type: none"> <li>Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period</li> </ul>

<sup>2</sup> 1995/96 to 2006/07 was selected as the reference period, corresponding to the first twelve years since the introduction of daily catch and effort reporting in the Tasmanian General Fishing Returns.



**Figure 1.2** Schematic illustration of the application of fishing mortality reference points for medium and minor reporting based on trends in catch. The shaded blue area represents the reference period. Note this figure is for illustrative purposes only.

Other measures are also taken into consideration in the determination of stock status, such as changes in biological characteristics of the stock, indicators of stock stress, significant changes in the catch of non-commercial species or undersize commercial-species, and significant external factors related to fishing activity.

Reporting	Performance indicators	Reference points
All	Stock stress	<ul style="list-style-type: none"> <li>• Significant change in the size/age composition of commercial catches</li> <li>• Significant numbers of unhealthy fish landed</li> </ul>

It should be noted that the assessment for Banded Morwong (*Cheilodactylus spectabilis*) for 2017/18 has been considered in a separate, stand-alone assessment report (Moore et al. 2018). This change from previous assessment reports reflects differences in the timetable for setting the annual total allowable catch for Banded Morwong (based on quota year) compared with routine assessment reporting for the other scalefish species (based on financial year).

For the purpose of analysis, shark net and bottom longline catch and effort have been excluded since these methods relate specifically to the School and Gummy Shark fishery, which is managed by the Commonwealth.

### Uncertainties and implications for management

While considerable attention has been directed at ensuring comparability of commercial data over time (refer Appendix 2), it is acknowledged that some recent administrative changes relating to the reporting of catches may have, nonetheless, exerted some influence on observed catch and effort trends.

Other uncertainties in this assessment relate to limitations in catch and effort data, mainly in terms of the level of detail provided and the lack of independent verification. Since the Commercial Catch, Effort and Disposal logbook (formerly the General Fishing Return) was

designed to accommodate a diverse range of fishing activities, compromises have been necessary, with data collection on a daily rather than operational (set or shot) basis.

In the past some fishers experienced problems in correctly interpreting or complying with reporting requirements, especially in terms of how effort information is reported. The introduction of new logbooks during the 2007/08 season has helped to clarify reporting, but there continues to be an on-going need to educate fishers in this regard. Further, the lack of catch verification remains an issue in relation to data quality.

Catch and effort are influenced by a combination of factors which include fishers matching their fishing operations to changing market requirements and/or resource availability, as well as responses to changing management arrangements. The latter adds further uncertainty regarding the underlying causes of any observed trends in catch and effort. There is, therefore, a need to take account of industry perceptions and information when interpreting fishery dependent information.

Limited information about the recreational fishery remains a major uncertainty and is especially significant in the scalefish assessment given the scale of recreational catches relative to commercial catches for some species. The 2000/01, 2007/08 and 2012/13 surveys provide critical information about this sector. A new state-wide survey is being undertaken during 2017/18 and the information obtained is expected to contribute to the 2019/20 assessment report.

Fish mortality due to disease, predation and fishery interactions with Australian and New Zealand fur seals is largely unknown but represents another source of uncertainty. Seals can cause substantial mortality to some of the fish species assessed in this report as well as causing gear damage and influencing fishers' behaviour, factors that impact on catches and catch rates. This tends to be caused predominantly by individual 'rogue' seals which learn to target particular fisheries or fishing methods (e.g. gillnetting), while the typical diet of seals includes mainly pelagic fish species (Goldsworthy et al. 2003).

## **Ecological Risk Assessment**

The current assessment incorporates an evaluation of the potential ecological risks posed by the Tasmanian Scalefish Fishery. Two recognised Ecological Risk Assessment (ERA) frameworks were used in this process: the first involved a qualitative approach suited to fisheries with limited data and is closely aligned with the standard risk assessment approach utilised in occupational health and safety; the second was a semi-quantitative approach that is suited to fisheries for which data relating to catch, discards, post release survival and technical aspects of the fishery are available. Risk analysis considers the source of risk, the possible consequences of the risk and how likely it is that the consequences will occur. Consequences and likelihood are assessed against specific objectives, which differ according to the component of the risk assessment. Consequence and likelihood are combined to produce an estimated level of risk associated with the particular hazardous event in question. The ERA was conducted as a snapshot in time, capturing the risk profile of the fishery in 2012/13; full details are provided in Bell et al. (2016).

## 2. General fishing trends

### Commercial fishing licences

Since 2001 the number of general Scalefish licences has gradually declined from almost 450 to about 280 in 2017, mainly driven by a two-thirds reduction in the number of Scalefish C licences issued (Table 2.1). Up until 2014 between 40–50% of total licence numbers were active each year. In 2015 there was a sharp increase in the number of active licences, in particular amongst Scalefish B and C licences, with activity reported by 195 or 69% of scalefish licences, up from 123 or 43% in 2014. This increase was mainly due to a rise in the number of new entrants activating existing licences along with several rock lobster fishers diversifying their operations, particularly among lease holders.

In addition to the general Scalefish licences, separate fishing licences allow the use of beach seine (a total of 50 licences in two categories, A and B), small mesh gillnet (10 licences), purse seine and Danish seine nets. Fishers with a rock lobster licence (but without Scalefish A or B licence) are also allowed to take scalefish with a limited amount of fishing gear.

**Table 2.1** Numbers of scalefish licences (FL) (total and active) by type (A, B or C) since 2001 (licence years run from March to February; 2017 is the last complete year covered by this assessment). na = not available at the time of writing.

Type	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015	2016	2017
<b>Licence numbers</b>																	
FLA	67	70	70	70	66	66	66	66	65	65	65	65	65	63	63	63	64
FLB	165	164	165	165	162	162	160	159	159	158	158	154	154	151	149	147	147
FLC	214	205	185	173	152	137	129	120	112	92	86	81	77	72	69	65	65
<b>Total</b>	<b>446</b>	<b>439</b>	<b>420</b>	<b>408</b>	<b>380</b>	<b>365</b>	<b>355</b>	<b>345</b>	<b>336</b>	<b>315</b>	<b>309</b>	<b>300</b>	<b>296</b>	<b>286</b>	<b>281</b>	<b>275</b>	<b>276</b>
<b>Active licences</b>																	
FLA	44	51	48	46	38	43	47	37	33	32	35	28	35	34	42	47	na
FLB	104	111	110	109	101	105	105	93	88	84	76	80	67	71	97	76	na
FLC	62	63	52	47	34	36	33	23	16	17	17	13	16	18	56	12	na
<b>Total</b>	<b>210</b>	<b>225</b>	<b>210</b>	<b>202</b>	<b>173</b>	<b>184</b>	<b>185</b>	<b>153</b>	<b>137</b>	<b>133</b>	<b>128</b>	<b>121</b>	<b>118</b>	<b>123</b>	<b>195</b>	<b>135</b>	<b>na</b>

## **Commercial catch trends**

### **General production**

Since the early 1990s, annual commercial catch of the major species has generally declined (Fig. 2.1). Overall, total scalefish production (excluding sharks, small pelagics and cephalopods) has declined from over 1,100 t in the mid-1990s to around 300 t in recent years. The commercial catch of 'key' scalefish species in 2017/18 (317.7 t) was largely consistent with 2016/17, when 311.9 t were landed (Appendix 3, Table A1). In addition to the scalefish catch, 698 t of cephalopods, 14.1 t of sharks, and 2.7 t of small pelagics were landed in 2017/18. Total production (all species) in the fishery was 1046.9 t, up from 2016/17 when 716.1 t was harvested, largely due to increased catches of Gould's Squid in 2017/18.

When assessing trends within the fishery it is important to recognise that some species occur seasonally in Tasmanian waters and that availability can differ markedly between years. Such variability may not, therefore, necessarily reflect changes in stock status. Species in this category include Blue Warehou, Barracouta and Gould's Squid. By contrast, species such as Banded Morwong, Striped Trumpeter, Bastard Trumpeter, Longsnout Boarfish, Southern Calamari and Wrasse are more 'resident' species, and variability in catches can reflect a combination of factors, including market demand, management intervention, stock status and intrinsic variability in life history.

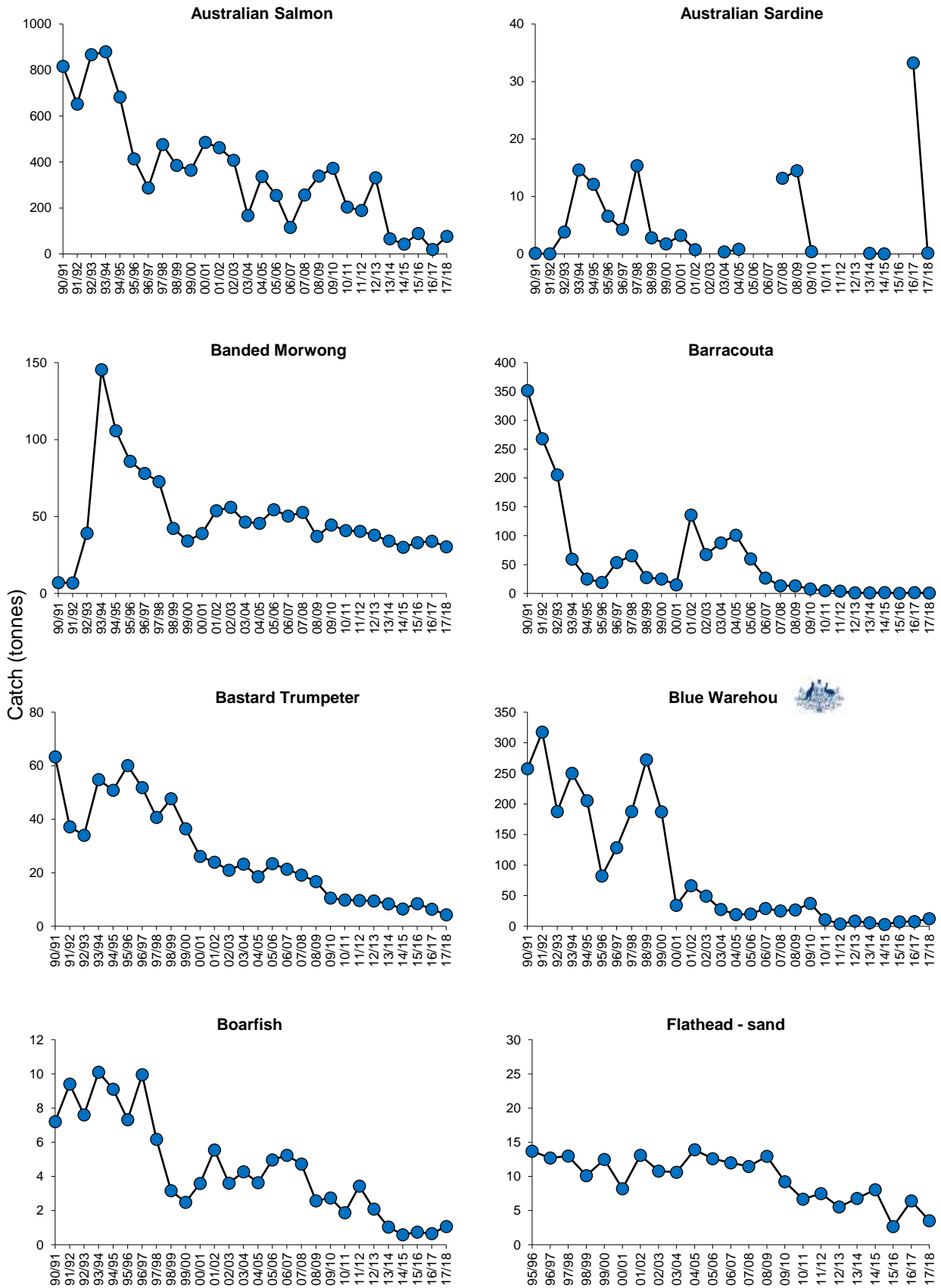



Figure 2.1 Annual catches (t) for selected scalefish species since 1990/91.  = Significant Commonwealth-managed fisheries

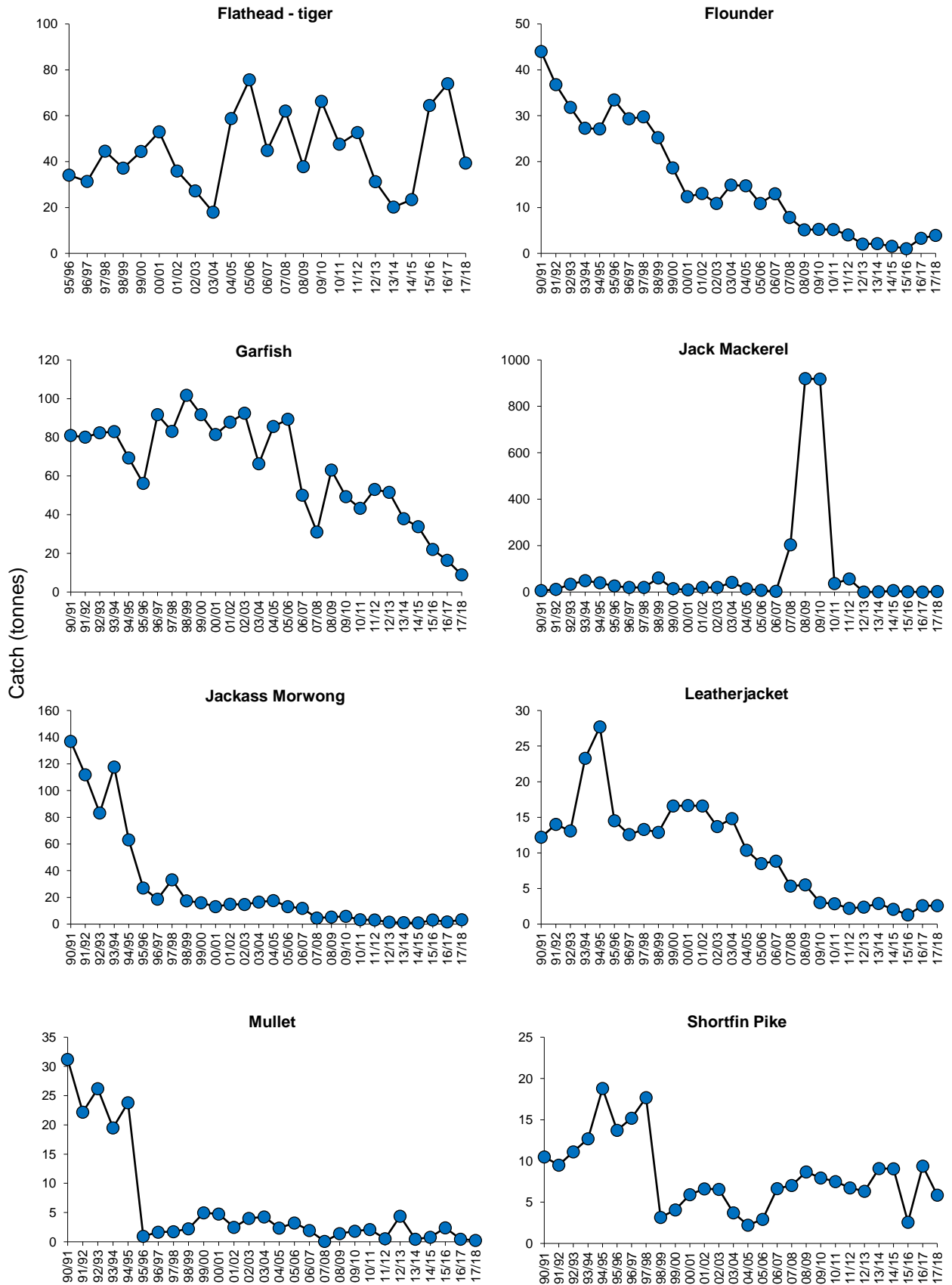


Figure 2.1 Continued.

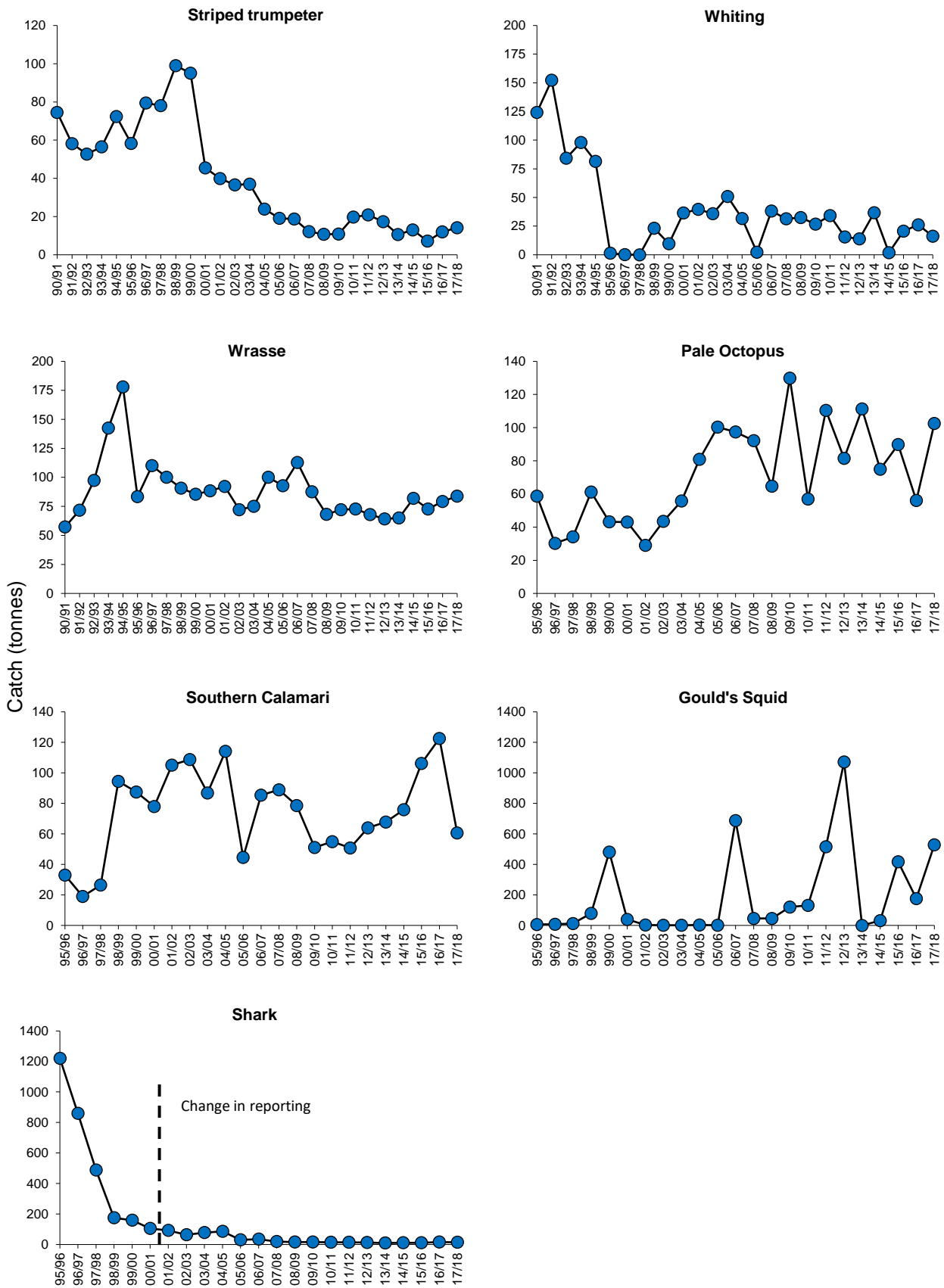


Figure 2.1 Continued.

## **Estuarine productions**

Estuarine production (as a subset of the whole general fishery) from selected monitoring estuaries totalled 41.0 t in 2017/18, a decline of 12.0 t from the previous year (Table 2.2). Catches came mainly from the southern estuaries including the Derwent River (ES1), Blackman Bay (ES17), Norfolk Bay (ES18), Frederick Henry Bay (ES19), and from the Tamar River (ES10) in the north of the state. Whiting, Southern Calamari, and Flounder were the main species captured. Port Davey (ES06) and Mersey River (ES08) are restricted to hand-line fishing, which accounts for the minimal catches in recent years. Georges Bay (ES12) and Ansons Bay (ES11) have been closed to commercial Scalefish fishing since 2004 and 2009, respectively.

**Table 2.2** Total commercial catches (t) in selected estuaries around Tasmania (a) by fishing year and (b) by key species in 2017/18.

a) By fishing year	ES01	ES06	ES07	ES08	ES09	ES10	ES11	ES12	ES17	ES18	ES19	ES20	Total	ES	Description
1995/96	17.39	0.67	4.43		0.41	10.75		0.43	2.92	26.44	14.12	3.22	80.78	ES01	Derwent River
1996/97	16.71	0.35	2.63		0.56	15.01		0.92	6.12	12.29	6.98	1.78	63.35	ES06	Port Davey
1997/98	14.28	0.16	1.41	<0.05	0.63	15.62		2.48	11.47	20.79	13.47	1.35	81.66	ES07	Macquarie Harbour
1998/99	14.21		1.38		0.90	19.60		1.59	10.04	36.50	23.19	4.87	112.28	ES08	Mersey River
1999/00	4.73		0.98		0.45	14.15	0.18	2.56	18.90	28.51	10.23	2.77	83.46	ES09	Port Sorell
2000/01	16.10		0.25		0.13	12.70	0.05	1.17	15.46	27.93	27.33	1.88	103.00	ES10	Tamar River
2001/02	13.88		2.23		0.19	73.82		1.19	8.86	64.06	32.33	2.00	198.56	ES11	Ansons Bay
2002/03	28.13		8.02		0.16	27.64	0.55	0.81	14.55	35.23	23.00	1.57	139.66	ES12	Georges Bay
2003/04	40.05		6.06		1.00	25.12			5.17	59.52	21.83	0.81	159.56	ES17	Blackman Bay
2004/05	25.99		4.93		1.76	34.47		<0.05	9.46	25.87	23.14	0.66	126.28	ES18	Norfolk Bay
2005/06	2.19	0.07	23.16		0.95	33.15	1.29		6.64	14.18	9.67	0.84	92.14	ES19	Frederick Henry Bay
2006/07	30.97	0.25	9.93		2.00	23.60	0.17		8.72	20.01	19.74	1.36	116.75	ES20	Pitt Water
2007/08	31.87	<0.05	3.16			15.26		<0.05	12.31	26.94	12.11	0.87	102.52		
2008/09	32.22		1.14		0.18	20.90		<0.05	8.38	15.75	10.45	2.07	91.09		
2009/10	26.91		0.72		0.46	15.22	<0.05	<0.05	3.93	15.57	4.39	2.07	69.27		
2010/11	27.84	0.11	0.44		0.60	10.25			5.65	5.82	13.71	1.69	66.11		
2011/12	13.88		0.28			8.39			4.95	6.88	6.70	1.89	42.97		
2012/13	12.19	0.07	0.13		<0.05	12.22	0.20		6.72	13.27	3.11	0.85	48.76		
2013/14	32.28		1.06		0.29	9.69			2.97	6.74	8.75	1.09	62.87		
2014/15	1.76	<0.05	<0.05		0.40	8.90		0.10	3.25	8.51	0.87	0.72	24.51		
2015/16	17.51				0.82	10.34		0.13	3.10	5.11	3.81	0.58	41.40		
2016/17	26.24		0.05		0.17	12.63			2.77	4.13	4.61	2.36	52.96		
2017/18	16.07		0.78			8.79			1.94	6.59	3.81	3.03	41.01		

Table 2.2 cont.

b) By species	ES01	ES06	ES07	ES08	ES09	ES10	ES11	ES12	ES17	ES18	ES19	ES20	Total
Australian Salmon						2.00						0.02	2.02
Calamari						1.82			1.60	3.41	3.01	0.03	9.86
Flathead						0.02							0.02
Flounder			0.23						0.24	0.38	0.04	2.87	3.76
Garfish						1.02				0.13	0.61	0.11	1.87
Mullet						0.14							0.14
Octopus						0.05			0.01	2.54			2.60
Trevally						0.53							0.53
Whiting	16.07												16.07
Wrasse						1.60			0.05		0.11		1.76

## **Recreational fishery**

### **Catch and effort**

Surveys of the recreational fishery conducted in 2000/01, 2007/08 and 2012/13 provide comprehensive snapshots of the Tasmanian recreational fishery (Henry and Lyle 2003, Lyle 2005, Lyle et al. 2009, Lyle and Tracey 2012, Lyle et al. 2014a). In addition, there have been targeted surveys of recreational gillnetting in 1996–98 and 2010 (Lyle 1999, Lyle and Tracey 2012) and offshore boat fishing in 2011/12 (Tracey et al. 2013). The most recent recreational fishing survey indicates that there has been a general decline in participation, in both absolute and relative terms (i.e. percentage of the population) (Lyle et al. 2014a). Nevertheless, recreational landings represent a significant component of the total harvest for many species, either as a proportion of the total harvest or in absolute quantities taken (Table 2.3). For instance, recreational catches likely exceed commercial catches of Australian Salmon, Blue Warehou, Striped Trumpeter, Bastard Trumpeter, Sand Flathead, Jackass Morwong, Barracouta and Flounder. By contrast, the commercial sector dominates the catches of Tiger Flathead, Banded Morwong, Southern Calamari, Southern Garfish, Pike, Wrasse and School Whiting.

A particularly conspicuous change in the recreational fishery has been the more than threefold increase in the landings of Southern Calamari since 2000/01, such that the recreational harvest is now comparable to that of the commercial sector (Table 2.3). Another is the significant recreational catch of Blue Warehou in 2012/13, which exceeds the Tasmanian commercial catch and recent estimates of Commonwealth commercial catches. Blue Warehou is classified as overfished in the latest national stock status reports (Woodhams et al. 2013, Flood et al. 2014, Patterson et al. 2017) and is subject to a rebuilding strategy at the Commonwealth level.

**Table 2.3** Estimated recreational harvest (numbers and weight) for key scalefish species taken by Tasmanian residents, commercial landings (weight) and proportion of the total (recreational plus commercial) catch represented by the recreational harvest (refer to Lyle et al. (2014a)). Note: the survey periods do not correspond with fishing years; 2000/01 represented the period May 2000 to Apr 2001, 2007/08 represented the period Dec 2007 to Nov 2008 and 2012/13 represented the period Nov 2012 to Oct 2013.\* estimated from the 2011/12 offshore recreational fishing (Tracey et al. 2013); \*\* estimated from the 2010 recreational gillnetting survey (Lyle and Tracey 2012).

Species	2000/01				2007/08				2012/13			
	Rec. harvest		Com.	Rec.	Rec. harvest		Com.	Rec.	Rec. harvest		Com.	Rec.
	No.	(t)	(t)	%	No.	(t)	(t)	%	No.	(t)	(t)	%
Flathead	1,236,675	322.0	63.4	83.5	1,066,293	293.0	73.2	80.0	924,932	235.9	39.9	85.5
Australian Salmon	300,456	105.0	485.0	17.8	110,312	48.1	299.8	13.8	144,712	63.7	270.8	19.0
Mullet	111,025	30.0	13.7	68.6	24,152	6.6	2.4	73.3	26,265	7.1	7.9	47.3
Flounder	50,582	15.2	10.5	59.1	32,436	10.1	7.8	56.3	23,238	7.2	2.1	77.4
Cod	65,115	30.6	4.0	88.4	14,263	8.2	2.5	76.7	10,464	6.1	2.2	73.5
Jackass Morwong	27,041	31.9	13.7	70.0	9,979	6.8	3.8	64.2	23,732	16.1	2.1	88.5
Garfish	15,669	1.9	81.4	2.3	14,568	2.0	51.0	3.7	15,260	2.0	50.6	3.8
Whiting	7,480	0.8	42.5	1.9	14,992	3.4	35.4	8.7	9,412	2.1	35.8	5.5
Black Bream	34,336	22.0	-	100.0	13,134	11.4	-	100.0	19,153	16.7	-	
Barracouta	24,320	46.9	15.1	75.7	11,577	10.8	13.9	43.8	32,954	31.0	1.1	96.6
Wrasse	23,083	13.6	88.4	13.3	11,640	10.3	68.5	13.1	7,223	6.4	65.2	8.9
Blue Warehou	16,359	14.6	36.3	28.6	8,723	7.0	26.6	20.8	10,757	15.4	8.8	63.6
Jack Mackerel	15,770	3.2	8.6	26.8	5,216	1.0	225.7	0.4	30,907	5.2	0.2	96.3
Striped Trumpeter	13,450	29.6	49.6	37.4	7,274*	31.9*	19.8	61.7	3,476	15.2	10.5	59.1
Bastard Trumpeter	29,130	37.0	26.2	58.5	27,527**	27.3**	10.5	72.2	7,573	7.5	9.8	43.4
Leatherjackets	18,706	8.2	16.7	33.0	7,619	2.6	4.2	38.0	5,389	1.8	2.5	41.9
Silver Trevally	16,812	4.7	1.6	74.6	10,636	4.2	2.0	67.9	4,826	1.9	2.8	40.4
Southern Calamari	29,473	17.7	76.6	18.8	40,525	44.6	102.6	30.3	57,728	63.5	60.2	51.3
Gould's Squid	9,903	5.0	39.7	11.1	73,236	36.6	45.8	44.4	42,853	21.4	1054.7	2.0

## Recreational gillnet fishery

The use of recreational nets in Tasmania has been subject to licensing since 1995, with fishers able to licence up to two graball nets (gillnets) prior to 2002, along with one mullet net and a beach seine. From November 2002 the number of graball nets that could be licensed was reduced to one per person. The number of recreational net licences issued rose rapidly from around 8,900 in 1995 to over 11,000 in 1999/2000, licence numbers then stabilised between 8,000 to 9,000 for several years. Licence numbers climbed again to around 10,000 in 2007/08 before trending downward and stabilising at around 7,800 in the past two years (Table 2.4). It is possible that the reduction in licence numbers since 2009/10 occurred in response to the introduction of maximum soak times for gillnets in 2009. Night netting, which was a common and popular practice amongst recreational fishers (Lyle 2000), was banned for recreational fishers (with the exception of Macquarie Harbour) in late 2004. While this appeared to have little discernible impact on licence numbers, a targeted survey of recreational gillnetting in 2010 revealed a concomitant reduction in overall gillnet effort (effort in 2010 was about 60% of the level in 1997 despite there being 40% more gillnet licence-holders, Lyle and Tracey 2012). Furthermore, only 73% of recreational licences were used during 2010 (Lyle and Tracey 2012).

The 2010 survey revealed that almost 65% of the gillnet catch (by number) was kept (Lyle and Tracey 2012). Bastard Trumpeter and Blue Warehou together represented 45% of the total retained catch, Atlantic salmon contributed a further 10%, with Australian Salmon, Jackass Morwong, Mullet and Wrasse of secondary importance. Wrasse was the most significant by-catch. Recreational gillnet catches of Bastard Trumpeter, Mullet, Jackass Morwong, Leatherjacket and Cod were higher when compared to the commercial catches, while Blue Warehou catches were similar between the two sectors. Recreational gillnet catch rates have fallen from an average of >6 fish retained per net set in 1997 to just over 4 fish per set throughout the past decade. While variability in the abundance of target species has contributed to this trend (especially Blue Warehou), changes in fishing practices (no night netting, shorter average set durations, reduction in the length of mullet nets, larger minimum size limits for some species influencing release/discarding rates, etc.) have been contributing factors.

**Table 2.4** The number of recreational gillnet licences issued by licensing year since 1995/96. na = not applicable.

Licence type	95/96	96/97	97/98	98/99	99/00	00/01	01/02	02/03	03/04	04/05	05/06
Graball 1	5615	6290	6685	6709	7477	7401	6960	7695	7313	7408	8054
Graball 2	2612	2678	2683	2426	2652	2515	1841	na	na	na	na
Mullet Net	656	684	738	739	879	845	608	754	753	754	816
<b>Total</b>	<b>8883</b>	<b>9652</b>	<b>10106</b>	<b>9874</b>	<b>11008</b>	<b>10761</b>	<b>9409</b>	<b>8449</b>	<b>8066</b>	<b>8162</b>	<b>8870</b>

Licence type	06/07	07/08	08/09	09/10	10/11	11/12	12/13	13/14	14/15	15/16	16/17
Graball 1	8677	9185	9172	8960	8162	8248	7995	7765	7887	7070	7092
Graball 2	na	na	na	na	na	na	na	na	na	na	na
Mullet Net	877	995	1080	922	886	888	890	841	855	789	793
<b>Total</b>	<b>9554</b>	<b>10180</b>	<b>10252</b>	<b>9882</b>	<b>9048</b>	<b>9136</b>	<b>8885</b>	<b>8606</b>	<b>8742</b>	<b>7859</b>	<b>7885</b>

Licence type	17/18
Graball 1	6670
Graball 2	na
Mullet Net	806
<b>Total</b>	<b>7476</b>

## 3. Commercial fishing gear

### General effort trends

Following the introduction of the new management arrangements in November 1998, beach seine, purse seine, gillnet and hand-line effort declined, whereas drop-line, squid-jig and dip-net effort increased sharply. While a range of factors, including availability of target species and market developments, have had an influence, there is little doubt that management changes have had a direct impact on effort levels. Specifically, there was a decline in the effort of fishing methods for which gear allocations, or access, became more regulated (beach seine, purse seine and gillnets), whereas there was an initial shift to or increase in effort for less regulated methods (hooks, jigs and dip-nets; i.e. gear that was generally available to most licence-holders).

Since the early 2000s effort for most fishing methods has declined, exceptions being hand-line, which has remained relatively stable, and automatic squid-jig which has peaked sporadically with the periodic occurrence of Gould's Squid in Tasmanian waters. For example, catches of Gould's Squid were at a record high level during 2012/13 whereas little fishing was reported for the automatic jig fishery in Tasmanian waters during 2013/14 and 2014/15, due in part to the saturation of local markets from the large catches in 2012/13.

Total effort (days fished) for automatic jig, beach seine, hand collection, hand-line and spear increased in 2017/18 relative to 2016/17. In contrast, total effort (days fished) for drop-line, dip net, Danish seine, fish trap, gillnet net, small mesh net, purse seine, squid jig and trolling declined in 2017/18 relative to 2016/17 (Figures 3.1–3.27; more detailed data available in Table A2, Appendix 4).

There is potential for future effort increases due to the levels of latent effort from licence-holders who are currently inactive in the fishery or participating at low levels. The 2004 management plan review in particular attempted to address this issue through several strategies including non-transferability of C-class licences.

The following section presents an overview of the catch composition, as well as the overall catch, effort, and number of vessels for each fishing method.

## Automatic squid-jig

Automatic squid-jig users target exclusively Gould's Squid, and have practically no by-catch.

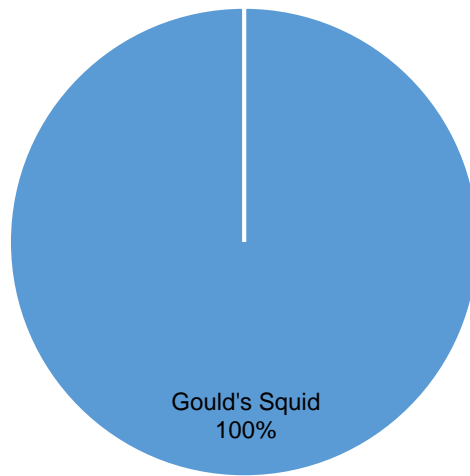


Figure 3.1 Automatic squid jig catch composition for 2017/18.

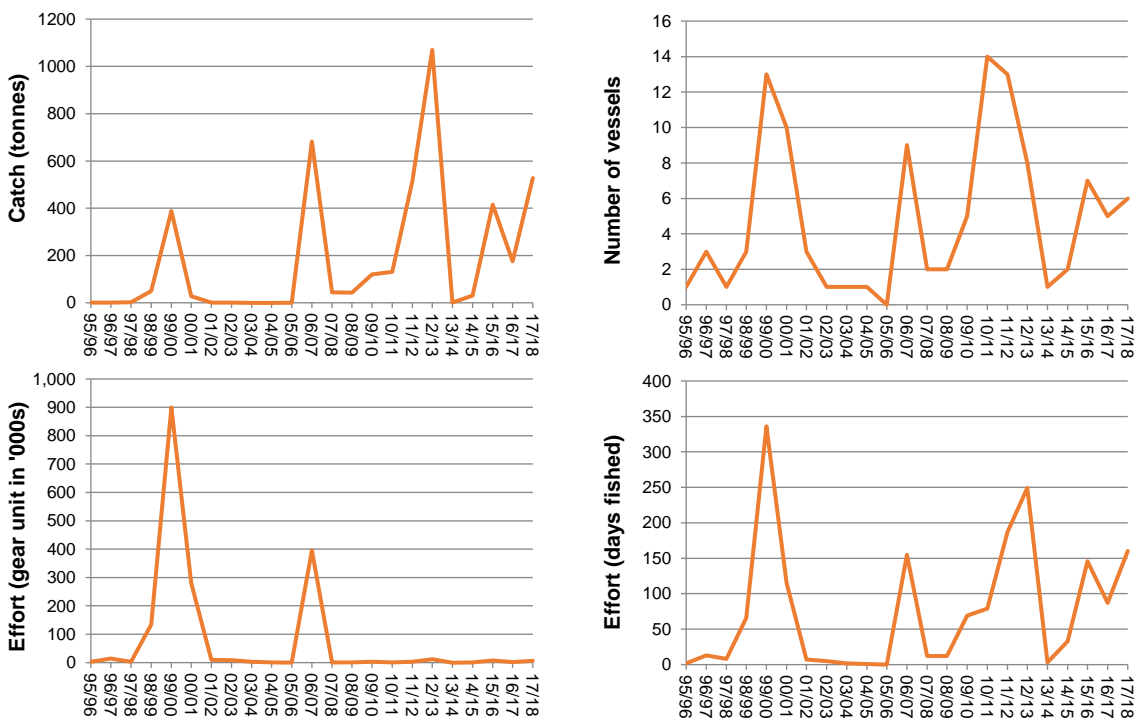
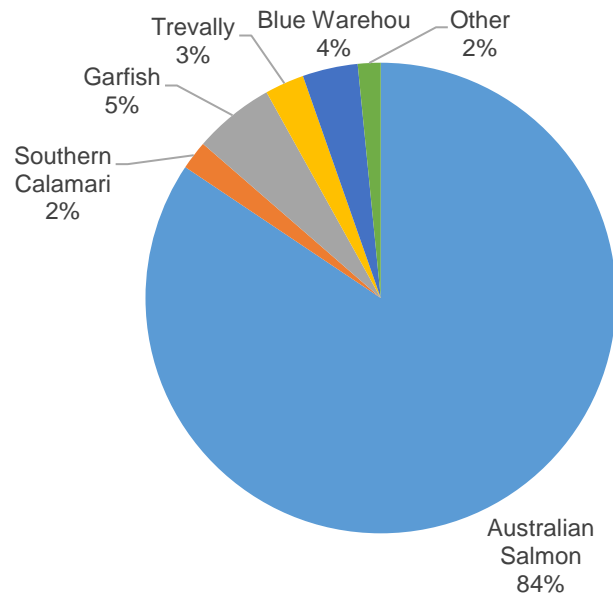
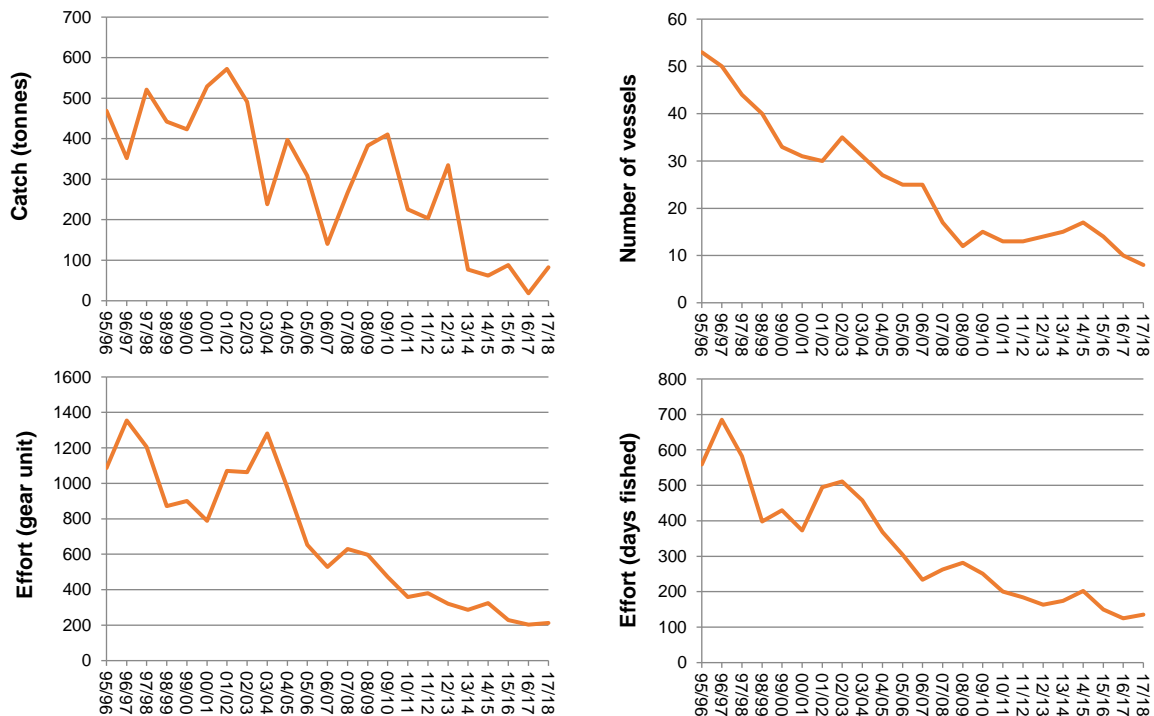


Figure 3.2 Overall catch, number of vessels using the gear, and effort (in gear unit and days fished) for automatic squid-jig.

**Beach seine**



**Figure 3.3** Beach seine catch composition for 2017/18.



**Figure 3.4** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for beach seine.

## Drop-line

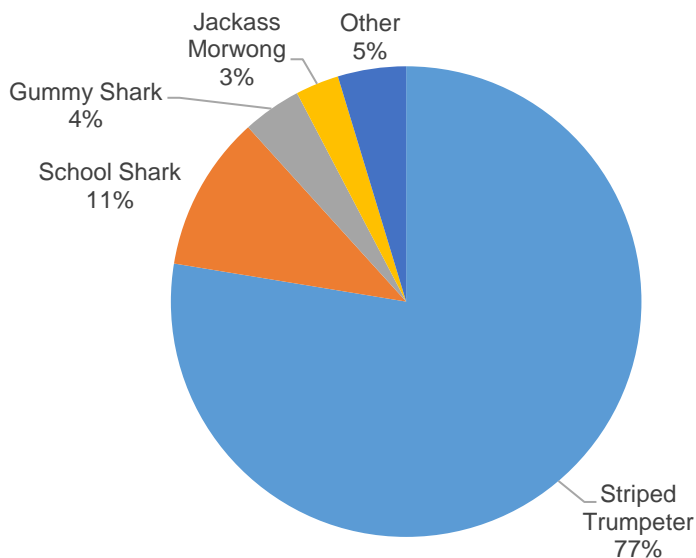


Figure 3.5 Drop-line catch composition for 2017/18.

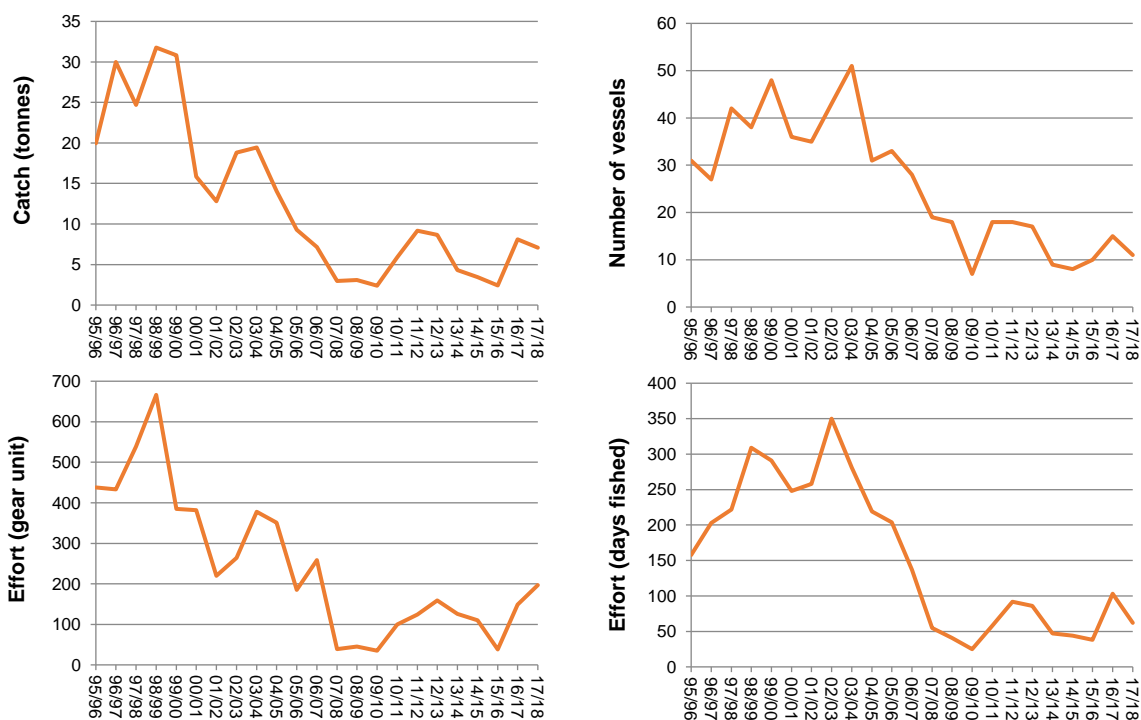
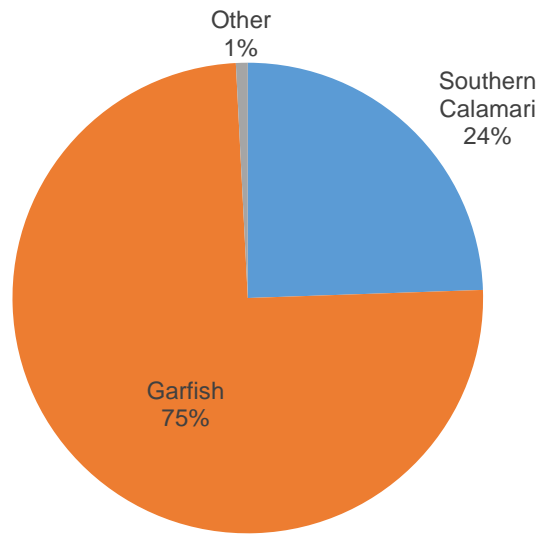
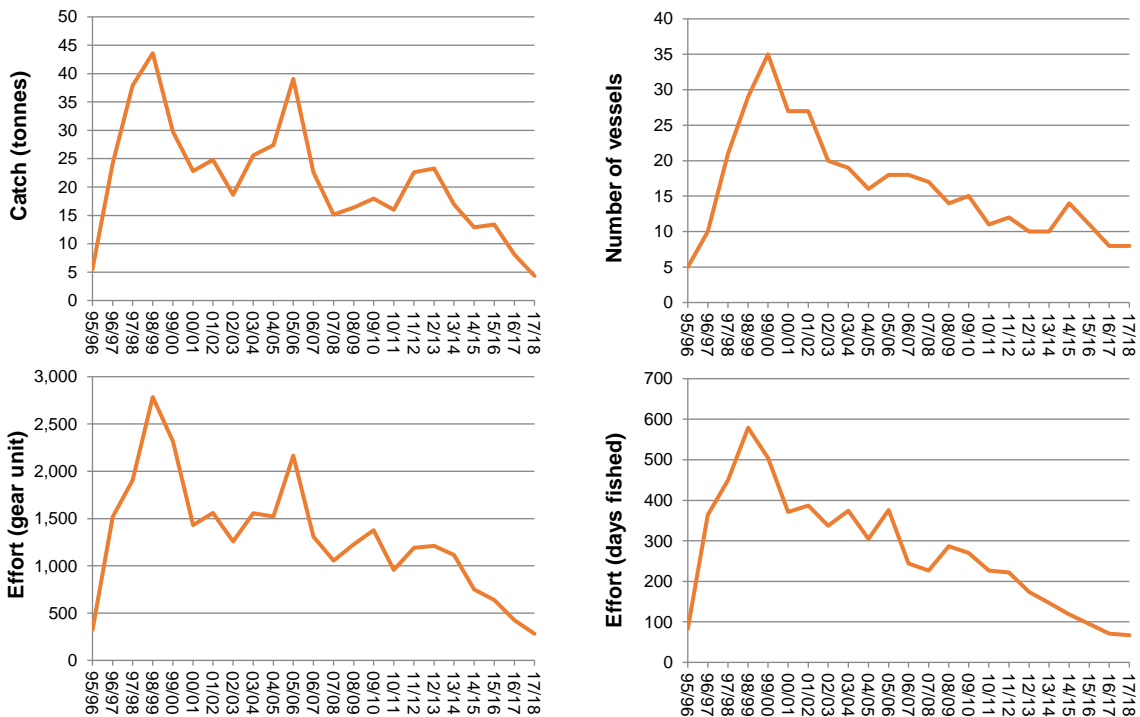


Figure 3.6 Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for drop-line. For the purpose of analysis, drop-line catch and effort up to 1998 was restricted to records that indicated a fishing depth of less than 200 m, effectively excluding effort targeting Blue-eye Trevalla but encompassing the target fishery for Striped Trumpeter (less than 1% of the Striped Trumpeter catch has been reported from depths greater than 200 m; since 1998 fishing for Blue-eye has been covered in Commonwealth catch returns).

**Dip-net**



**Figure 3.7** Dip-net catch composition for 2017/18.



**Figure 3.8** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for dip-net.

## Danish seine

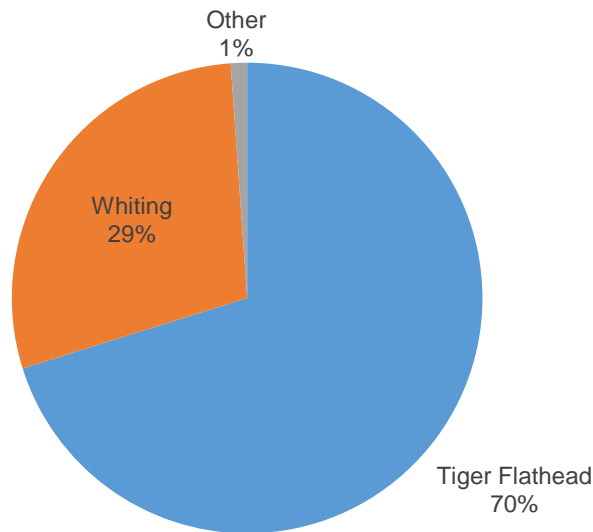


Figure 3.9 Danish seine catch composition for 2017/18.

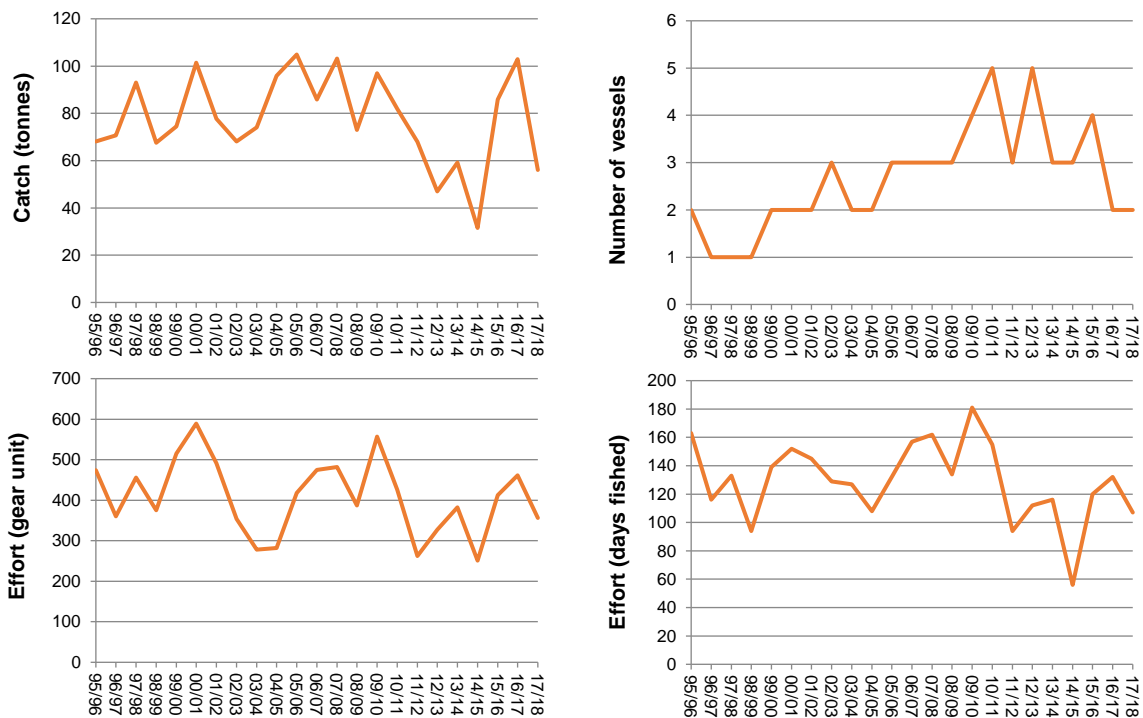
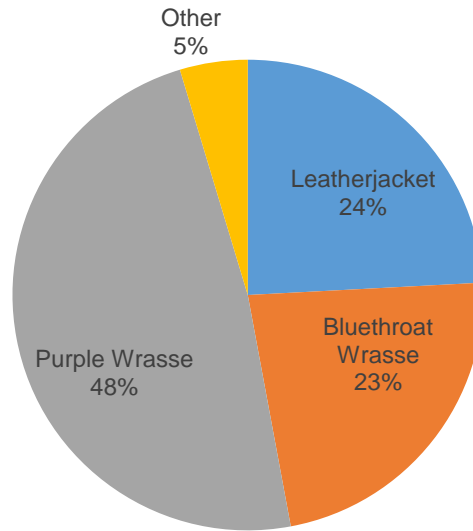
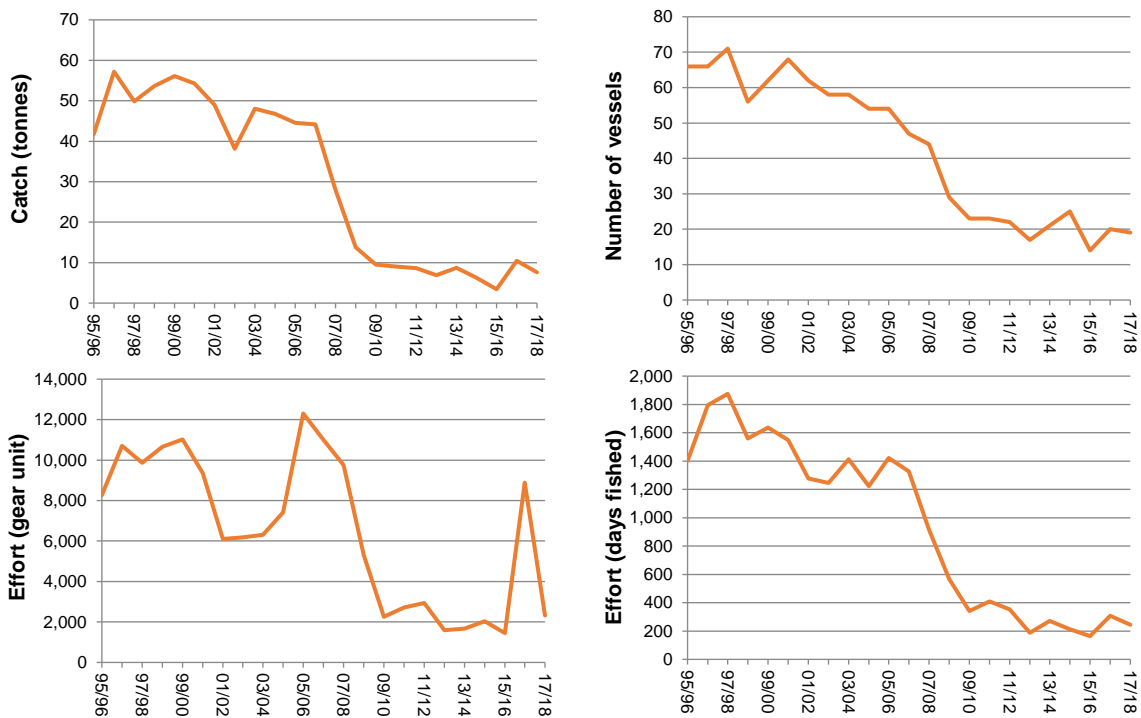


Figure 3.10 Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for Danish seine.

**Fish trap**



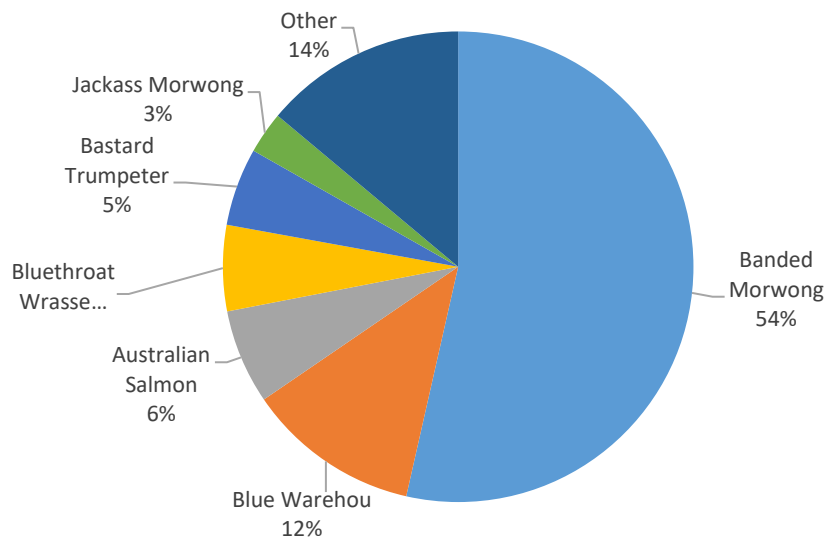
**Figure 3.11** Fish trap catch composition for 2017/18.



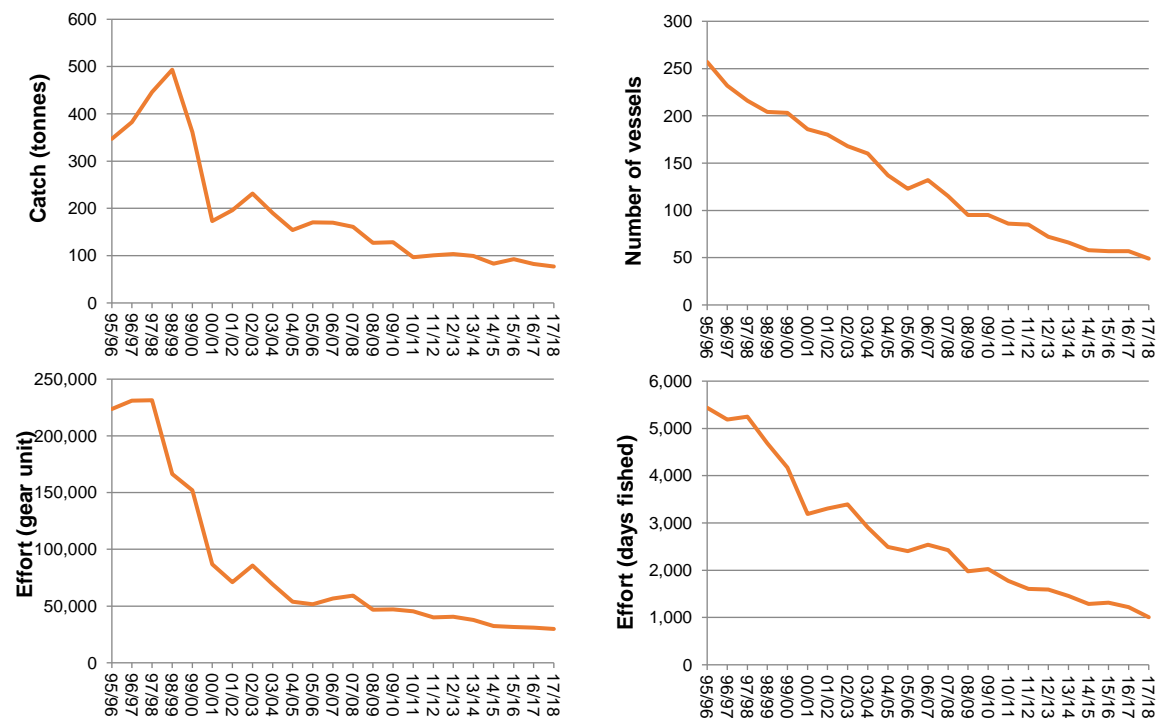
**Figure 3.12** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for fish trap.

## Gillnet

Gillnets in this analysis include both traditional gillnets (~110 mm mesh size) and 'Banded Morwong' nets (~140 mm mesh size).

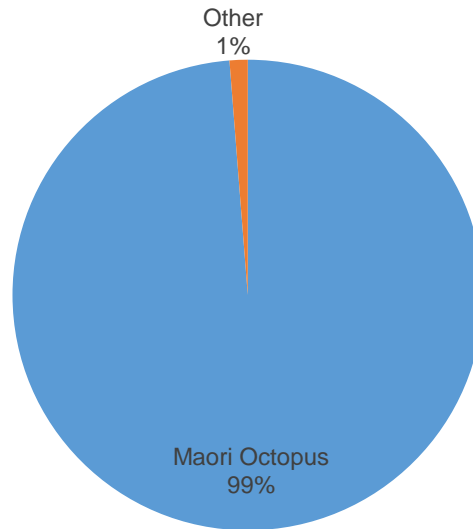


**Figure 3.13** Gillnet catch composition for 2017/18. Note figure includes Banded Morwong that were released.

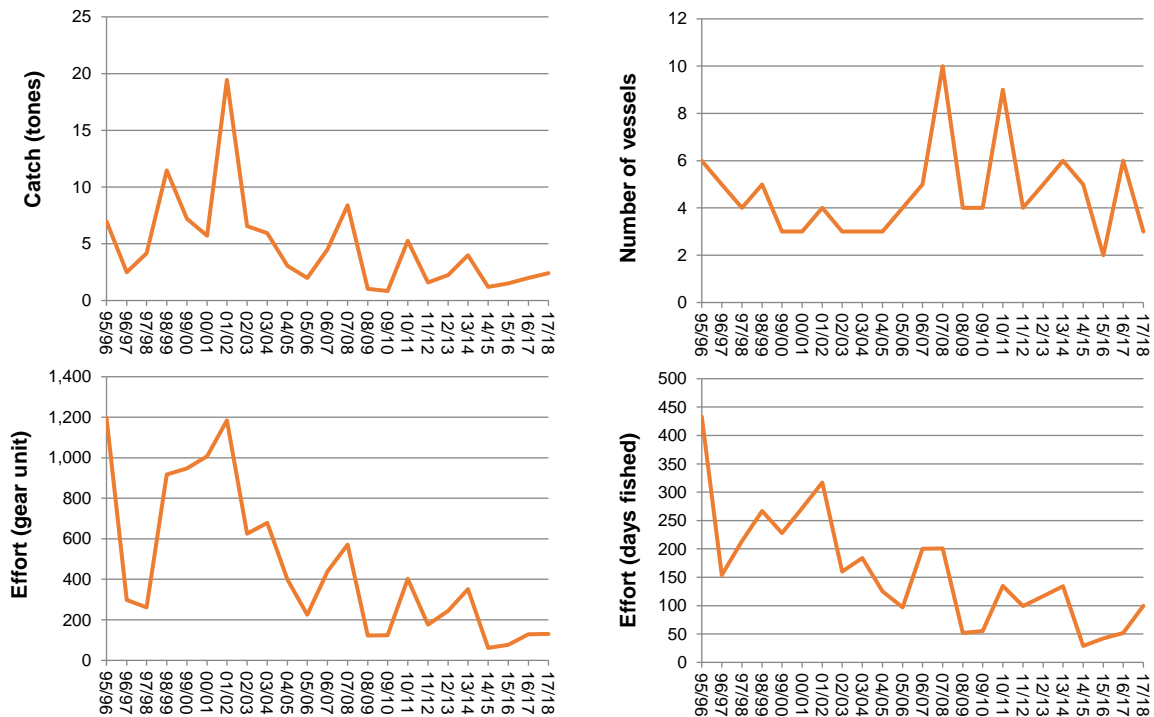


**Figure 3.14** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for gillnet.

**Hand collection**



**Figure 3.15** Hand collection catch composition for 2017/18.



**Figure 3.16** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for hand collection.

## Hand-line

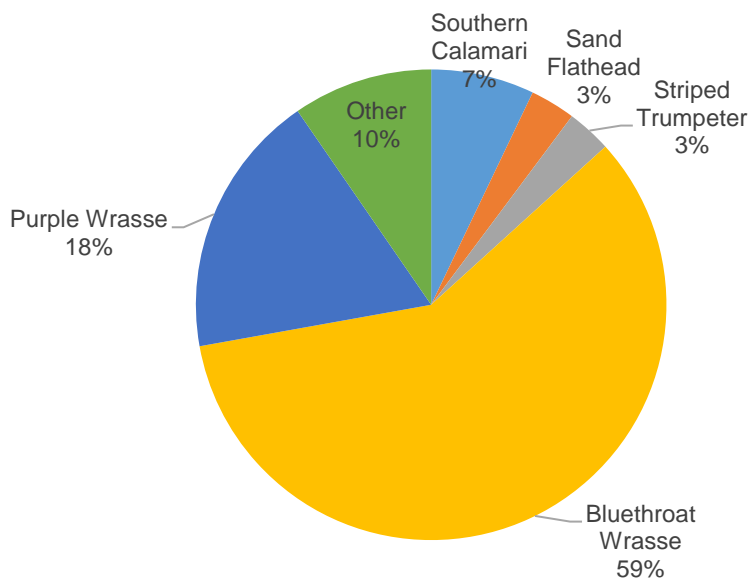


Figure 3.17 Hand-line catch composition for 2017/18.

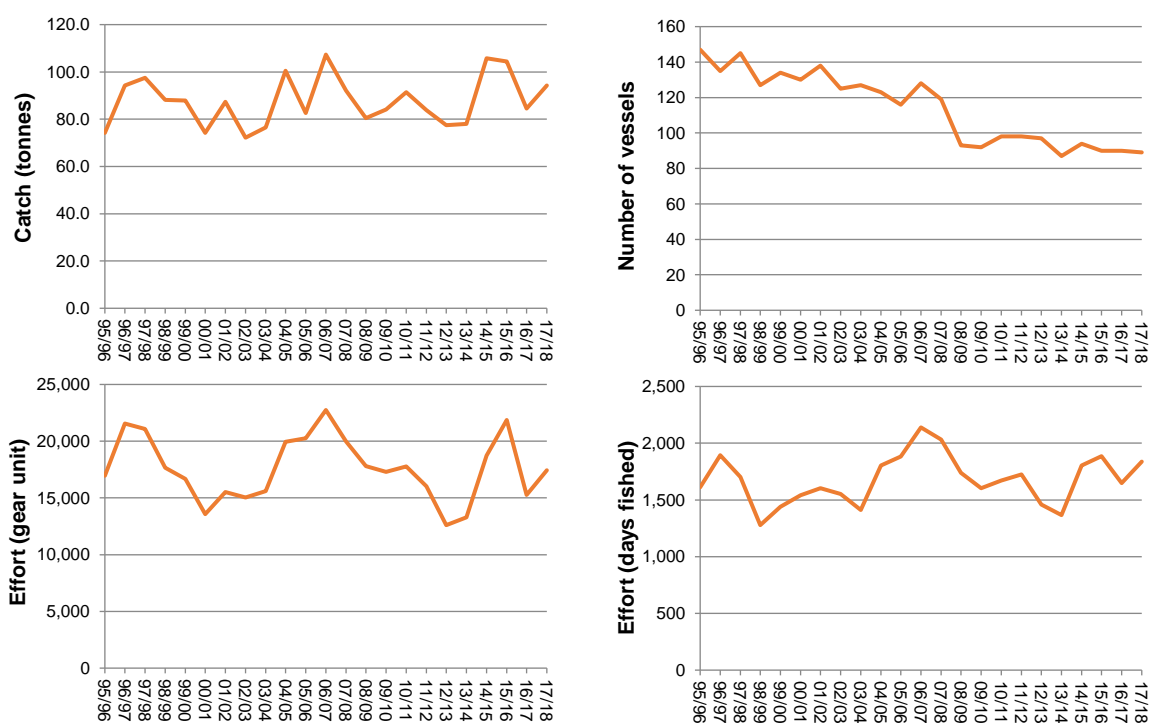
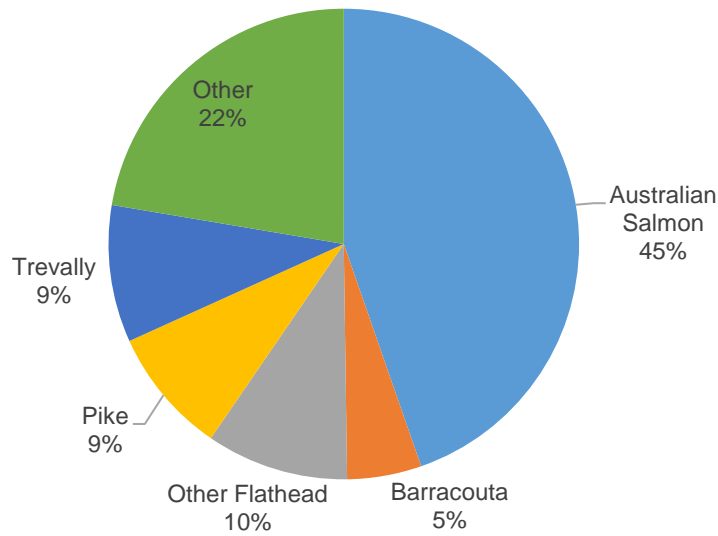


Figure 3.18 Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for hand-line.

**Small mesh net**

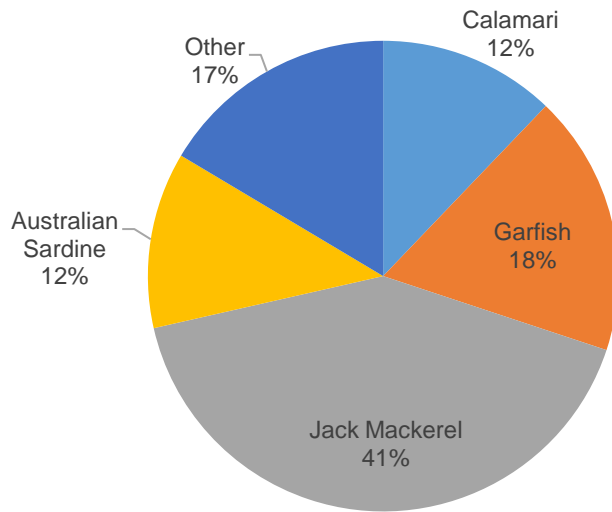


**Figure 3.19** Small mesh net catch composition for 2017/18.

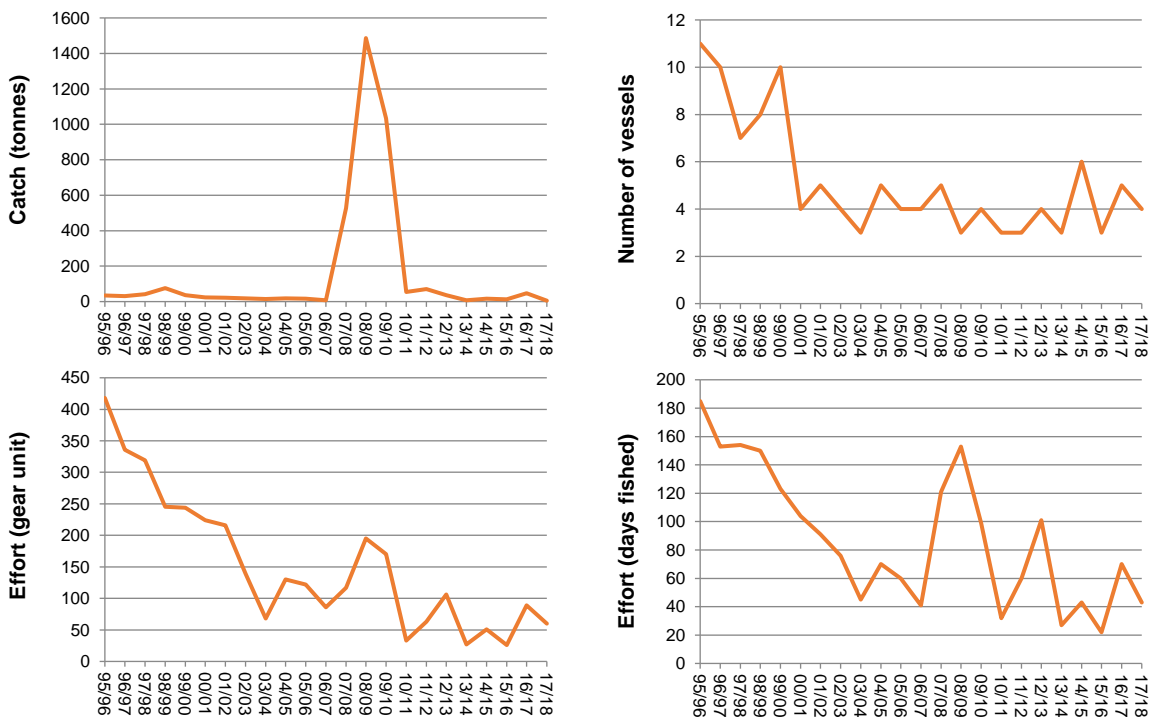


**Figure 3.20** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for small mesh net.

**Purse seine**

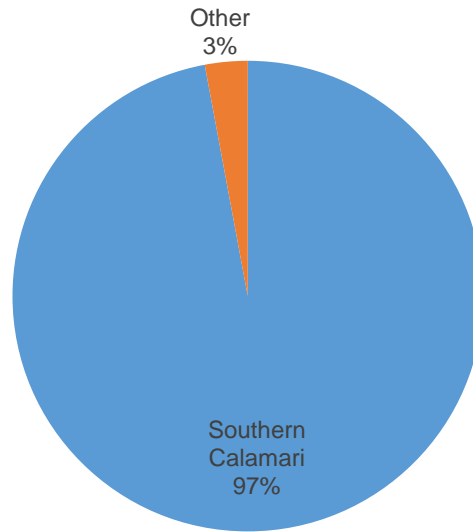


**Figure 3.21** Purse seine catch composition for 2017/18.

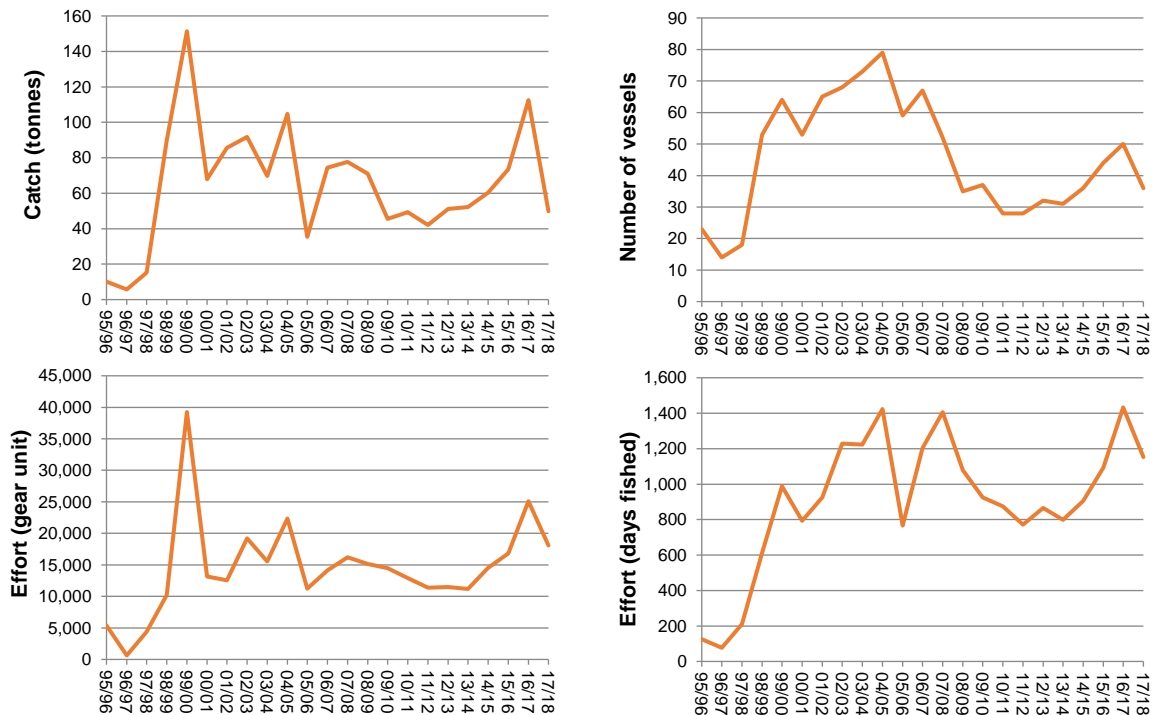


**Figure 3.22** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for purse seine.

**Squid jig**

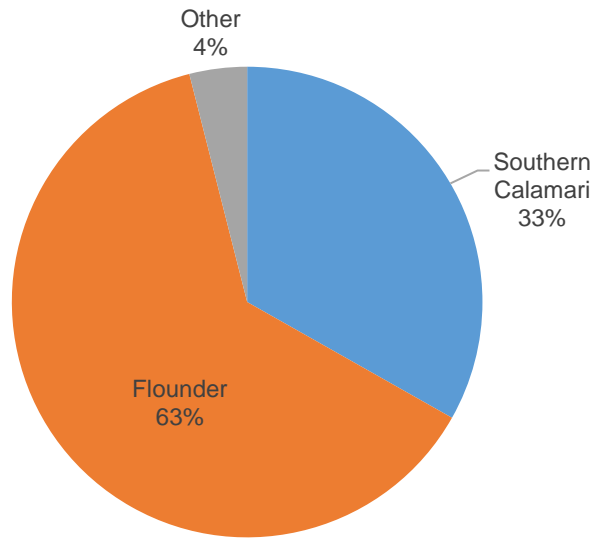


**Figure 3.23** Squid jig catch composition for 2017/18.

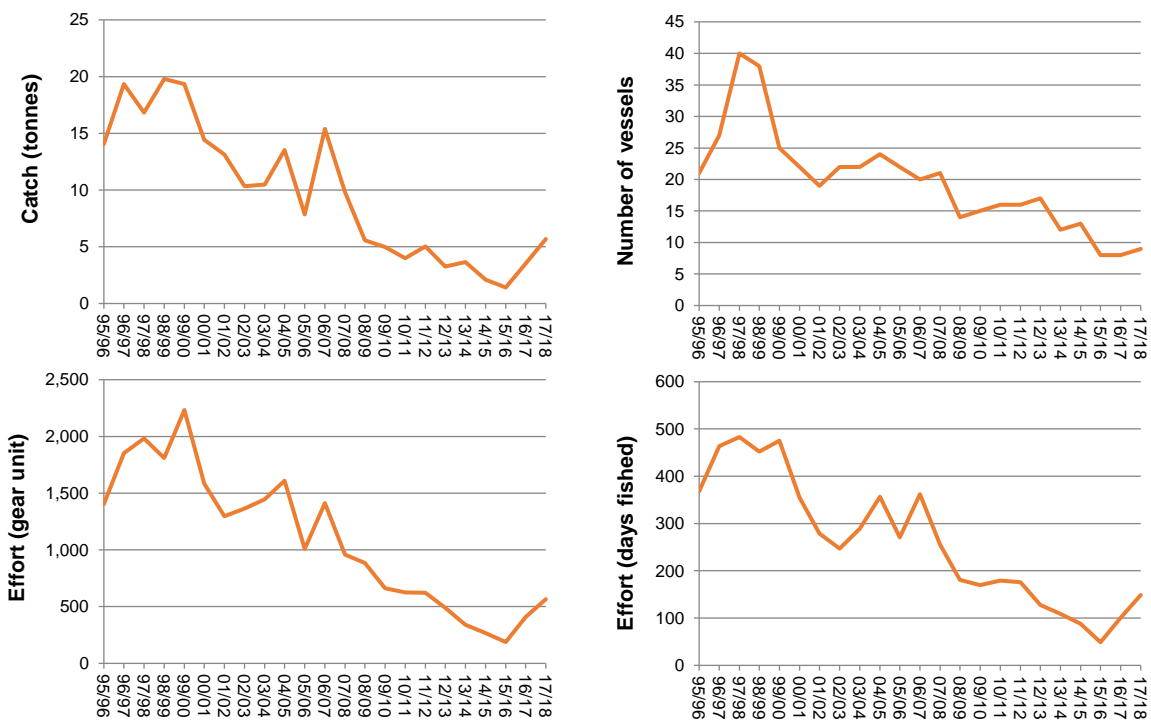


**Figure 3.24** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for squid jig.

**Spear**

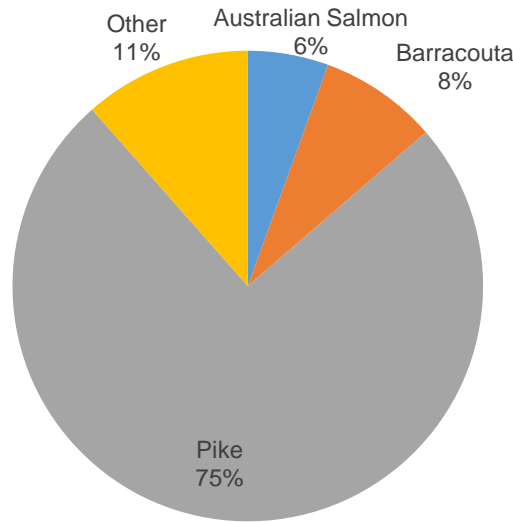


**Figure 3.25** Spear catch composition for 2017/18.

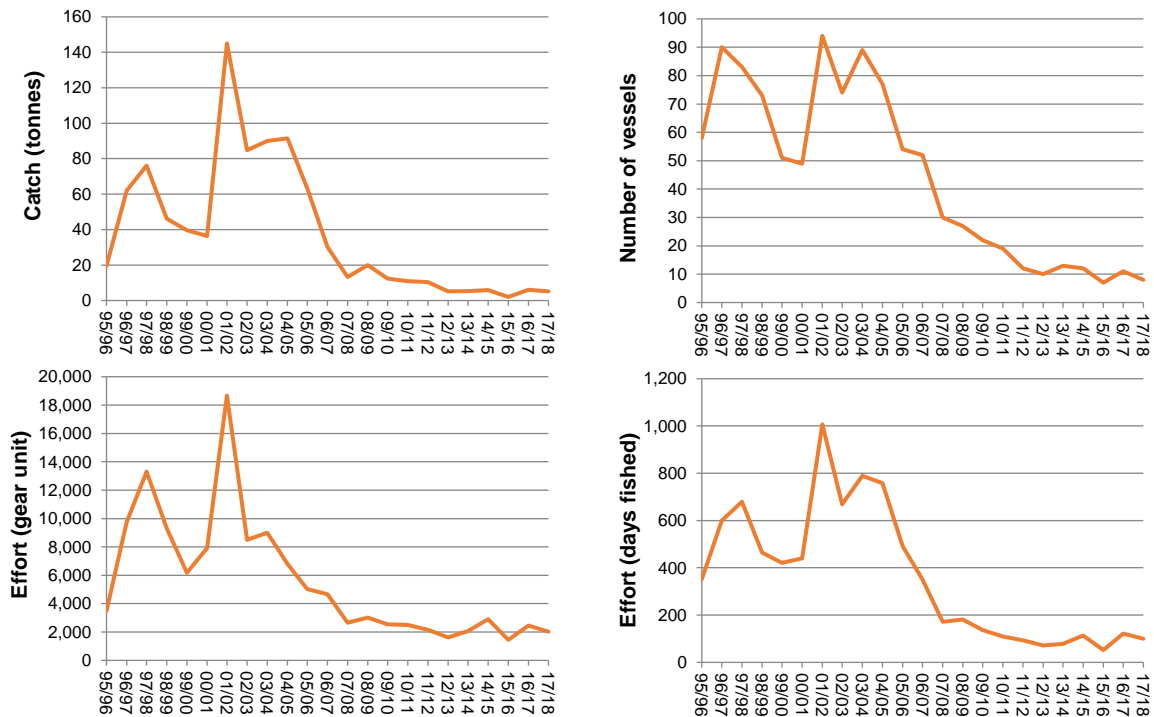


**Figure 3.26** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for spear.

**Trolling**



**Figure 3.27** Trolling catch composition for 2017/18.

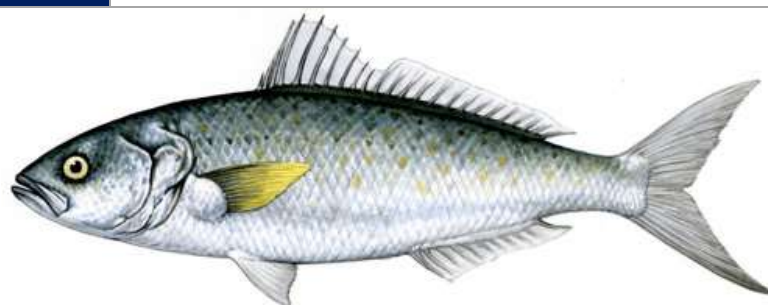


**Figure 3.28** Overall catch (t), number of vessels using the gear, and effort (in gear unit and days fished) for trolling.

## 4. Eastern Australian Salmon

### *Arripis trutta*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
Species has a long history of exploitation across south-eastern Australian. Low commercial landings in Tasmania in recent years are driven by market demand not abundance. The current level of fishing pressure in Tasmania is well below historical levels and unlikely to cause the biological stock to become recruitment impaired.	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Arripis trutta*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Open water. Down to 30 m depth.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>From Victoria to Queensland (Brisbane) and around Tasmania.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Fish (pelagic predominantly).</li> </ul>	Stewart et al. (2011)
Movement and stock structure	<ul style="list-style-type: none"> <li>One single well-mixed stock along southeast Australia</li> <li>Travel great distances between the different States.</li> <li>Increased population mixing occurs with both increasing age and decreasing latitude.</li> </ul>	Stewart et al. (2011)
Natural mortality	<ul style="list-style-type: none"> <li><math>M</math> between 0.35 and 0.50.</li> </ul>	Stewart et al. (2011)
Maximum age	<ul style="list-style-type: none"> <li>Maximum sampled is 12 years but potentially up to 26 years.</li> </ul>	Stewart et al. (2011)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 89.0 cm.</li> <li>Maximum weight: 9.4 kg</li> <li>Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math> where <math>L</math> is the fork length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</li> </ul>	Edgar (2008) Frimodt (1995) Stewart et al. (2011)

	Parameter estimates are:				
	<b>Sex</b>	$L_{\infty}$	$k$	$t_{0\infty}$	
	Females	65.0	0.24	-0.15	
Males	61.3	0.27	-0.13		

Maturity	<ul style="list-style-type: none"> <li>• Size-at-50% maturity: 42.13 cm</li> <li>• Age at 50% maturity: 2.19 years</li> </ul>	Stewart et al. (2011)
Spawning	<ul style="list-style-type: none"> <li>• From October to March, off New South Wales.</li> <li>• The relationship between batch fecundity and fork length is exponential with <math>F = 14581e^{0.0659L}</math>, where <math>F</math> is the fecundity (in number of eggs) and <math>L</math> is the fork length (cm).</li> <li>• The relationship between batch fecundity and age is exponential with <math>F = 96604e^{0.227A}</math>, where <math>F</math> is the fecundity (in number of eggs) and <math>A</math> is the age (in years).</li> </ul>	Stewart et al. (2011)
Early life history	<ul style="list-style-type: none"> <li>• Eggs, larvae and juveniles drift and migrate from spawning grounds to Tasmania and Victorian waters during autumn and winter.</li> <li>• Juveniles (4–6 cm fork length) appear in shallow Tasmanian waters between January and September.</li> </ul>	Kailola et al. (1993)
Gillnet post release survival	<ul style="list-style-type: none"> <li>• Low: 20–62% depending on soak duration</li> </ul>	Lyle et al. (2014b)

## **Background**

There are two species of Australian Salmon cohabiting in Tasmanian waters: *Arripis trutta* (Eastern Australian Salmon) and *Arripis truttaceus* (Western Australian Salmon). Eastern Australian Salmon constitutes approximately 94% of Tasmanian commercial catches.

Australian Salmon have a long history of exploitation in Tasmania, with large-scale commercial fishing occurring at least since 1958 (Stewart et al. 2011). There are two distinct sectors in the commercial fishery: a small number of large vessels specifically equipped to capture and store large quantities of Australian Salmon, and a large number of small vessels which target the species on an opportunistic basis or take them as by-product. A single company operating up to three vessels has typically accounted for around 85% of landings for the species. A total commercial catch trigger of 435 t (120% of the average annual catch between 1996/97 and 2006/07) applies to the Tasmanian fishery.

Most commercially caught Australian Salmon are frozen whole and sold as rock lobster bait with production levels linked to the demand for bait. Some Australian Salmon are sold fresh for human consumption.

Australian Salmon is the second most commonly captured fish species by recreational fishers (Lyle 2005, Lyle et al. 2009, Lyle et al. 2014b), who mainly use line fishing methods to target the species.

<b>FISHING METHODS</b>	Mainly beach seine, also purse seine and gillnet. Line for recreational.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence, Beach seine licence).</li> <li>• Species licence (Australian Salmon licence) – 8 issued, 1 commonly used.</li> <li>• Spatial and temporal area closures for Australian Salmon licence.</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Trip limit of 500 kg for operators with Scalefish licences but no Australian Salmon licence.</li> <li>• Possession limit of 30 and bag limit of 15 individuals for recreational fishers.</li> <li>• Minimum size (200 mm TL).</li> <li>• Total commercial catch trigger of 435 t.</li> </ul>
<b>MAIN MARKET</b>	Local and interstate.

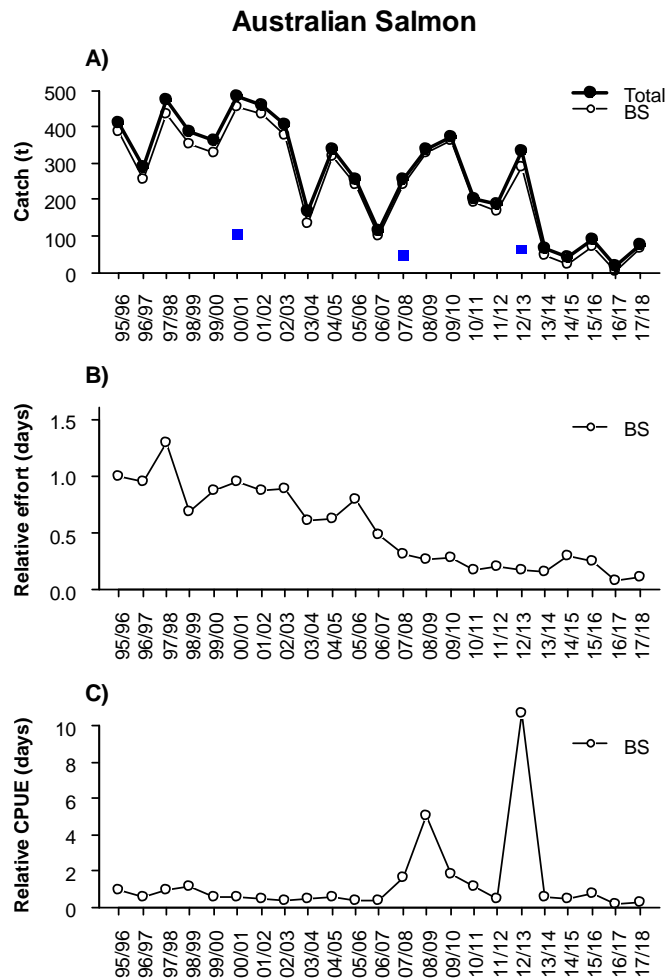
## Current assessment

### **Catch, effort and CPUE**

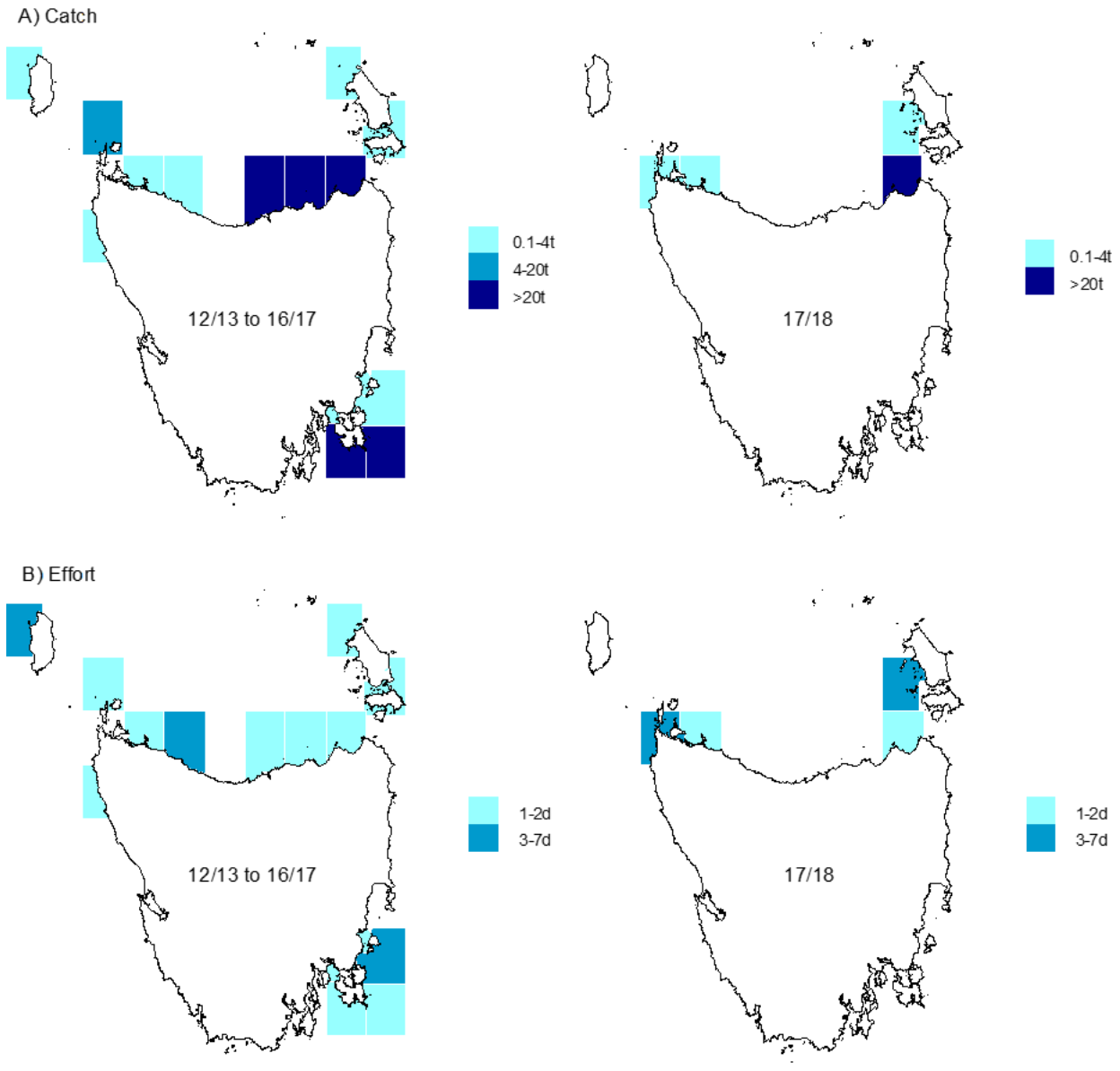
Following a large catch in 2012/13 (331 t) landings for the last few years have been low, with 76.1 t landed in 2017/18 (Fig. 4.1A). The low catch in recent years has been due to a dramatic decline in the landings by beach seine fishers that have historically landed the majority of the catch (Fig. 4.1A). The majority of the catch in 2017/18 was taken using beach seine, followed by gillnetting. Catches in 2017/18 came from the northwest and northeast coasts, and from around Flinders Island (Fig. 4.2). Effort and catch rates remain low (Fig. 4.1B and 4.1C), however it should be noted that catch rate is influenced by the skewed nature of the data (i.e. the majority of catches are small but the total catch is influenced by a small number of extremely large catches). In addition, catch rate is not a particularly sensitive indicator of stock condition for schooling species such as Australian Salmon, especially if search time is not taken into account.

### **Ecological Risk Assessment**

In the 2012/13 ecological risk assessment (ERA) of the Tasmanian scalefish fishery, beach seining was considered a very low risk activity with regard to Australian Salmon. Beach seining was also considered a low risk activity with regards to non-retained species as bycatch is usually released alive and 'herded' not 'meshed/gilled' and a very low risk in regards to the general ecosystem (Bell et al. 2016). Given the low catch and effort since 2012/13 there is no evidence to suggest this level of risk has increased.



**Figure 4.1** A) Annual commercial catch (t) by gear, and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. BS=beach seine.



**Figure 4.2** (A) Annual Australian Salmon catches (t) and (B) effort (days) for beach seine, gillnet, small mesh net and purse seine fishing methods by fishing block averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Given that beach seine catch rates are not a sensitive indicator of stock status due to the schooling behaviour of the species, the biomass performance indicators (based on CPUE and CPUE trends) were not calculated for Australian Salmon.

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Total commercial catch >435 t	No	
	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (462.1 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (254.2 t)	Yes	↓ 178.1 t (70.1%)
	• Catch variation from the previous year above the greatest inter-annual increase from the reference period (188.7 t)	No	
	• Catch variation from the previous year above the greatest inter-annual decrease from the reference period (240.0 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (105.2 t)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate (13.8% in 2007/08)	Yes	Latest estimate (2012/13) 19.0%
<b>Stock stress</b>	• Significant change in the size/age composition	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

**SUSTAINABLE**

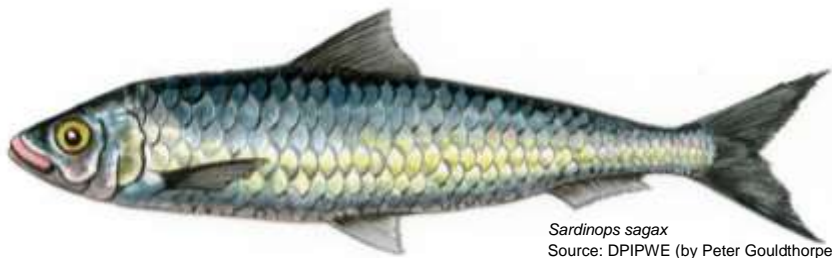
Although the third lowest catch reference point was breached as a result of the low landings in 2017/18, annual production of Australian Salmon is strongly linked to the fishing practices of a single beach seine operator who is responsible for the overwhelming majority of landings. Catch and effort for this sector increased slightly in 2017/18 relative to 2016/17 when they reached a historically low level, reflecting low market demand rather than changes in abundance.

Eastern Australian Salmon represent a single well-mixed stock along southeast Australia and the Tasmanian fishery catches mostly sub-adults. There has been little change in size and age composition evident in other states (NSW and VIC) suggesting the species is not overexploited. The eastern Australian biological stock is sustainably fished as a whole and the current level of commercial and recreational fishing pressure in Tasmania is well below historical levels and unlikely to cause the biological stock to become recruitment impaired.

# 5. Australian Sardine

## *Sardinops sagax*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
<p>The fishery is in a developmental phase in Tasmania, with low catches reported to date. The species is classified as not overfished nor subject to overfishing by ABARES for 2017 (Patterson et al. 2018), while all Australian stocks are similarly classified as sustainable in the 2016 Status of Australian Fish Stocks. The current level of fishing pressure in Tasmania is unlikely to cause the biological stock to become recruitment impaired.</p>	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



### Species biology

Parameter	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Lower reaches of estuaries to continental shelf waters.</li> </ul>	Paxton and Hanley (1989)
Distribution	<ul style="list-style-type: none"> <li>Found throughout southern Australia from Queensland (Rockhampton) to Western Australia (Shark Bay) and northern Tasmania.</li> </ul>	Izzo et al. (2017)
Diet	<ul style="list-style-type: none"> <li>Phytoplankton and planktonic crustaceans.</li> </ul>	Stewart et al. (2011)
Movement and stock structure	<ul style="list-style-type: none"> <li>Four separate biological stocks are considered present in Australian waters: south-west coast of Western Australia, Great Australian Bight–Spencer Gulf, Bass Strait–Port Phillip Bay (including Tasmania and likely southern New South Wales) and eastern Australia. Four stocks were considered in the 2016 Status of Australian Fish Stocks reporting (Eastern Australia, Western Australia west coast, Western Australia south coast and Southern Australia). The Tasmanian fishery is likely to be part of the Bass Strait–Port Phillip Bay stock.</li> </ul>	Izzo et al. (2017) Ward et al. (2017)
Natural mortality	<ul style="list-style-type: none"> <li><math>M</math> of 0.43 calculated for stock on south coast of Western Australia.</li> <li>Mean <math>M</math> estimated as 0.52 for populations of the west coast of the United States.</li> </ul>	Fletcher (1995) Zwolinski et al. (2013)

Maximum age	<ul style="list-style-type: none"> <li>8 years</li> </ul>	Stewart et al. (2010)								
Growth	<ul style="list-style-type: none"> <li>Maximum length: 23.0 cm</li> <li>Maximum weight: 0.14 kg</li> <li>Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math> where <math>L</math> is the fork length (mm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</li> </ul> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>K</math></th> <th><math>t_{0\infty}</math></th> </tr> </thead> <tbody> <tr> <td>Combined</td> <td>236.1</td> <td>0.37</td> <td>-0.28</td> </tr> </tbody> </table>	Sex	$L_{\infty}$	$K$	$t_{0\infty}$	Combined	236.1	0.37	-0.28	<p>Neira et al. (1999)</p> <p>Stewart et al. (2010)</p>
Sex	$L_{\infty}$	$K$	$t_{0\infty}$							
Combined	236.1	0.37	-0.28							
Maturity	<ul style="list-style-type: none"> <li>Size-at-50% maturity: 14.5 cm</li> <li>Age at 50% maturity: 1–2 years</li> </ul>	<p>Stewart et al. (2010)</p> <p>Ward and Grammer (2016)</p>								
Spawning	<ul style="list-style-type: none"> <li>Main spawning area for the eastern Australian stock occurs off southern Queensland and northern NSW during late winter and early spring. Spawning has also been recorded to occur between Tasmania and southern Victoria during summer.</li> <li>The relationship between batch fecundity and weight is linear with <math>F = -12042 + 452.69W</math>, where <math>F</math> is the fecundity (in number of eggs) and <math>W</math> is the female weight (g).</li> </ul>	<p>Ward et al. (2015)</p> <p>Lo et al. (2005)</p>								
Early life history	<ul style="list-style-type: none"> <li>Juveniles (4–6 cm fork length) have been recorded in Port Phillip Bay, Victoria, between February and May, and within bays elsewhere in south-eastern Australia in winter through to summer months.</li> </ul>	<p>Hoedt and Dimmlich (1995)</p> <p>Neira et al. (1999)</p>								
Post release survival	<ul style="list-style-type: none"> <li>Unknown, but likely low.</li> </ul>									

## Background

Australian Sardine have had a relatively long history of commercial exploitation in mainland state and Commonwealth waters, but have been seldom harvested in Tasmanian waters. The majority of the total Australian Sardine catch is derived from state waters. Total landings in Victoria for 2015/16 were 1, 682 t, while catches of Australian Sardine in Commonwealth waters totalled 114 t in 2015/16 (Patterson et al. 2017). There is also a significant fishery for the species in South Australia, where most of the catch is used as feed for the Southern Bluefin Tuna sea ranching operations.

The fishery for Australian sardine in Tasmania is likely based on the Bass Strait–Port Phillip Bay stock. There is evidence to suggest that the species may be present in large quantities in Tasmanian waters in some years. Ichthyoplankton surveys conducted during 2014 suggested that a spawning biomass of approximately 10,962 t was present off northern Tasmania and in Bass Strait during summer (Ward et al. 2015). It was also noted that the actual spawning area was likely to be larger than surveyed (possibly extending further into Bass Strait and off northern Tasmania), implying that the estimate may be negatively biased. In relation to recommended harvest rates, Smith et al. (2015a) has recommended a harvest rate of 24–27% for Australian

Sardine. However, given uncertainty in the biomass estimate and taking into account that very little is known about the dynamics of the sardine stocks off south-eastern Australia, a more conservative harvest rate (around 20%) has been recommended. Applying a harvest rate of 20% to the 2014 biomass estimate suggests that the stock (some of which occurs outside of Tasmanian waters) could support catches in the range of 1,600–3,000 tonnes per year. It should be noted that this is likely to be a shared stock and a fishery already operates in Victoria (annual catches > 1,500 t in recent years), and accordingly, collaborative management will be needed.

In 2015 the Tasmanian Government released a framework to support developmental fishing program for Australian Sardine by identifying two sectors based on scale of operations; a large-scale sector (capable of handling catches in excess of several tonnes at a time) and a small-scale sector taking limited catches (< one tonne maximum at a time) (DPIPWE 2015). Managed as a developmental fishery, a total annual catch limit of 600 tonnes was applied to the large-scale sector (with a maximum of 300 tonnes to be taken from Bass Strait and 300 tonnes from the east coast) and a maximum of 50 tonnes applied to the small-scale sector. Two large-scale and two small-scale developmental permits were issued initially.

Australian Sardine is not a significant recreational species in Tasmania (Lyle et al. 2014a).

<b>FISHING METHODS</b>	Mainly purse seine, also beach seine.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence, beach seine licence and purse seine licence).</li> <li>• Developmental Australian Sardine permit.</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• A catch limit for the large-scale commercial sector of 600 tonnes for the developmental fishery, with a maximum of 300 t to be taken from Bass Strait and 300 t from the east coast. <ul style="list-style-type: none"> <li>○ A 'soft cap limit' of 60 t per month for the large-scale commercial sector for each area described above.</li> </ul> </li> <li>• A total limit of 50 t for the small-scale sector during the developmental process.</li> <li>• A trip limit of 10 kg of Australian Sardine for general Scalefish Licence holders.</li> <li>• Spatial restrictions.</li> <li>• Possession limit of 100 and bag limit of 50 individuals for recreational fishers (all baitfish species combined).</li> </ul>
<b>MAIN MARKET</b>	Local and interstate.

## Current assessment

### Biological characteristics

In 2017, a sampling program was implemented to obtain baseline biological information, including size and age composition of the catch and insights into spawning dynamics (in both space and time) of Australian Sardine in Tasmanian waters. In 2017, a total of 201 fish were sampled (including 152 individuals from the north-east and 49 individuals from the north coast) during January and February, with approximately 20–50 fish sampled from each commercial purse seine catch landed.

### *Size composition*

Sampled fish ranged from 125 mm to 216 mm fork length, with a modal length of 160 mm (Fig. 5.1). Samples from the north coast were on average larger than those caught for the north-east (206±1 mm vs. 162±1 mm, respectively). Few differences in length structure were observed between males and females, although there were proportionally more females in the larger length classes (Fig. 5.1).

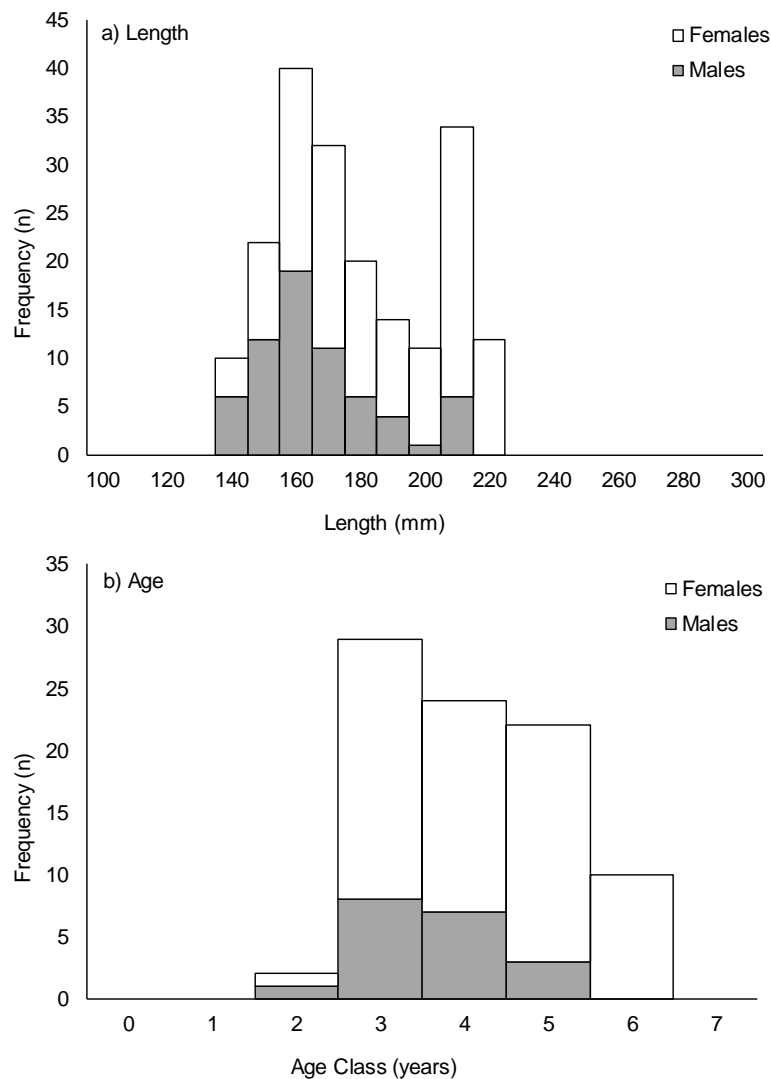
### *Age composition*

In the absence of a validated ageing protocol for the species in Tasmanian waters, ages of sampled individuals were estimated using the linear otolith weight-age relationship of Stewart et al. (2010) that is applied to ageing the species in NSW:

$$\text{Age (years)} = 1.95 \times \text{average otolith weight (mg)} + 0.43$$

The resulting age frequency histogram from the fishery-dependent samples revealed that ages ranged from 2–6 years, with a modal age of 3 years (Fig. 5.1). This is largely consistent with observations of the stock elsewhere across its distribution. For example, Stewart et al. (2010) observed an age range of 0–5 years off the coast of NSW, with the majority of individuals being between 1 and 4 years old.

A large overlap in age structure was evident between males and females, however, as with length, there were proportionally more females in the older age classes (Fig. 5.1).



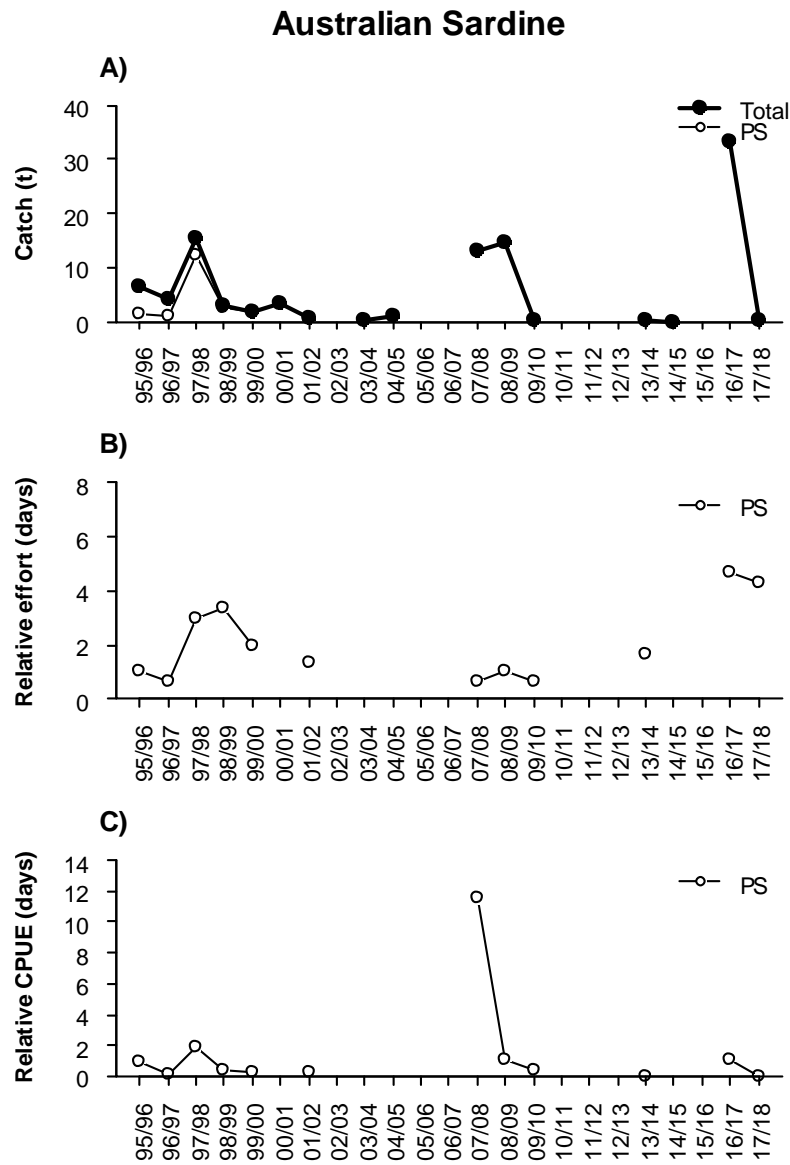
**Figure 5.1** Length (top) and age (bottom) frequency histograms for Australian Sardine sampled from the commercial purse seine catch in 2017.

### Catch, effort and CPUE

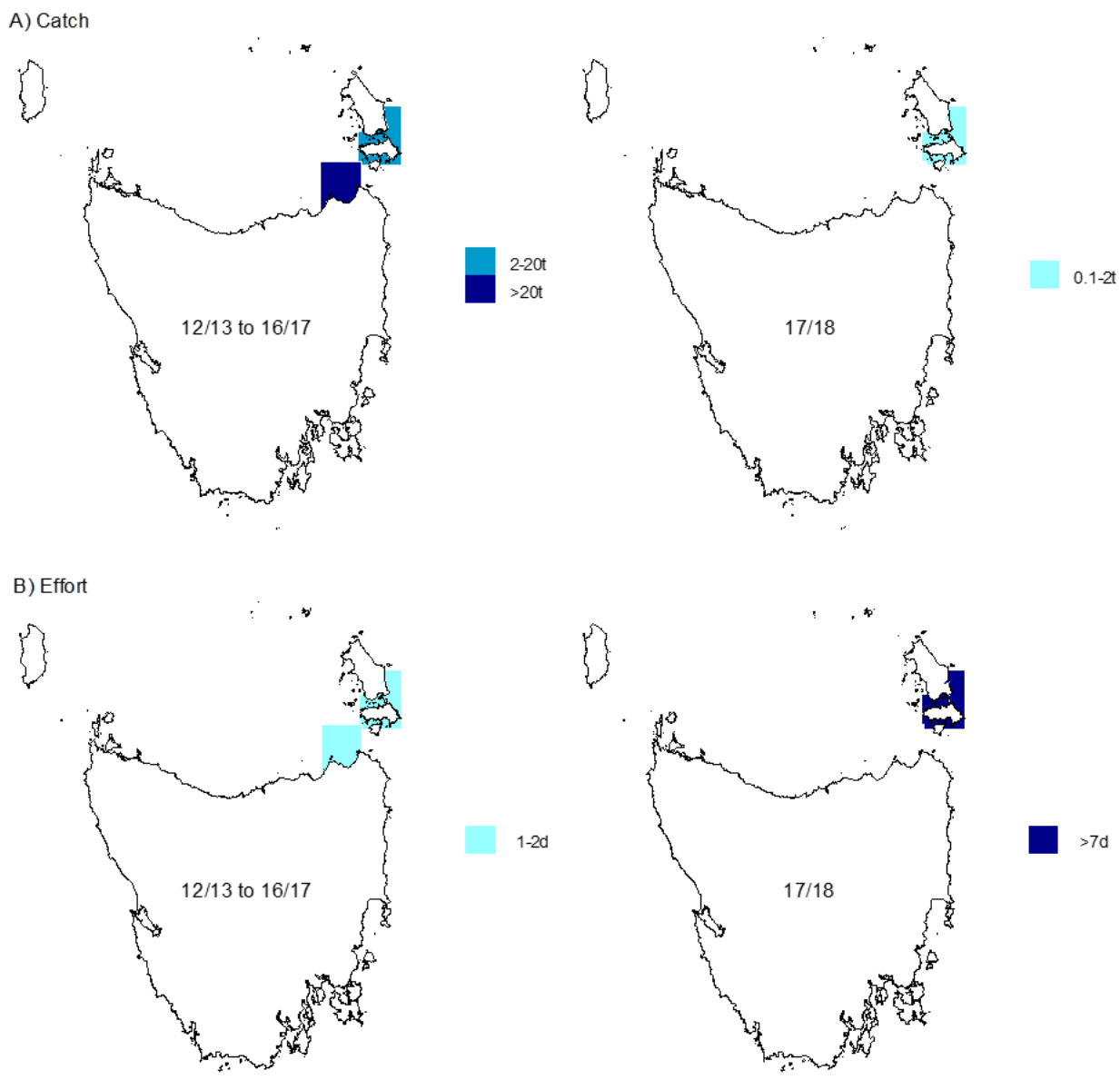
The 2017/18 total commercial catch of Australian Sardine in Tasmanian waters was 0.1 t (Fig. 5.2). Historically, the species has constituted a minor and sporadic component of the Tasmania scalefish fishery, with peak catches of 15.4 t recorded in 1997/98, 14.5 t in 2008/09 and 33.3 t in 2016/17, interspersed among years of little or no catch. The earlier peak catches largely reflect incidental take of Australian Sardine by fishers targeting other small pelagic fishes (e.g. redbait). Targeted fishing for the species under the developmental fishery permit commenced in 2016/17, with fishing activity based around the north and northeast coasts (Fig. 5.3). All Australian Sardine landed in 2016/17 and 2017/18 were caught by purse seine.

### Ecological Risk Assessment

In the 2012/13 ecological risk assessment (ERA) of the Tasmanian scalefish fishery, purse seining was considered a negligible risk activity to Australian Sardine due to the small quantities taken in the fishery (Bell et al. 2016). Since 2012/13 there has been minimal catch of Australian Sardine, and the risk is still considered negligible at currently levels of fishing activity. If catches were to increase markedly in the future the risks would need to be reassessed, including the potential for interactions with marine mammals.



**Figure 5.2** A) Annual commercial catch (t) by gear, and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. BS=beach seine.



**Figure 5.3** (A) Australian Sardine catches (t) and (B) effort (days) for beach seine and purse seine by fishing block averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right). Note: only minor landings of Australian Sardine occurred in the previous 10 years.

## Reference points

As this fishery is a developmental fishery in Tasmania waters, a full suite of reference points is yet to be defined. The development phase reference points (see management methods above) were not breached for 2017/18.

## Stock status

**SUSTAINABLE**

Overall, catches of Australian Sardine in Tasmania waters reflect only a minor proportion of the Bass Strait–Port Phillip Bay stock, with surveys conducted in 2014 indicating a spawning biomass of approximately 10,962 t off northern Tasmania.

Since 2008, Australian Sardine populations in the Commonwealth Small Pelagic fishery have been considered to be not overfished nor subject to overfishing (Patterson et al. 2018), and all four Australian stocks considered during the 2016 Status of Australian Fish Stocks (Eastern Australia, Western Australia west coast, Western Australia south coast and Southern Australia) were classified as sustainable (Ward et al. 2017). Given that current levels of effort are unlikely to result in recruitment overfishing, this ranking has been applied to the Tasmanian fishery.

# 6. Barracouta

## *Thyrsites atun*

<b>STOCK STATUS</b>	<b>UNDEFINED</b>
Catches of Barracouta have declined steadily since the mid-2000s due to a decrease in targeted effort as a result of a lack of market demand. Catches and catch rates are not considered indicative of stock status and there is insufficient information to confidently classify the stock.	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Thyrsites atun*  
Source: DPIPWE (bv Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Open water and coastal bays. Down to 550 m depth.</li> </ul>	Edgar (2008) Kailola et al. (1993)
Distribution	<ul style="list-style-type: none"> <li>From midwestern Australia to southern Queensland, and around Tasmania. Also widely distributed in the southern hemisphere in temperate latitudes.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Pelagic crustaceans, cephalopods, fishes (e.g. anchovy and Jack Mackerel).</li> </ul>	Nakamura and Parin (1993)
Movement and stock structure	<ul style="list-style-type: none"> <li>Schooling fish.</li> <li>Some stocks undertake annual migrations lasting 6–9 months and covering several hundreds of km.</li> <li>Also moves through the water column from 200 m depth to the surface.</li> <li>At least 5 stocks: 3 in south-eastern waters, 1 in South Australia and 1 in Western Australia.</li> </ul>	Paul (2000) Kailola et al. (1993) Blackburn and Gartner (1954)
Natural mortality	<ul style="list-style-type: none"> <li><math>M = 0.3</math></li> </ul>	Hurst et al. (2012)
Maximum age	<ul style="list-style-type: none"> <li>At least 10 years, potentially up to 15 years.</li> </ul>	Kailola et al. (1993) Hurst et al. (2012)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 1.4 m</li> <li>Maximum weight: 6 kg</li> <li>Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math></li> </ul>	Edgar (2008) Nakamura and Parin (1993)

	<p>where <math>L</math> is the fork length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</p> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_{0\infty}</math></th> </tr> </thead> <tbody> <tr> <td>Combined</td> <td>91</td> <td>0.42</td> <td>-0.25</td> </tr> </tbody> </table> <ul style="list-style-type: none"> <li>Length-weight relationship was estimated at <math>W = 0.1064 L^{2.2385}</math> for females and males combined where <math>W</math> is weight (g) and <math>L</math> is the fork length (cm).</li> </ul>	Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Combined	91	0.42	-0.25	<p>Grant et al. (1978)</p> <p>Blackburn (1960)</p>
Sex	$L_{\infty}$	$k$	$t_{0\infty}$							
Combined	91	0.42	-0.25							
Maturity	<ul style="list-style-type: none"> <li>Sexual maturity at about 50–60 cm FL and about 2–3 years of age.</li> </ul>	Hurst et al. (2012)								
Spawning	<ul style="list-style-type: none"> <li>October to March in Tasmania.</li> </ul>	Kailola et al. (1993)								
Early life history	<ul style="list-style-type: none"> <li>Little data. Eggs are pelagic and juveniles inhabit sheltered waters of southern bays and estuaries.</li> </ul>	<p>Kailola et al. (1993)</p> <p>Hurst et al. (2012)</p>								
Gillnet post release survival	<ul style="list-style-type: none"> <li>NA</li> </ul>									

## Background

Barracouta were the subject of a large commercial trolling fishery in Tasmania in the 1960s and 1970s, with catches ranging from 600–1,600 t (Kailola et al. 1993). Demand for Barracouta, however, declined in the mid-1970s and there is now little commercial fishing for the species. Barracouta abundance also fluctuates greatly in State waters annually.

<b>FISHING METHODS</b>	Mostly troll and hand-line.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>Gear licence (Scalefish fishing licence).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>Possession limit of 30 and bag limit of 15 individuals for recreational fishers.</li> </ul>
<b>MAIN MARKET</b>	Local.

## Current assessment

### Catch, effort and CPUE

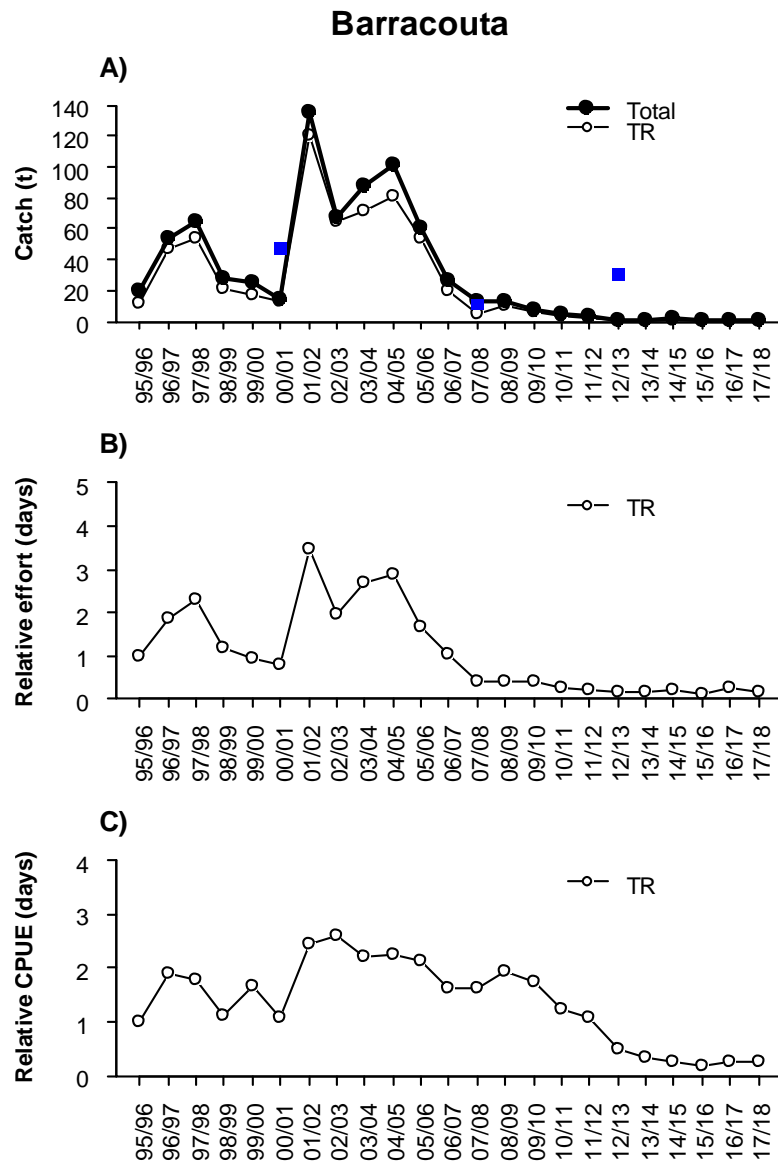
Total commercial catches peaked in the early 2000s with maximum of 136 t but gradually declined from 101 t in 2004/05 to a historical low of 0.4 t in 2015/16 (Fig. 6.1A). Commercial catches in 2017/18 were 0.9 t (Fig. 6.1A). Trolling and small mesh nets were the main fishing methods used to target Barracouta in 2017/18. Catches and fishing effort were traditionally concentrated off southern Tasmania (Emery et al. 2017); however, over the last few fishing seasons have been largely concentrated off the north coast (Fig. 6.2).

After the peak in the early 2000s, effort declined and has stabilised since 2007/08 at a low level (Fig. 6.1B). Catch rates have been relatively stable over the most recent fishing years (Fig. 6.1C); however, it is likely that fishers utilising fishing gears historically used to target Barracouta are now targeting other species and as a result, catch-based statistics are unlikely to be a reliable indicator of abundance.

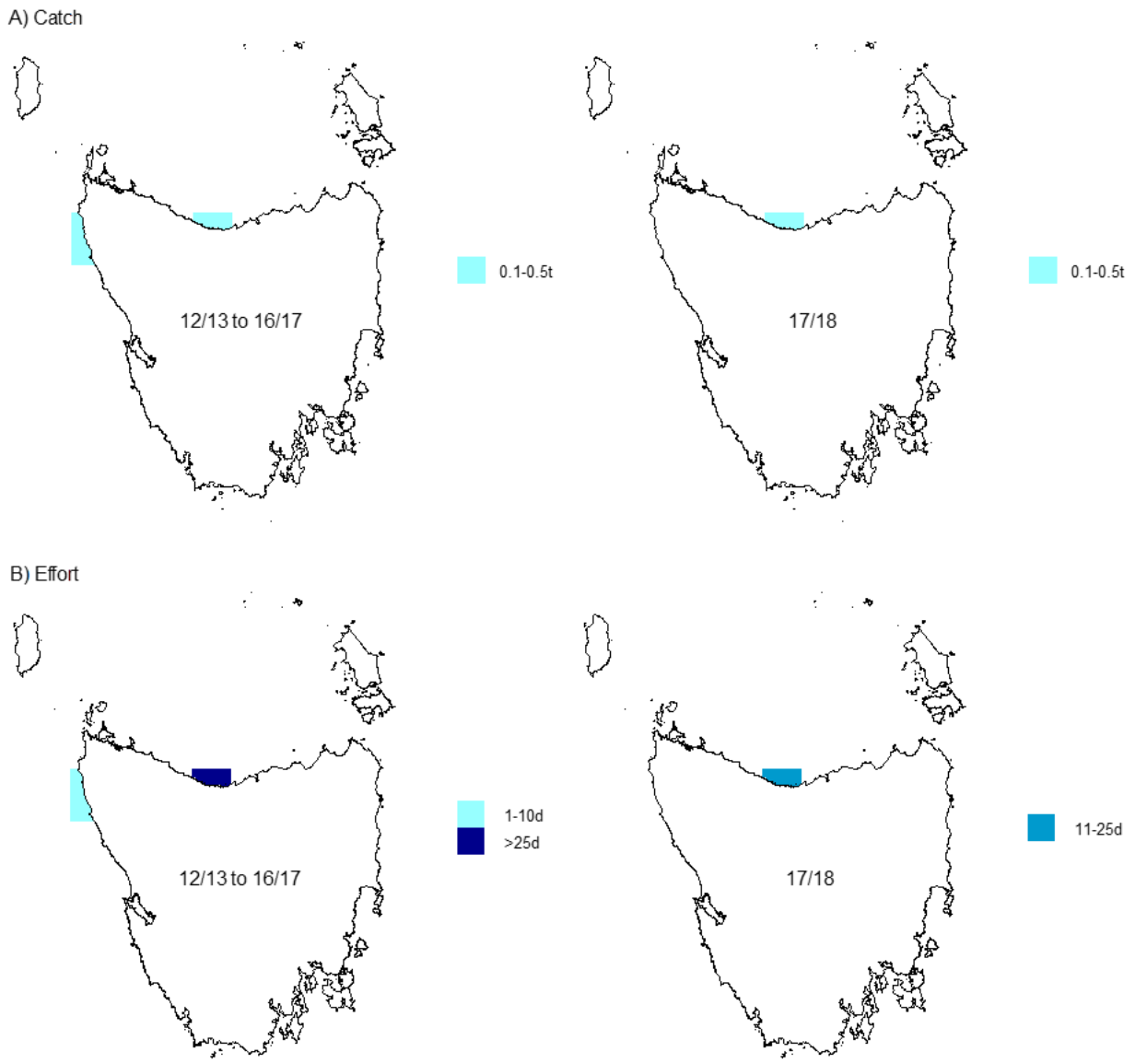
Barracouta are targeted and also taken as by-product by the recreational sector. Catches were estimated at 46.9 t in 2000/01 (Lyle 2005), 10.8 t in 2007/08 (Lyle et al. 2009) and 31 t in 2012/13 (Lyle et al. 2014b), the latest estimate being considerably greater than the commercial harvest in recent years (Fig. 6.1A).

### **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian scalefish fishery, trolling was considered a negligible risk activity with regard to Barracouta, by-product species, non-retained species and the general environment (Bell et al. 2016). Since 2012/13 there has been minimal catch of Barracouta and the risk is still considered negligible.



**Figure 6.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. TR=troll.



**Figure 6.2** (A) Barracouta catches (t) and (B) effort (days) for troll and hand-line fishing methods by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (87.5 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (25.0 t)	Yes	↓ 24.1 t (96.3%)
	• Latest recreational catch estimate > recreational catch estimate from the reference period (46.9 t)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate (43.8% in 2007/08)	Yes	Latest estimate (2012/13): ↑ 96.6%
<b>Biomass</b>	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0091)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

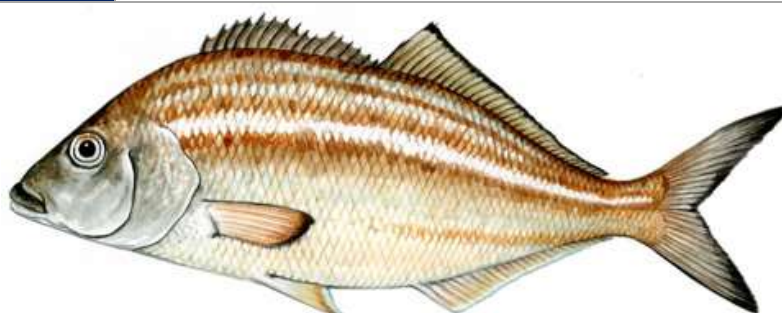
UNDEFINED

Historically, the population of Barracouta has undergone large fluctuations in size and availability, possibly linked to recruit variability and environmental factors. Catches of Barracouta in Tasmanian waters have been declining steadily since the mid-2000s due to a decrease in targeted effort as a result of a lack of market opportunities. The increase in recreational catch proportion mainly reflects the sharp fall in commercial landings rather than increased targeting by recreational fishers. Discards of Barracouta in the South East Trawl Fishery sector of the Southern and Eastern Scalefish and Shark Fishery (SESSF) have previously estimated to be around 12% of the total discarded non-quota catch (Knuckey 2006), equating to roughly 1,356–1,920 t annually. The fate of such discards is unknown. While this suggests that Barracouta may be locally abundant within the SESSF, due to a lack of targeting commercial catches and catch rates in Tasmania are not considered indicative of stock status. As such, there is insufficient information to confidently classify this stock.

# 7. Bastard Trumpeter

## *Latridopsis forsteri*

<b>STOCK STATUS</b>	<b>DEPLETED</b>
Trends in commercial catch suggest record low population levels and that the species is recruitment overfished. The current minimum legal size limit is below the size of maturity. Current low levels of fishing pressure may be too high to allow stock to recover.	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Latridopsis forsteri*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Exposed reefs and sandy bottom down to 160 m depth.</li> </ul>	May and Maxwell (1986) Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>Sydney (New South Wales) to southern South Australia, Tasmania, southern New Zealand.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Small fish, invertebrates.</li> </ul>	Edgar (1997)
Movement and stock structure	<ul style="list-style-type: none"> <li>Schooling fish, usually in small numbers. Large individuals occur in deeper waters.</li> <li>Juveniles tend to remain associated with areas of reef for periods of time.</li> <li>No information on the stock structure.</li> </ul>	Edgar et al. (2004) Gomon et al. (2008)
Natural mortality	<ul style="list-style-type: none"> <li>Undetermined.</li> </ul>	
Maximum age	<ul style="list-style-type: none"> <li>Up to 20 years.</li> </ul>	Murphy and Lyle (1999)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 65 cm</li> <li>Maximum weight: 4 kg</li> <li>Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math></li> </ul>	Edgar (2008) Gomon et al. (2008)

	<p>where <math>L</math> is the fork length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</p> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_{0\infty}</math></th> </tr> </thead> <tbody> <tr> <td>Combined</td> <td>59.4</td> <td>0.144</td> <td>-2.9</td> </tr> </tbody> </table>	Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Combined	59.4	0.144	-2.9	Murphy and Lyle (1999)
Sex	$L_{\infty}$	$k$	$t_{0\infty}$							
Combined	59.4	0.144	-2.9							
Maturity	<ul style="list-style-type: none"> <li>Sexual maturity at sizes &gt; 45 cm and ages &gt; 4 years.</li> </ul>	Murphy and Lyle (1999)								
Spawning	<ul style="list-style-type: none"> <li>Believed to spawn in late winter.</li> </ul>	Murphy and Lyle (1999)								
Early life history	<ul style="list-style-type: none"> <li>Small juveniles settle from the plankton on reefs in large numbers at intervals of several years.</li> </ul>	Edgar (2008)								
Recruitment	<ul style="list-style-type: none"> <li>Variable. No-stock recruitment relationship established.</li> </ul>									
Gillnet post release survival	<ul style="list-style-type: none"> <li>High: 95 – 83% depending on gillnet soak duration.</li> </ul>	Lyle et al. (2014a)								

## Background

Bastard Trumpeter was one of the first fish species to have been commercially exploited in Tasmania. Their apparent abundance around reefs close to newly established Hobart meant that they were an important source of seafood for the fledgling colony. Their exploitation was further aided by the relative ease at which they could be caught using gillnets set within accessible shallow inshore reefs.

Bastard Trumpeter are taken today almost exclusively by gillnet. Bastard Trumpeter have also long been recognised as an important fish for recreational fishers. The species resides on inshore reefs until about 4–5 years of age (and approximately 50 cm long) before moving offshore into deeper water as they approach maturity, apparently remaining in that habitat for the remainder of their lives (Harries and Lake 1985, Murphy and Lyle 1999). Hence, both commercial and recreational fisheries are based almost entirely on juvenile fish.

<b>FISHING METHODS</b>	Gillnet.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>Gear licence (Scalefish fishing licence, rock lobster licence).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>Possession limit of 10 and bag limit of 5 individuals for recreational fishers.</li> <li>Trip limit 200 kg for commercial scalefish licence holders.</li> <li>Trip limit of 30 fish for commercial rock lobster licence holders.</li> <li>Legal minimum length (380 mm TL).</li> </ul>
<b>MAIN MARKET</b>	Local.

## **Current assessment**

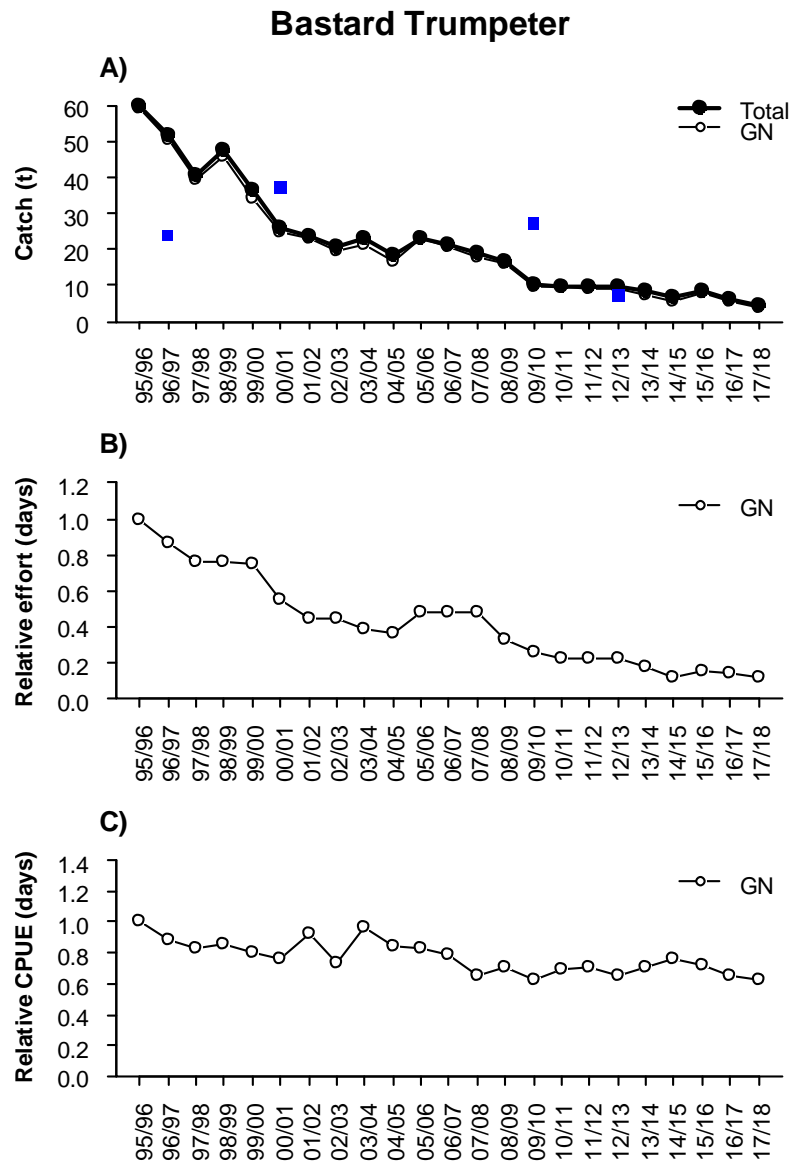
### **Catch, effort and CPUE**

Bastard Trumpeter catches have been declining steadily since the mid-1990s and have been < 10 t since 2010/11, with a record low of 4.3 t landed in 2017/18 (Fig. 7.1A). Bastard Trumpeter are taken almost exclusively by gillnet from inshore waters off the east, south and west coasts (Fig. 7.2). Catches and effort in 2017/18 contracted spatially relative to previous years, being concentrated primarily around the southeast and southwest coasts (Fig. 7.2). Bastard Trumpeter have been predominantly taken by recreational fishers in recent years, although the estimated catch in 2012/13 was also a historic low of 9.8 t (Lyle et al. 2014b) and is considerably lower than the 2010 estimate of 27.3 t (Lyle and Tracey 2012).

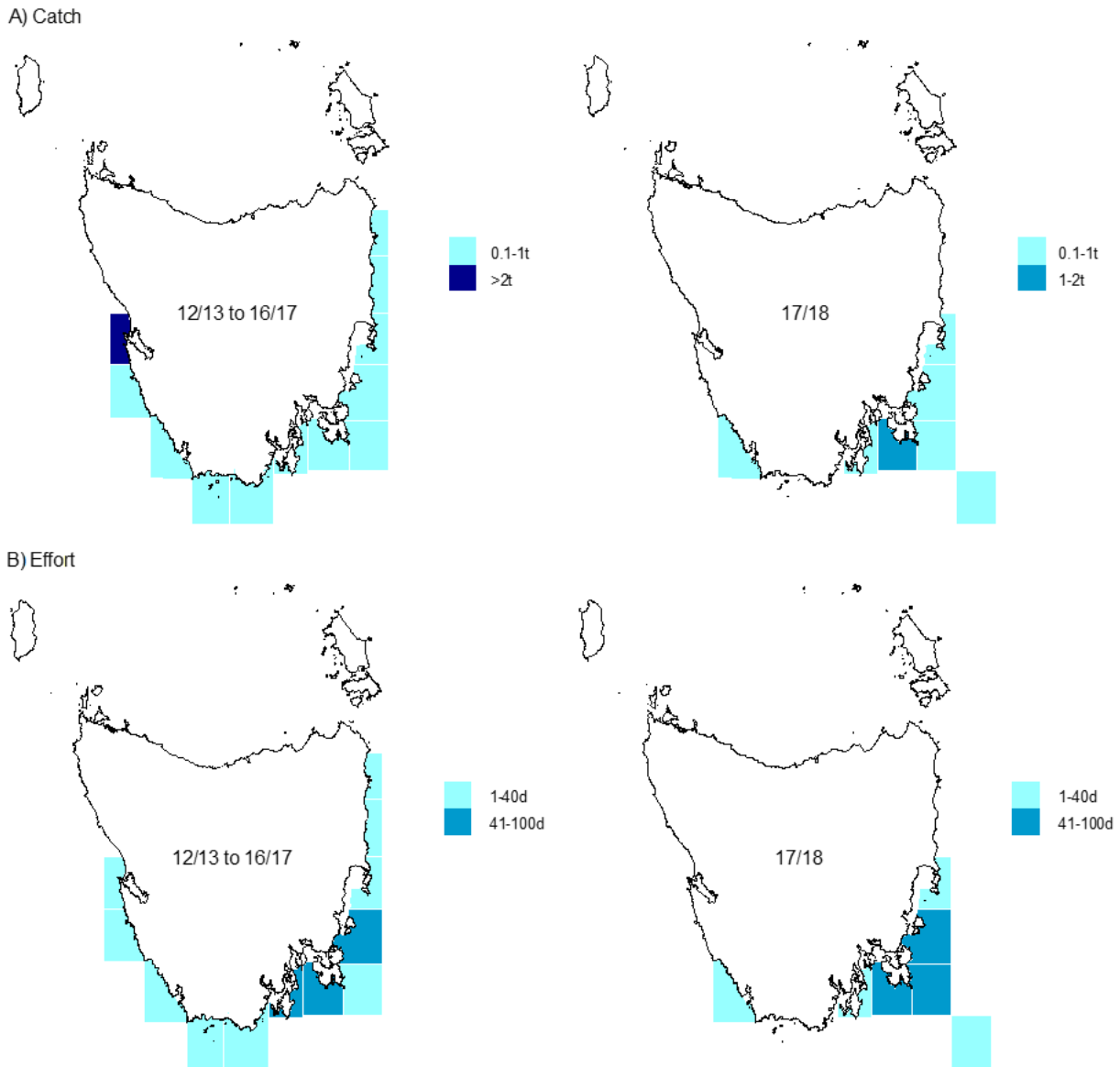
Commercial gillnet effort has followed a similar downward trend to catches since the mid-1990s (Fig. 7.1B). Daily catch rates have remained relatively stable since 2006/07 at a reduced level (Fig. 7.1C). This lack of obvious trend, despite the sharp decrease in catches, presumably reflects the fact that Bastard Trumpeter are taken primarily taken as by-product rather than as a target species. This is exacerbated by the fact that the majority of gillnetting effort now targets Banded Morwong with 140 mm mesh sizes, selecting only the largest Bastard Trumpeter. Previously, a larger proportion of fishers used smaller mesh sizes (<114 mm) to target Bastard Trumpeter and/or Blue Warehou.

### **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian scalefish fishery, gillnetting was considered a medium risk activity with regard to Bastard Trumpeter, due to the population status for this species being uncertain, the potential for growth overfishing given the fishery is based on juveniles, and the possibility that current fishing pressure is too high (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.



**Figure 7.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on days fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. GN=gillnet. Data includes Australian Fisheries Management Authority (AFMA) catch in State waters.



**Figure 7.2** (A) Bastard Trumpeter catches (t) and (B) effort (days) for gillnet fishing by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right). Data includes Australian Fisheries Management Authority (AFMA) catch taken in State waters.

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (47.7 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (21.3 t)	Yes	↓ 17.0 t (79.7%)
	• Catch variation from the previous year above the greatest inter-annual increase from the reference period (7 t)	No	
	• Catch variation from the previous year above the greatest inter-annual decrease from the reference period (-11.3 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (24 t)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate (73.6% in 2010)	No	
<b>Biomass</b>	• CPUE < 3 <sup>rd</sup> lowest CPUE value from the reference period (0.0108 t/days fished)	Yes	0.0031 t/day fished (↓ 28.7%)
	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0011)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

**DEPLETED**

Due to the low catches in 2017/18, the lowest catch and catch rate reference points were breached. As Bastard Trumpeter is a by-product species, catch rather than catch rate may be a better indicator of abundance/availability for the species. Consequently, the trend in commercial production suggests that current inshore populations are at historically low levels. In accordance with this observation, industry and recreational representatives have expressed concerns about the scarcity of the species in recent years (Emery et al. 2017), although a lack of market demand for Bastard Trumpeter appears to be a factor influencing landings. On board observations suggest that legal sized Bastard Trumpeter are sometimes discarded by Banded Morwong fishers and research suggests post release survivability is high (Lyle et al. 2014b). Given the majority of gillnet effort is now targeted at Banded Morwong using larger mesh sizes than those used historically to target Bastard Trumpeter, both catch and catch rate trends may be confounded and not representative of population status. Nevertheless, fishing practices are likely to have remained fairly consistent in recent years (2007/08 – present) so the low, stable catch rate trend and declining catches are likely to be representative of a population that has not substantially rebuilt despite significant reduction in both commercial and recreational gillnet effort.

The Tasmanian Bastard Trumpeter fishery is based almost entirely on juveniles. As fish grow, they appear to move offshore and are rarely caught. No information is available on the adult portion of the population but it is clear that fishing pressure exerted on those individuals that evade the inshore fishery is very low (by-catch in shark nets, trawl, Danish seine or deep-water fish traps used by the Commonwealth SESSF appears to be negligible). The species exhibits high recruitment variability resulting in short-term variation in catches, which has been a feature of this fishery over the past century (Harries and Croome 1989). Anecdotal reports and low inshore catches suggest that recruitment has been low in recent years and limited length frequency data for 2011 and 2012 indicate a reduction in the number of smaller-sized individuals in the fishery relative to the late 1990s (Emery et al. 2016). Studies have demonstrated significantly higher abundances of Bastard Trumpeter in unfished marine reserves relative to fished sites around Tasmania (Edgar and Barrett 1999), which in combination with the fact that commercial and recreational fisheries are based entirely on juveniles, suggests that recruitment as well as growth overfishing may be occurring.

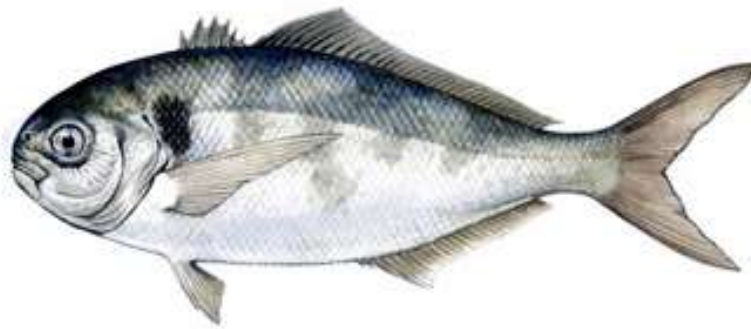
It is worth noting that the stabilisation of the catch since 2009/10 corresponds to the introduction of several management measures for the species (increase in the minimum legal size, introduction of commercial trip limits and reduction in recreational bag and possession limits). However, the current minimum size limit of 38 cm TL is still well below the size at maturity (>45 cm FL, Murphy and Lyle 1999). While there has been discussion about increasing the minimum size limit to recover stocks this would effectively close down the current commercial and recreational fisheries for the species. Further reductions in the recreational bag limit for this species were introduced in 2015.

Given the continued reduction in catches, the current rating of medium risk from gillnetting in the ERA and the current minimum legal size limit below the size of maturity, Bastard Trumpeter are classified as depleted.

# 8. Blue Warehouse

## *Seriolella brama*

<b>STOCK STATUS</b>	<b>DEPLETED</b>
<p>This is predominately a Commonwealth-managed species that is classified as overfished by ABARES in 2017 (Patterson et al. 2018). It is sporadically abundant in Tasmanian waters. Despite a reduction in Total Allowable Catch (TAC) for the Commonwealth fishery to 118 t and the initiation of a rebuilding strategy in 2008, stocks have shown little evidence of recovery.</p>	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery/Southern and Eastern Scalefish and Shark Fishery (Commonwealth)
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Seriolella brama*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Inshore reefs/harbours and open water down to 400 m depth.</li> </ul>	Edgar (2008) Smith (1994)
Distribution	<ul style="list-style-type: none"> <li>New South Wales to South Australia, Tasmania, New Zealand.</li> </ul>	Edgar (2008) Gomon et al.(2008)
Diet	<ul style="list-style-type: none"> <li>Invertebrates (mainly salps), krill, crabs, squids.</li> </ul>	Gavrilov and Markina (1979) Annala (1994) Bulman et al. (2001)
Movement and stock structure	<ul style="list-style-type: none"> <li>Schooling fish, highly mobile.</li> <li>Small juveniles pelagic, commonly in association with jellyfish in open coastal waters. Sub-adults often found in sheltered waters of large marine embayments.</li> <li>Although genetics has not confirmed separate stocks, there are indications of population structuring. It is likely that two stocks occur in southern Australian waters (east and west of Bass Strait).</li> </ul>	Gavrilov and Markina (1979) AFMA (2008, 2011)  Bruce et al. (2001a) Robinson et al. (2008)

Natural mortality	<ul style="list-style-type: none"> <li>Estimated between <math>M=0.30</math> and <math>M=0.45</math>.</li> </ul>	Knuckey and Sivakumaran (2001)								
Maximum age	<ul style="list-style-type: none"> <li>Up to 15 years.</li> </ul>	AFMA (2012)								
Growth	<ul style="list-style-type: none"> <li>Maximum length: 90 cm</li> <li>Maximum weight: 4 kg</li> <li>Growth (in New Zealand) described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math> where <math>L</math> is the length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</li> </ul> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_{0\infty}</math></th> </tr> </thead> <tbody> <tr> <td>Combined</td> <td>54.65</td> <td>0.37</td> <td>-0.67</td> </tr> </tbody> </table> <ul style="list-style-type: none"> <li>Length-weight relationship was estimated at <math>W = 0.03L^{2.9}</math> for females and males combined, where <math>W</math> is weight (g) and <math>L</math> is the total length (cm).</li> </ul>	Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Combined	54.65	0.37	-0.67	Gomon et al. (2008) Frimodt (1995) BWAG (1998)
Sex	$L_{\infty}$	$k$	$t_{0\infty}$							
Combined	54.65	0.37	-0.67							
Maturity	<ul style="list-style-type: none"> <li>Size-at-50% maturity estimated at 36 cm (3.67 years) for females.</li> <li>Batch fecundity (BF): <math>\ln(BF) = 2.614\ln(L) + 2.366</math>, where <math>L</math> is length in cm.</li> </ul>	Knuckey and Sivakumaran (2001)								
Spawning	<ul style="list-style-type: none"> <li>Peak spawning in winter, with major regional differences in the magnitude and timing of spawning.</li> <li>Major spawning ground on the central west and northwest coasts in Tasmania.</li> </ul>	Bruce et al. (2001a)								
Early life history	<ul style="list-style-type: none"> <li>Larvae restricted to shelf and slope waters.</li> <li>Larvae likely to be transported by Zeehan Current from spawning grounds of western Tasmania to southeastern Tasmania nursery areas.</li> <li>Larvae settle to the bottom at length &gt; 14.5 mm BL.</li> </ul>	Bruce et al. (2001a) Neira et al. (1998)								
Recruitment	<ul style="list-style-type: none"> <li>Variable. No-stock recruitment relationship established.</li> </ul>									
Gillnet post release survival	<ul style="list-style-type: none"> <li>Low: 35%, but small legal minimum length so rarely discarded.</li> </ul>									

## Background

Blue Warehou occur seasonally in Tasmanian inshore waters, the region representing the southern-most extent of the species' distribution. The availability of the species in coastal waters is assumed to be influenced by prevailing oceanographic conditions and availability of prey species, in particular salps. These factors produce marked inter-annual variability in abundance and hence catch taken from State waters.

Blue Warehou is a Commonwealth managed species and a Memorandum of Understanding (MoU) exists to cover catches from Tasmanian State Fishing Waters. Within the context of this MoU, State catches of Blue Warehou are to be managed within historical levels. The species is

assessed at the Commonwealth level with Tasmanian (and other state) catches included in the stock assessment modelling.

<b>FISHING METHODS</b>	Mainly gillnets, also small mesh nets, seine nets and hook and line (recreational).
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence, rock lobster fishing licence, small mesh gillnet licence, class seine licences).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Possession limit of 20 and bag limit of 10 individuals for recreational fishers.</li> <li>• Minimum size (250 mm TL).</li> </ul>
<b>MAIN MARKET</b>	Local and interstate.

## Current assessment

### Catch, effort and CPUE

In Tasmania, Blue Warehou is taken primarily in gillnets with lesser quantities taken by small mesh nets and seine (Fig. 8.1A). They are captured using a variety of methods by Commonwealth fisheries including other gillnet categories (e.g. shark gillnets), Danish seine and trawl.

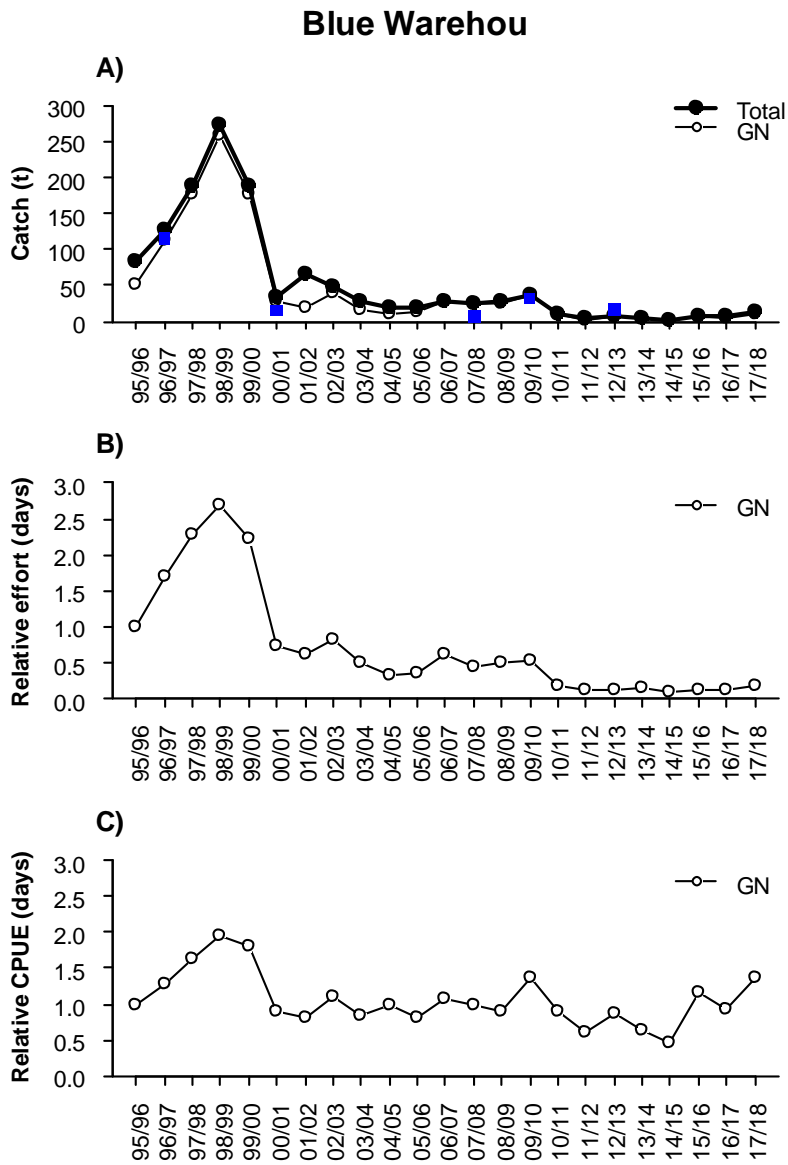
Due to low availability since the early 2000s, the species has been rarely targeted. Catch increased from a historic low of 2.8 t in 2014/15 to 12.6 t in 2017/18. Peak Tasmanian landings were 317.6 t in 1991/92, which corresponded with the peak Australia wide landings of almost 3,000 t (AFMA 2012). Commonwealth commercial catches have also been down in recent years with only 25 tonnes harvested in the 2017/18 fishing season (Patterson et al. 2018). Two stocks of Blue Warehou are believed to occur in southern Australian waters: the east and the west Bass Strait stocks (Bruce et al. 2001b), which has led to the species being managed by AFMA as two stocks. The Tasmanian fishery is now mainly centred off the southeast coast (Fig. 8.2) and thus probably targets the eastern stock. Historically, catches have also been taken off the north and northwest coasts, probably involving the western stock.

Blue Warehou are also targeted by recreational fishers using gillnets, and to a lesser extent line fishing. Historically, recreational catches have been lower than commercial catches (Fig. 8.1A), although in 2010 catch estimates were similar for both sectors (32.5 t for recreational and 37.5 t for commercial) and the 2012/13 the recreational catch of 15.4 t (Lyle et al. 2014a) exceeded the commercial catch of 8.5 t in that year.

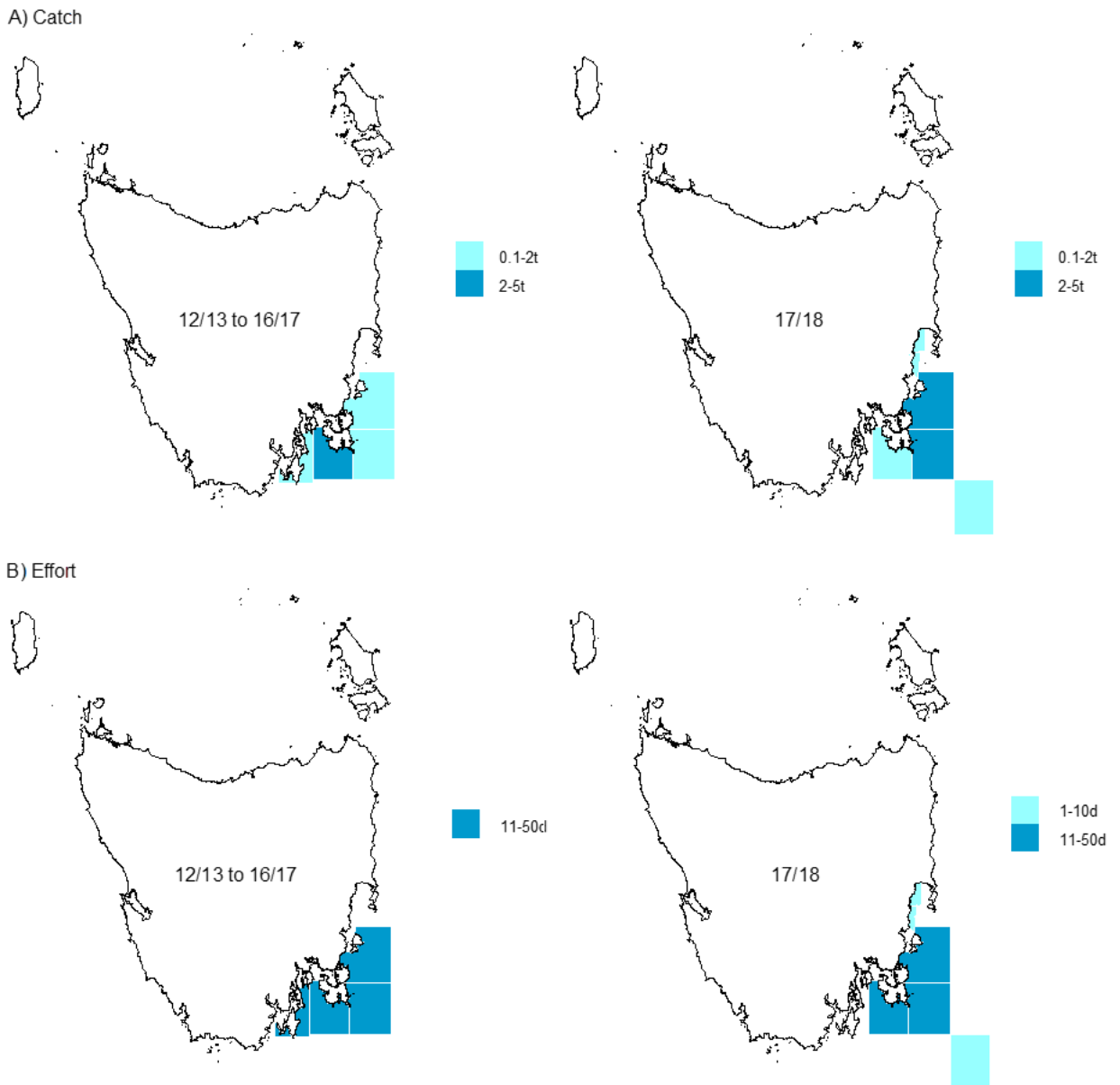
Following the increase in commercial gillnet effort and catch rates between 1995/96 and 1998/99 that resulted in an increase in landings, effort has fallen to substantially lower levels and in recent years has remained low (Fig. 8.1B, C), influenced by the limited availability of Blue Warehou in Tasmanian waters. Catch rates stabilized after 2000/01 at a lower level, with minor peaks in 2009/10 and again 2015/16.

### Ecological Risk Assessment

In the 2012/13 ERA of the Tasmanian scalefish fishery, gillnetting was considered a high risk activity with regard to Blue Warehou due to the depleted stock status (the consequence of historical overfishing) (Bell et al. 2016). There is no new information to suggest that this situation has changed since then. Post-release survival of any by-catch of this species is very low (Lyle et al. 2014b).



**Figure 8.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. GN=gillnet.



**Figure 8.2** (A) Blue Warehouse catches (t) and (B) effort (days) for gillnet fishing by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Commercial catch limit of 318 t as per Memorandum of Understanding (MoU)	No	
	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (187 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (27.6 t)	Yes	↓ 15 t (54.3%)
	• Catch variation from the previous year above the greatest inter-annual increase from the reference period (84.7 t)	No	
	• Catch variation from the previous year above the greatest inter-annual decrease from the reference period (152.8 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (65.3 t)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate (57.3%, in 2010)	Yes	Latest estimate (2012/13): ↑ 63.6%
<b>Biomass</b>	• CPUE < 3 <sup>rd</sup> lowest CPUE value from the reference period (0.0229 t/days fished)	No	
	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0126)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

DEPLETED

The decreasing catch of Blue Warehou over the last 20 years is almost certainly linked to reduced biomass. This is predominantly a result of overfishing by Commonwealth and State fisheries during the 1990s when >2,500 t were caught in several years and >1000 t annually between 1987 and 1998 (AFMA 2012). These figures include all state landings and the Tasmanian landings accounted for around 10% of the total Australian Blue Warehou landings throughout much of this period (AFMA (2012); Appendix 3). In recent years, catches have declined substantially and it is possible that the Tasmanian recreational catch of Blue Warehou now exceeds the catches from Commonwealth and Tasmanian commercial fisheries. While the reduced Commonwealth and Tasmanian catches may be advantageous for the stock(s) in the long-term, it has been noted that it makes it more difficult to assess the “true” state of stock(s) due to a lack of fishery dependent data.

Blue Warehou is under a Commonwealth rebuilding strategy (first introduced in 2008 and later reviewed in 2014), which aims in the first instance to rebuild both east and west stocks to or

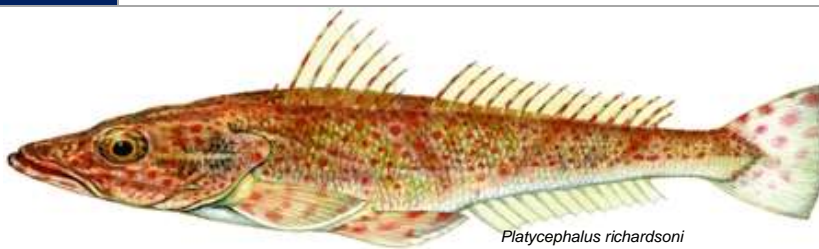
above the default limit reference biomass point ( $B_{LIM}$ ) of 20 per cent of the unfished spawning biomass by 2024 (AFMA 2014). Consequently, the Commonwealth Total Allowable Catch (TAC) for Blue Warehou has been progressively reduced since 2003, and was further reduced to 118 t (split 27 t in the east and 91 t in the west) in 2012/13 (AFMA 2012). AFMA considers the reduction in recent catches (just 1.9 t in 2015/16, 16 t in 2016/17, and 25 t in 2017/18) to be in part a result of their active management and education program. Further management measures include: SESSF fishery closures and gear requirements. There was also a voluntary Commonwealth industry closure implemented between 2008 and 2012 in areas of high Blue Warehou abundance that were believed to be spawning grounds but this was discontinued following review in 2013 due to the patchiness and unpredictability of the species in these areas (AFMA 2014). In Tasmania, management measures include bag and possession limits and a minimum size limit. However, if Blue Warehou stocks start to recover these may not be enough to prevent targeting from State commercial and recreational fishers (SFAC 2015).

Despite the Commonwealth and Tasmanian management measures outlined above, there have been few signs of recovery for the species and the Fishery Status Reports classified the status of Blue Warehou stocks in 2017 as overfished (for biomass) and uncertain (for fishing mortality (Patterson et al. 2018). As such, Blue Warehou is classified as depleted in Tasmanian waters.

# 9. Tiger Flathead

## *Platycephalus richardsoni*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
This is a predominately Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES in 2017 (Patterson et al. 2018). In Tasmania, Tiger Flathead are caught predominately by the commercial sector and while catches have fluctuated between 20–60 tonnes in recent years they have typically represented under 5% of Commonwealth trawl landings.	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery/ Southern and Eastern Scalefish and Shark Fishery (Commonwealth)
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Platycephalus richardsoni*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	Exposed sand and silt habitat. 10–400 m depth.	Edgar (2008) Tilzey et al. (1990)
Distribution	Victoria to New South Wales and around Tasmania.	Edgar (2008) Gomon et al.(2008)
Diet	Fish.	Coleman and Mobley (1984)
Movement and stock structure	Young inhabit shallow waters of the continental shelf and move into the outer shelf zone as they reach maturity.	Kailola et al. (1993) Jordan (1998)
Natural mortality	$M$ between 0.21 and 0.46.	Klaer (2010)
Maximum age	12 years	Rowling (1994) Bani (2005)
Growth	Maximum length: 650 mm (FL) Maximum weight: 2.9 kg Growth (in NSW) described by von Bertalanffy growth function $L = L_{\infty}(1 - e^{-k(t-t_0)})$	Edgar (2008) Gomon et al. (2008) Barnes et al. (2011)

	<p>where <math>L</math> is the length (mm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</p> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_{0\infty}</math></th> </tr> </thead> <tbody> <tr> <td>Females</td> <td>750.2</td> <td>0.13</td> <td>-1.0</td> </tr> <tr> <td>Males</td> <td>418.9</td> <td>0.26</td> <td>-1.0</td> </tr> </tbody> </table> <ul style="list-style-type: none"> <li>Length-weight relationship was estimated at <math>W = 3.25 \times 10^{-6} L^{3.13}</math> for females and males combined where <math>W</math> is weight (g) and <math>L</math> is the total length (mm).</li> </ul>	Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Females	750.2	0.13	-1.0	Males	418.9	0.26	-1.0	
Sex	$L_{\infty}$	$k$	$t_{0\infty}$											
Females	750.2	0.13	-1.0											
Males	418.9	0.26	-1.0											
Maturity	Reach sexual maturity at 4-5 years and total length of 30 cm for males and 36 cm for females.	Fairbridge (1951)												
Spawning	December to February.	Kailola et al. (1993) Jordan (2001a)												
Early life history	Unknown.	Jordan (2001a)												
Recruitment	No-stock recruitment relationship established.													
Gillnet post release survival	Moderate: 50% (all soak durations and including both graball and mullet nets).	Lyle et al. (2014b)												
Rod and-line post release survival	High: >99% for circle hooks and 94–97% for conventional hooks.	Lyle et al. (2007)												

## Background

Several flathead species occur in Tasmanian waters, with Tiger Flathead dominating commercial catches taken by Danish seine. Since Sand Flathead and Tiger Flathead were not routinely distinguished in commercial catch returns until 2007; species-specific catches prior to 2007 have been inferred using the species proportions, by method, for catches taken between 2007/08 and 2011/12. Species-specific fishing effort and catch rates are therefore only considered reliable since 2007/08.

Tiger Flathead constitute a minor component of the recreational harvest of flathead (5 -10%, Lyle et al. 2014a).

<b>FISHING METHODS</b>	Danish seine.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>Gear licence (Scalefish fishing licence, Danish seine licence).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>Possession limit of 30 and bag limit of 20 individuals (Sand and Tiger Flathead) for recreational fishers.</li> <li>Minimum size (320 mm TL).</li> </ul>
<b>MAIN MARKET</b>	Mostly local.

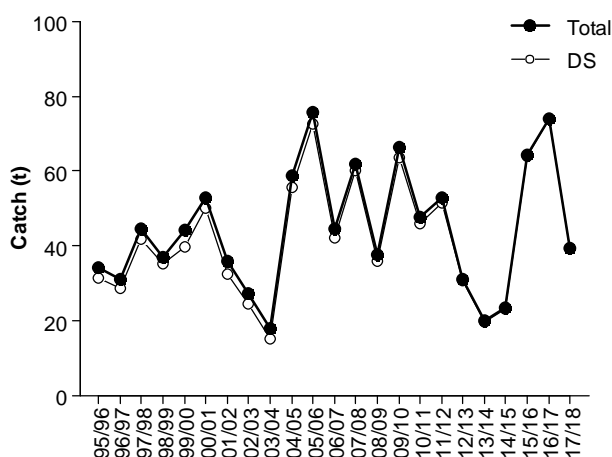
## Current assessment

### Catch, effort and CPUE

The 2017/18 total commercial catch of Tiger Flathead was 39.4 t, down from 74.0 t in 2016/17 and 64.4 t in 2015/16. As the commercial fishery for Flathead has not undergone major changes in its operations since 1995/96 it was feasible to back-calculate catches prior to 2007 (Fig. 9.1). Tiger Flathead landings have been variable over time, fluctuating between 20 and 80 t per annum without an obvious trend (Fig. 9.1 and 9.2).

Danish seine fishing effort and catch rates in 2017/18 declined slightly relative to 2016/17, although have historically fluctuated widely, presumably reflecting the degree of targeting of the species (Fig. 9.2). Peaks in Danish seine catches, effort and CPUE are influenced by a small number of operators that have primarily targeted Tiger Flathead during those years. As with recent years, catches in 2017/18 were derived mainly from the southeast and east coasts (Fig. 9.3).

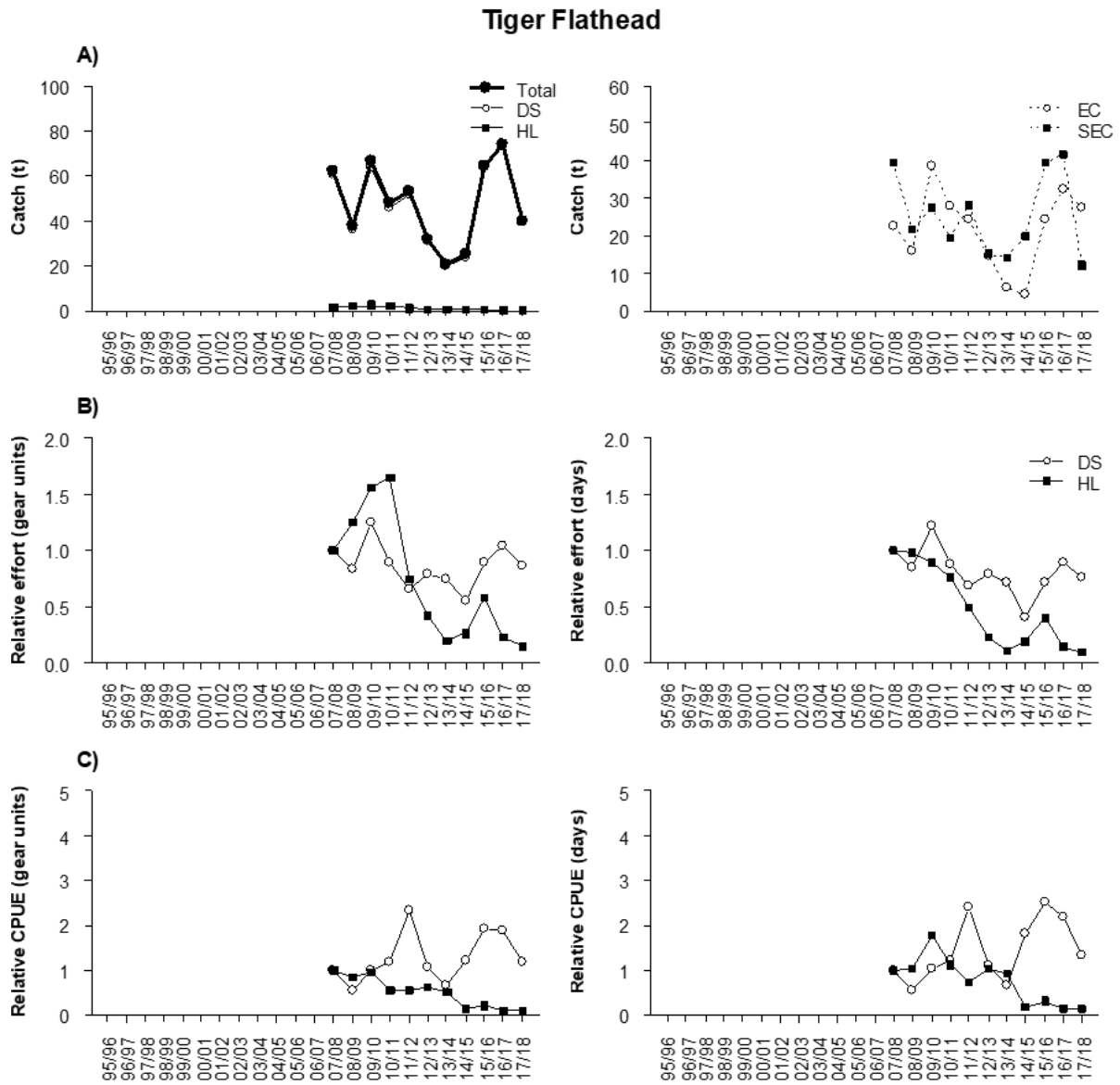
Recreational Flathead catches were estimated at 361 t in 2000/01 (Lyle 2005), 292 t in 2007/08 (Lyle et al. 2009) and 235.9 t in 2012/13 (Lyle et al. 2014a). Tiger Flathead constitute a minor component of the recreational flathead harvest (around 10%, with Sand Flathead constituting around 90% of the recreational harvest).



**Figure 9.1** Back-calculated annual commercial catch (t) by gear for Tiger Flathead. DS=Danish seine.

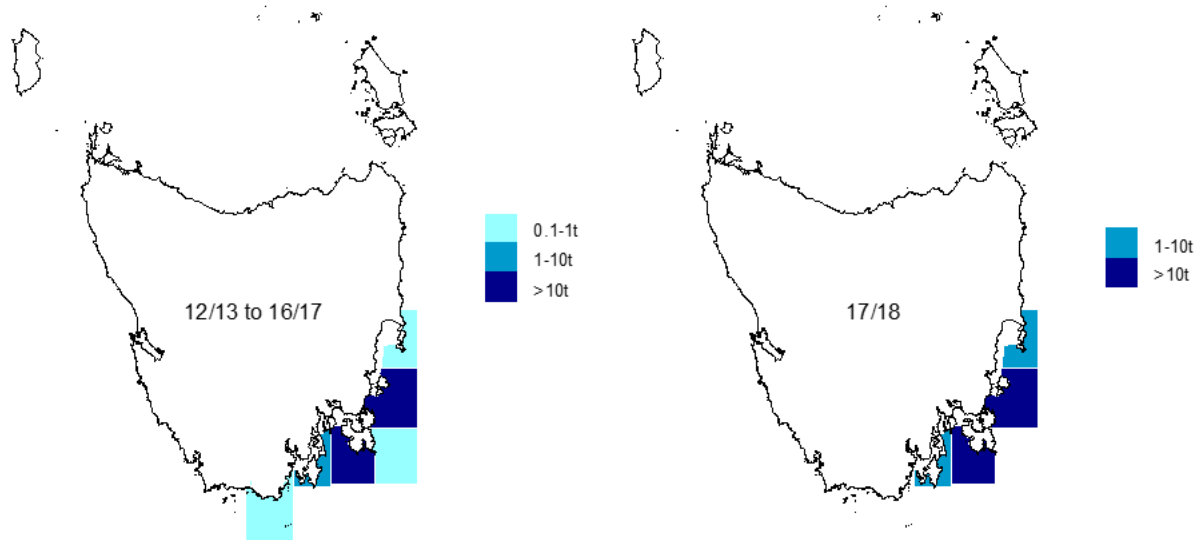
### Ecological Risk Assessment

In the 2012/13 ERA of the Tasmanian scalefish fishery, Danish seining was considered a low risk activity with regard to Tiger Flathead and very low risk to by-product species, such as sharks and mixed fish. Risks to the general ecosystem varied from very low, in terms of discarded fish attracting wildlife, to medium, for possible changes to the seafloor from the net dragging (Bell et al. 2016).

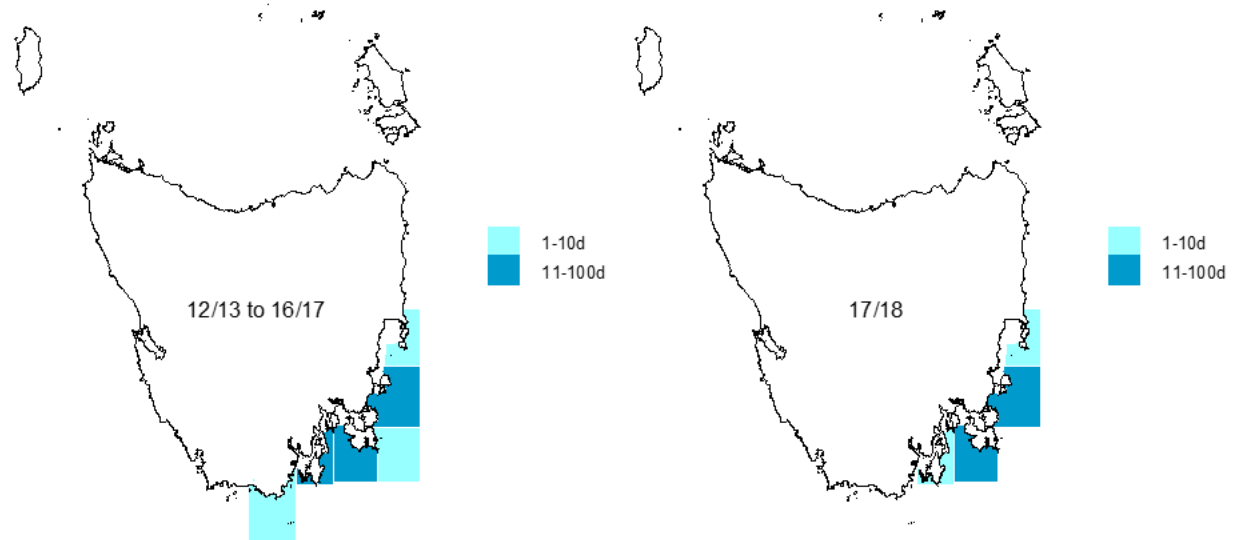


**Figure 9.2** A) Tiger Flathead annual commercial catch (t) by gear (left) and region (right). B) Tiger Flathead commercial effort by method based on gear units (left) and day fished (right) relative to 2007/08. C) Tiger Flathead commercial catch per unit effort (CPUE) based on weight per gear unit (left) and weight per day fished (right) relative to 2007/08. DS=Danish seine, HL=hand-line, SEC=southeast coast, EC=east coast.

A) Catch



B) Effort



**Figure 9.3** (A) Tiger Flathead catches (t) and (B) effort (days) by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

**Reference points for Flathead (combined).**

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (63.1 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (50.5 t)	Yes	↓ 6.6 t (13.1%)
	• Catch variation from the previous year above the greatest inter-annual increase from the reference period (43.5 t)	No	
	• Catch variation from the previous year above the greatest inter-annual decrease from the reference period (-31.9 t)	Yes	↓ 6.1 t (19.1%)
	• Latest recreational catch estimate > recreational catch estimate from the reference period (361 t)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate (80.0% in 2007/08)	Yes	Latest estimate (2012/13) 85.5%
<b>Biomass</b>	• CPUE < 3 <sup>rd</sup> lowest CPUE value from the reference period (0.013 t/days fished)	Yes	0.0013 t/day fished (↓ 10%)
	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0020)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

**Stock status****SUSTAINABLE**

Danish seine catches are highly variable and have historically tended to be inversely related with catches of School Whiting (refer Fig. 2.1), which are also targeted using the method. Tiger Flathead catches were 40.3 t in 2017/18, while School Whiting catches also increased (to 26.0 t tonnes), resulting from an increase in Danish seine fishing effort relative to previous years (up 19 days in 17/18 from 16/17) (Fig. 3.10, Table A2). While commercial catches of Tiger Flathead have been maintained at comparable levels in the past, there are additional and significant trawl catches of flathead (almost exclusively Tiger Flathead) taken from Commonwealth waters as part of the South East Trawl Fishery (Patterson et al. 2018). In the 2017/18 fishing season the total Commonwealth catch of flathead, which is almost entirely Tiger Flathead, was 2,434 t, slightly down from 2,873 t in 2016/17 (Patterson et al. 2018). Accordingly, Tasmanian catches only constitute a fraction (less than 3%) of the overall landings from the stock, which has been classified as sustainably fished in recent years (Flood et al. 2014, Patterson et al. 2015, 2016, 2017, 2018). In accordance with the Commonwealth assessment, Tiger Flathead is assessed as sustainable in Tasmanian waters.

While Tiger Flathead constitute a minor component of the recreational flathead catch, various management changes were introduced in 2015 to improve the sustainability of flatheads in State waters, including an increase in the minimum size limit from 300 mm to 320 mm and introduction of a daily bag limit of 20 per fisher with a possession limit of 30.

# 10. Southern Sand Flathead

## *Platycephalus bassensis*

<b>STOCK STATUS</b>	<b>DEPLETING</b>
Recreational catches of Sand Flathead dominate landings of this species in Tasmania. Fishery independent surveys suggest relatively low abundances of legal sized fish, particularly in south-eastern Tasmania, with populations subject to heavy fishing pressure. While a recent increase in minimum size limit and reduction in bag limit should benefit stocks, current levels of fishing pressure are likely to cause the stock to become recruitment impaired.	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Platycephalus bassensis*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	Sheltered sand and silt habitat. Down to 100m depth.	Edgar (2008) Tilzey et al. (1990)
Distribution	From Western Australia to New South Wales and around Tasmania.	Edgar (2008) Gomon et al. (2008)
Diet	Fish and crustaceans.	Ayling et al. (1975)
Movement and stock structure	Seasonal movements between inshore and offshore in east and southeast Tasmania.	Kailola et al. (1993) Jordan 1998
Natural mortality	$M$ between 0.28 and 0.59 (per year).	Bani (2005)
Maximum age	20 years	Bani (2005)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 51.5 cm</li> <li>Maximum weight: 3.1 kg</li> <li>Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math> where <math>L</math> is the length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</li> </ul>	Edgar (2008) Gomon et al. (2008) Bani (2005) Gomon et al. (2008)

	<p>Growth appears very variable according to location and year.</p> <p>Parameter estimates (for Coles Bay) are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_{0\infty}</math></th> <th>Year</th> </tr> </thead> <tbody> <tr> <td rowspan="2">Females</td> <td>39.01</td> <td>0.25</td> <td>-1.57</td> <td>2001/02</td> </tr> <tr> <td>45.35</td> <td>0.15</td> <td>-2.23</td> <td>2002/03</td> </tr> <tr> <td rowspan="2">Males</td> <td>37.43</td> <td>0.21</td> <td>-1.76</td> <td>2001/02</td> </tr> <tr> <td>34.40</td> <td>0.39</td> <td>-2.51</td> <td>2002/03</td> </tr> </tbody> </table>	Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Year	Females	39.01	0.25	-1.57	2001/02	45.35	0.15	-2.23	2002/03	Males	37.43	0.21	-1.76	2001/02	34.40	0.39	-2.51	2002/03	
Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Year																					
Females	39.01	0.25	-1.57	2001/02																					
	45.35	0.15	-2.23	2002/03																					
Males	37.43	0.21	-1.76	2001/02																					
	34.40	0.39	-2.51	2002/03																					
Maturity	<ul style="list-style-type: none"> <li>• Size-at-50% maturity estimated at 21.7 cm for males, and between 24.7 and 26.3 cm for females depending on location.</li> <li>• Age at 50% maturity variable: 2.5–3.5 years for males and 2.6–5.2 years for females depending on location.</li> </ul>	Bani and Moltschaniwsky (2008)																							
Spawning	<ul style="list-style-type: none"> <li>• From October to March, with a peak from October-December. Spawning occurs throughout their range in southern and eastern Tasmania, including on the shelf.</li> </ul>	Kailola et al. (1993) Jordan (2001a)																							
Early life history	<ul style="list-style-type: none"> <li>• Settlement occurs over an extended period, between 4 to 14 months after spawning.</li> <li>• Size at settlement around 2.1 cm.</li> </ul>	Jordan (2001a)																							
Recruitment	<ul style="list-style-type: none"> <li>• No-stock recruitment relationship established.</li> </ul>																								
Gillnet post release survival	<ul style="list-style-type: none"> <li>• Moderate: 50% (all soak durations and including both graball and mullet nets).</li> </ul>	Lyle et al. (2014b)																							
Rod and line post release survival	<ul style="list-style-type: none"> <li>• High: &gt;99% for circle hooks and 94–97% for conventional hooks.</li> </ul>	Lyle et al. (2007)																							

## Background

Sand Flathead are mainly caught commercially by hand-line, with the species also taken as a by-product in gillnets and Danish seines. Since Sand Flathead and Tiger Flathead were not routinely distinguished in commercial catch returns until 2007 species specific catches prior to 2007 have been inferred using the species proportions, by method, for catches taken between 2007/08 and 2011/12. Species-specific fishing effort and catch rates are therefore only considered reliable since 2007/08.

Sand Flathead are targeted recreationally by hand-line or rod and line and constitute around 90% of the total flathead recreational harvest in Tasmania.

<b>FISHING METHODS</b>	Hand-line, rod and line, and gillnet (Sand Flathead).
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence and Danish seine licence).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Possession limit of 30 and bag limit of 20 individuals (Sand and Tiger Flathead) for recreational fishers.</li> <li>• Minimum size (320 mm TL).</li> </ul>
<b>MAIN MARKET</b>	Mostly local.

## Current assessment

### Biological characteristics

Concerns surrounding the abundance of Sand Flathead led to the establishment of an annual fishery-independent survey, which has been conducted since 2012 (Ewing and Lyle 2015). The survey uses fishing gear and targeting practices typical of recreational fishers in areas of significant effort, including the D'Entrecasteaux Channel, Norfolk Bay and Frederick Henry Bay, and Great Oyster Bay; with sampling occurring during February and March. Fishing was generally conducted over three (not necessarily consecutive) days per region, with from 19 to 21 standard sites fished in each region. Sites sampled in 2018 represent a range of suitable habitats (including depths) for targeting Sand Flathead and provide wide spatial coverage within the given region.

#### *Size composition*

Length frequency histograms from the fishery-independent survey (Fig. 10.1) indicate that the majority of Sand Flathead in the D'Entrecasteaux Channel and Norfolk Bay and Frederick Henry Bay were below the minimum size limits, indicating a low abundance of legal sized fish (Fig. 10.1A and 10.1B). Conversely, in Great Oyster Bay approximately half of the fish were larger than 300 mm in each of the years sampled (Fig. 10.1C).

The proportion of fish exceeding both the original size limit (300mm) and the new size limit (320mm), appears to have stabilised when compared with 2017 catches in the D'Entrecasteaux and Norfolk and Frederick Henry Bays regions (Fig. 10.1A and 10.1B). The proportion of catches from Great Oyster Bay above both the original and the new size limits were lower than those encountered in both 2016 and 2017 (Fig 10.1C).

#### *Age composition*

Age frequency histograms from the fishery-independent survey (Fig. 10.2) indicate that up until 2016 four year olds represented the peak age class of Sand Flathead in the D'Entrecasteaux Channel and Norfolk Bay and Frederick Henry Bay and that the abundance and proportion of females declines rapidly in the older age classes (Fig. 10.2A and 10.2B). In 2018 the modal age classes at D'Entrecasteaux Channel and Norfolk Bay and Frederick Henry Bay was two and five years, respectively. Conversely, in Great Oyster Bay the age structure was strongly influenced by older age classes (> six years) up until 2016 (Fig. 10.2C). In 2016, 2017 and 2018 there was a sharp reduction in the proportional representation of age six years and above (Fig 10.2C). It is unclear whether the poor representation of the older age classes (i.e. 6+) in Great Oyster Bay in recent years is indicative of past fishing pressure or an artefact of sampling. If the former the data

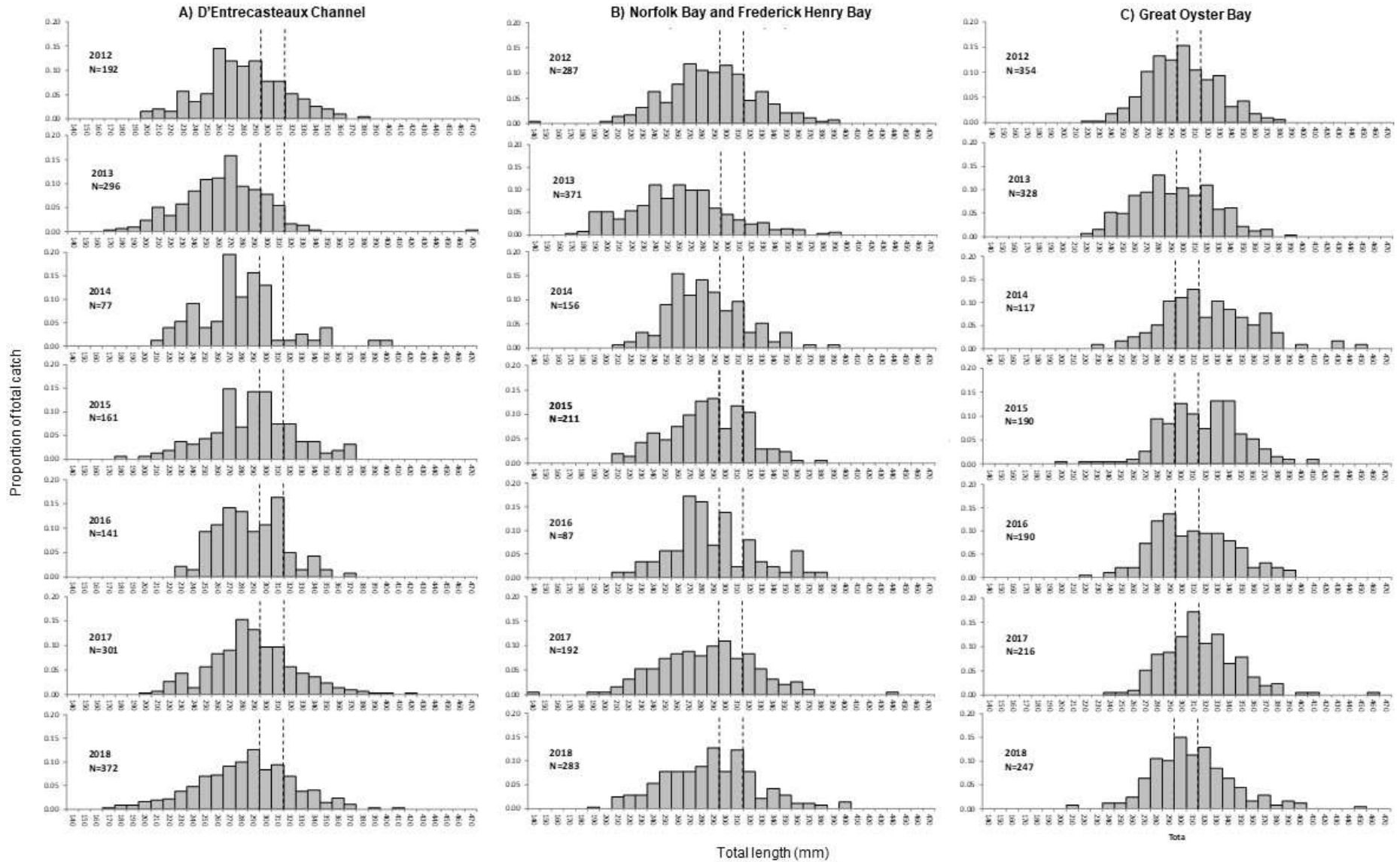
suggest that the cohorts that have been particularly abundant in the past have now been fished to lower levels and the age composition is more consistent with the other two regions. There remains an obvious bias towards males in age classes older than around seven years, which is likely to be the result of higher fishing pressure on females who reach the minimum size limit faster than the slower growing males (Ewing and Lyle 2015).

### Mortality

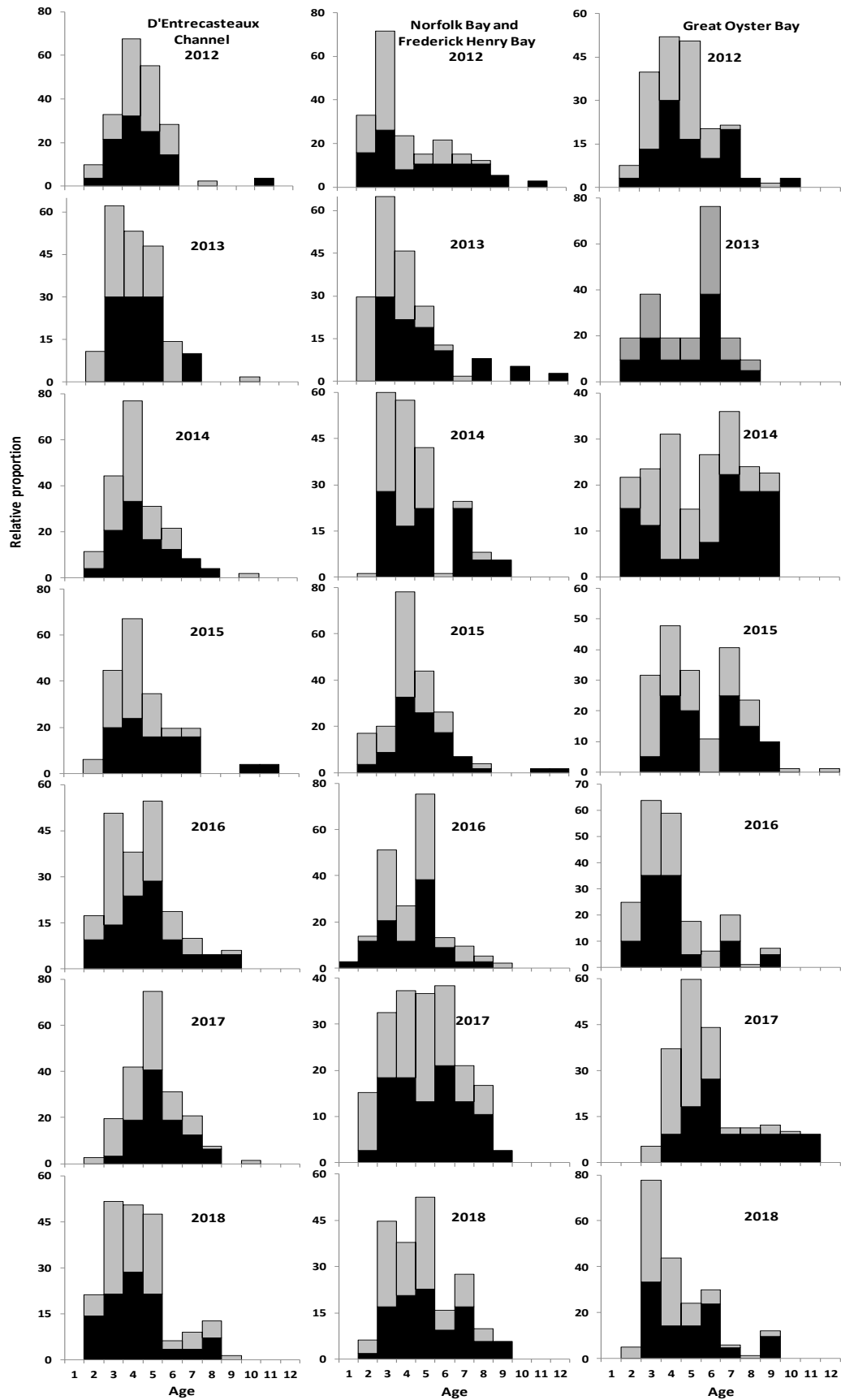
A comparison of estimates of  $F$  from the periods prior to the increase in the MSL (2012–14), and after recovery (2017/18), are presented in Table 8. Female fishing mortality was marginally lower after the increase in the MSL (range of ratios by region of  $F_{2012-14} / F_{2017/18}$  from 0.88 to 0.95). Male fishing mortality experienced more marked reductions after the increase in the MSL (ratios of  $F_{2012-14} / F_{2017/18}$  from 0.13 to 0.59) (Table 8). In 2017/18,  $F$  for females was estimated to be more than three times that of  $M$  in the D'Entrecasteaux Channel, and more than two times that of  $M$  in the Norfolk Bay and Frederick Henry Bay and Great Oyster Bay regions (Table 10.1). For males,  $F$  was estimated to be below  $M$  in 2017/18 at both the D'Entrecasteaux Channel and Great Oyster Bay, but greater than  $M$  within Norfolk Bay and Frederick Henry Bay (Table 10.1).

**Table 10.1.** Sand Flathead fishing mortality estimates ( $F$ ) by region for the years prior to, and after recovery from, the increase in the MSL. DEC is the D'Entrecasteaux Channel region, FHNB is the Frederick Henry-Norfolk Bay region, and GOB is the Great Oyster Bay region.  $Z$  is total mortality derived from catch curves,  $M$  is the mean of two estimates of natural mortality (Hoenig and Lawing 1982 and Ewing and Lyle 2014),  $F$  is fishing mortality [ $Z - (\text{mean of 2 estimates for } M)$ ]. Parameters with a 2012–14 subscript represent mortality prior to the increase in the MSL and the 2017/18 subscript represent mortality after the recovery period. From Ewing and Lyle (2018).

Parameter	DEC		FHNB		GOB	
	Fem	Male	Fem	Male	Fem	Male
$M_{\text{Mean}}$	0.20	0.20	0.20	0.20	0.20	0.20
$Z_{2012-14}$	0.99	0.51	0.74	0.45	0.71	0.32
$Z_{2017/18}$	0.95	0.37	0.69	0.45	0.65	0.22
$F_{2012-14}$	0.79	0.30	0.53	0.25	0.51	0.12
$F_{2017/18}$	0.75	0.17	0.49	0.24	0.45	0.01
$F_{2012-14} / F_{2017/18}$	0.95	0.54	0.91	0.59	0.88	0.13



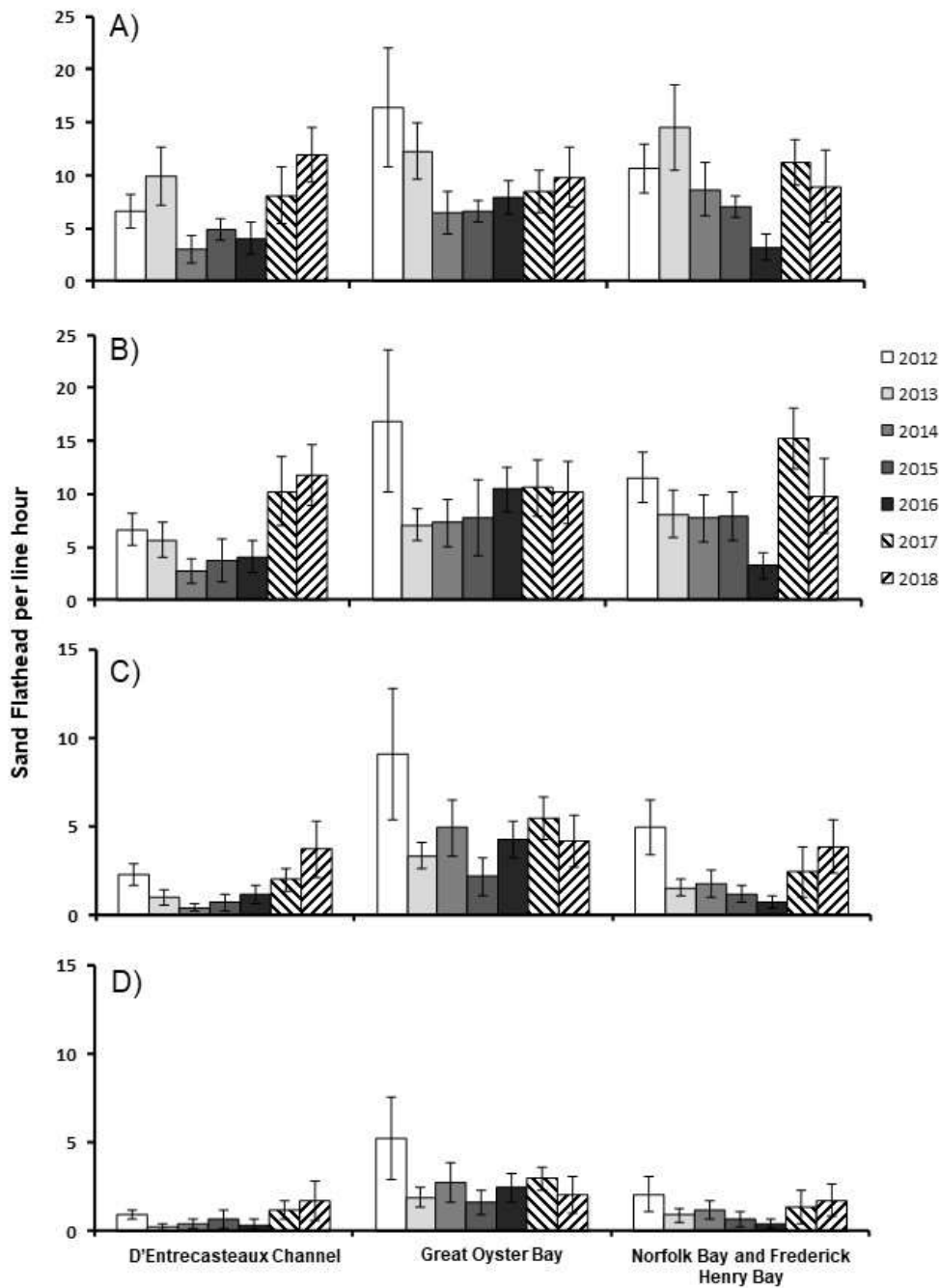
**Figure 10.1** Length frequency histograms for Sand Flathead captured in: A) D'Entrecasteaux Channel; B) Norfolk and Frederick Henry Bay; and C) Great Oyster Bay. Dotted lines indicated minimum legal size limits (300 mm applied to 2015, 320 mm thereafter).



**Figure 10.2** Age frequency histograms for aged Sand Flathead for A) D'Entrecasteaux Channel; B) Norfolk Bay and Frederick Henry Bay; and C) Great Oyster Bay. The black bars indicate males and grey bars indicate females.

### Survey CPUE

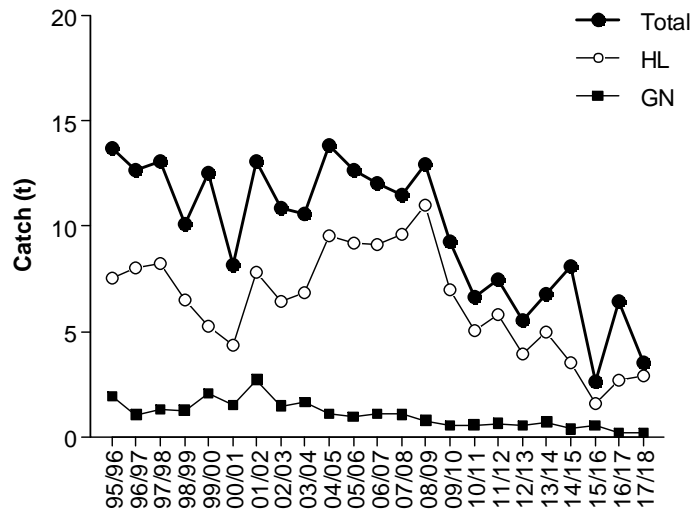
Standardised catch rates from the fishery-independent survey suggest that the populations of legal sized Sand Flathead in the D'Entrecasteaux Channel and Great Oyster Bay were stable but at depleted levels in 2013–2016 compared with 2012 (Fig. 10.3C & D). By contrast, the standardised catch rate in Norfolk-Frederick Henry Bay declined to its lowest level in 2016. In 2018, standardised catch rates of fish greater than 300 mm TL and 320 mm TL increased within the D'Entrecasteaux Channel and Norfolk Bay and Frederick Henry Bay regions relative to previous years (Fig. 10.3C & D), suggesting that the management measures introduced in late 2015 may be having a positive effect on the stock. Within Great Oyster Bay, standardised catch rates in 2018 were generally similar to those observed in 2016 and 2017 (Fig. 10.3)



**Figure 10.3** Mean catch rates (fish per line hour) by region and year for Sand Flathead: (A) raw catch rates; (B) standardised catch rates; (C) standardised catch rates for fish above the original 300mm MLS, and (D) standardised catch rates for fish above the new 320mm MLS. Error bars represent 95% confidence intervals.

## Commercial catch, effort and CPUE

The commercial fishery for Flathead has not undergone major changes in its operations since 1995/96. It was therefore possible to back calculate catches for Sand Flathead prior to 2007, when species were not distinguished, based on the average proportion of species by gear type from 2007/08 to 2011/12 (Fig. 10.4). Sand Flathead catches remained relatively stable between 10 and 15 t per annum until 2008/09 but have generally declined since then to reach a historical low of 2.7 t in 2015/16 (Fig. 10.4 and 10.5). In 2017/18, catches were 3.5 t, down from 2016/17 when 6.4 t was harvested (Fig. 10.5). In 2017/18, 2.9 t of the Sand Flathead catch was taken by hand-line, which is consistent with 2016/17, when 2.7 t was landed. Catches in 2017/18 were derived mainly from the southeast and east coasts (Fig. 10.6), consistent with previous years.



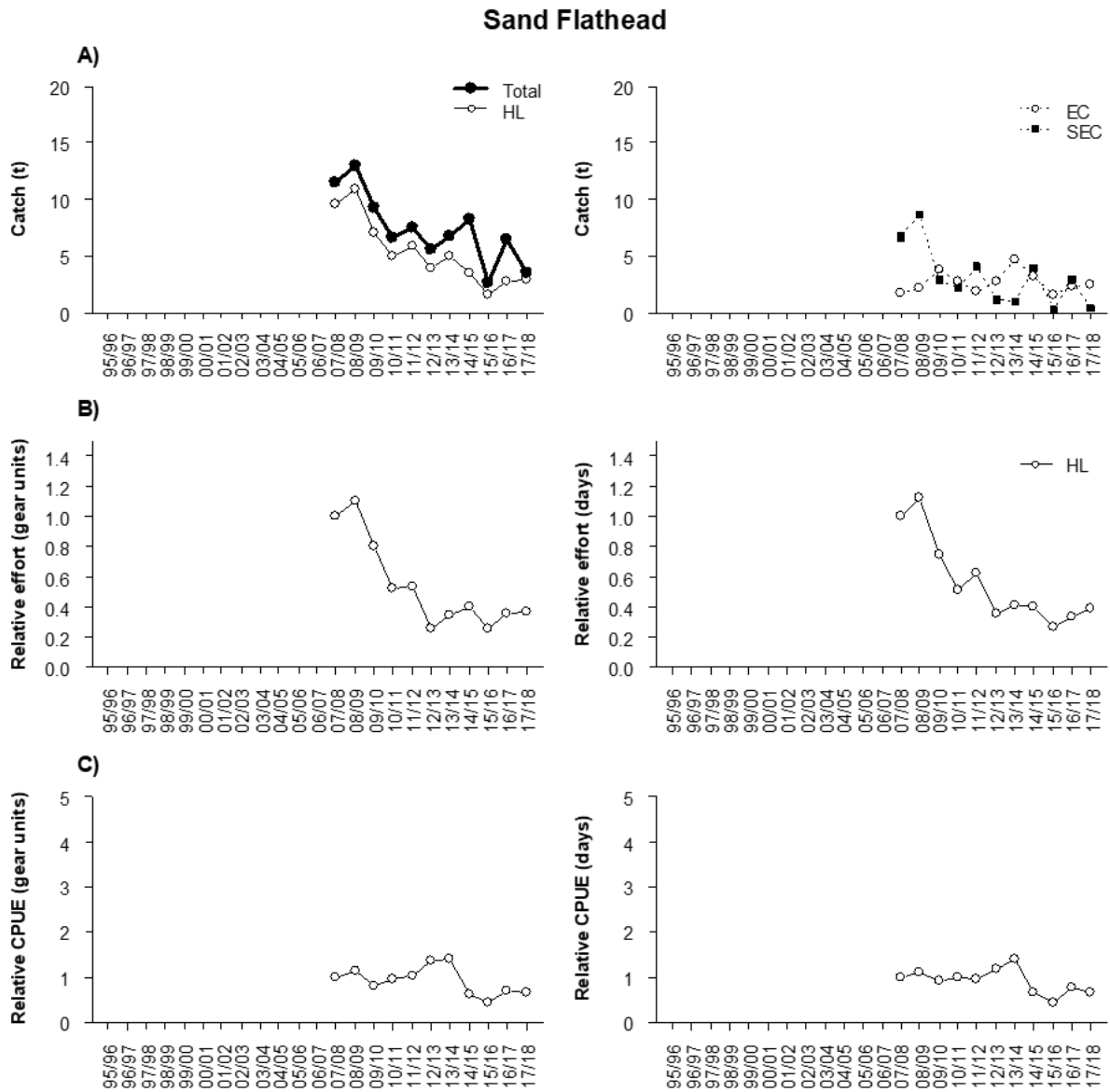
**Figure 10.4** Back-calculated annual commercial catch (t) by gear for Sand Flathead. HL=hand line and GN=gillnet (catches taken by other methods are not shown).

Hand-line fishing effort increased slightly in 2017/17 to 2016/17 and 2015/16, (Fig. 10.5). Catch rates in 2017/18 were similar to those observed in 2016/17, and higher than those in 2014/15 and 2015/16, potentially as a result of the management measures introduced in late 2015, but remain low relative to historical levels (Fig. 10.5).

Recreational Flathead (all species) catches were estimated at 361 t in 2000/01 (Lyle 2005), 292 t in 2007/08 (Lyle et al. 2009) and 235.9 t in 2012/13 (Lyle et al. 2014a). Sand Flathead are by far the dominant flathead species taken by recreational fishers (>90% in numbers).

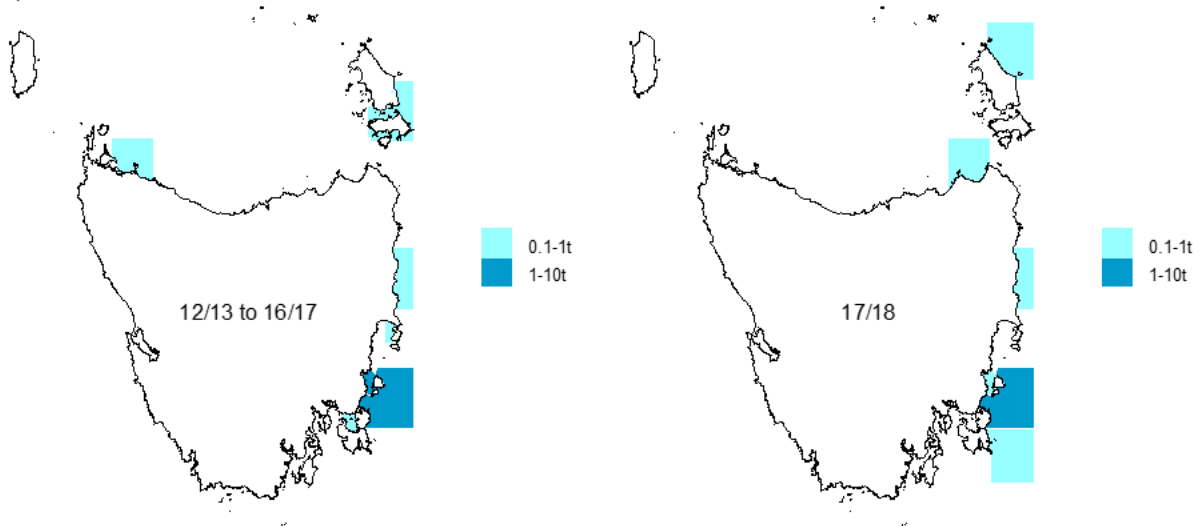
## Ecological Risk Assessment

In the 2012/13 ERA of the Tasmanian scalefish fishery, hand-lining was considered a medium risk with regard to Sand Flathead due to evidence of the population being subject to heavy fishing pressure. Hand-lining was considered a medium risk to by-product mixed fish species due to the uncertainty surrounding their population status. Impacts on communities and protected species were generally low or negligible although heavy fishing pressure on Sand Flathead populations in inshore and estuarine waters was considered to represent a medium risk to the trophic structure of these systems (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.

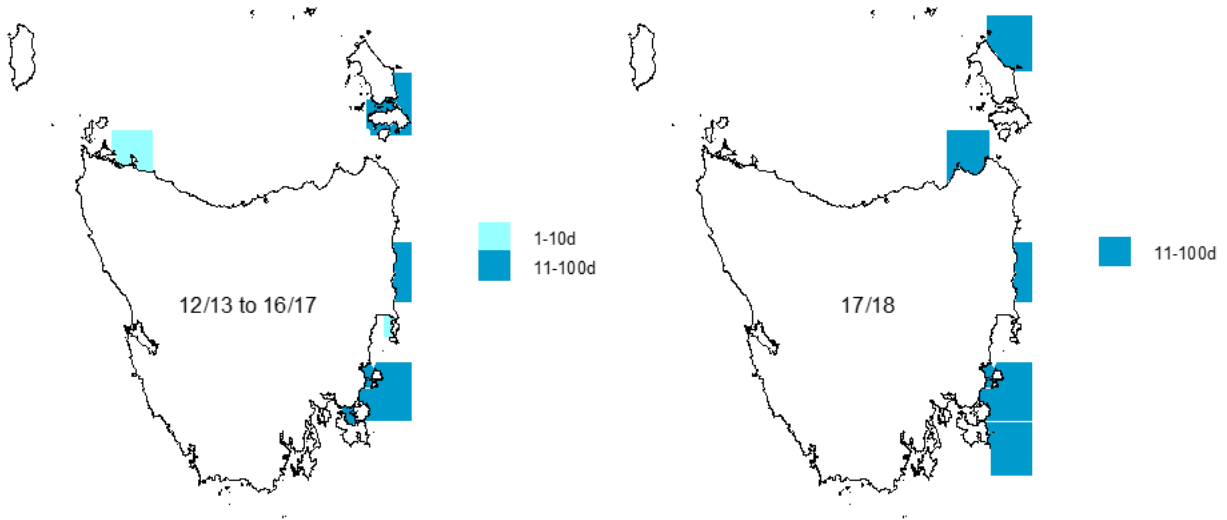


**Figure 10.5** A) Sand Flathead annual commercial catch (t) by gear (left) and region (right). B) Sand Flathead commercial effort by method based on gear units (left) and day fished (right) relative to 2007/08. C) Sand Flathead commercial catch per unit effort (CPUE) based on weight per gear unit (left) and weight per day fished (right) relative to 2007/08. HL=hand-line, SEC=southeast coast, EC=east coast.

A) Catch



B) Effort



**Figure 10.6** (A) Sand Flathead catches (t) and (B) effort (days) by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

### Reference points for Flathead (combined)

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (63.1 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (50.5 t)	Yes	↓ 6.6 t (13.1%)
	• Catch variation from the previous year above the greatest inter-annual increase from the reference period (43.5 t)	No	
	• Catch variation from the previous year above the greatest inter-annual decrease from the reference period (-31.9 t)	Yes	↓ 6.1 t (19.1%)
	• Latest recreational catch estimate > recreational catch estimate from the reference period (361 t)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate (80.0% in 2007/08)	Yes	Latest estimate (2012/13) 85.5%
<b>Biomass</b>	• CPUE < 3 <sup>rd</sup> lowest CPUE value from the reference period (0.013 t/days fished)	Yes	0.0013 t/day fished (↓ 10%)
	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0020)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

### Stock status

**DEPLETING**

The main impact on Sand Flathead stocks is from the recreational sector with catches >20 times that of the commercial sector. Due to an absence of targeting among commercial fishers, a Sand Flathead fishery-independent survey commenced in 2012 to support classification of this species.

The survey over recent years has identified a low relative abundance of legal-size fish, particularly in the D'Entrecasteaux Channel and Frederick Henry-Norfolk Bay with a higher proportion of older age class males, which is likely a factor of females reaching the minimum size limit faster than slower growing males (Ewing and Lyle 2015). In the initial years of the survey, catch rates suggested that the abundances of legal sized fish in each of the major fishing areas were depleted relative to historic levels.

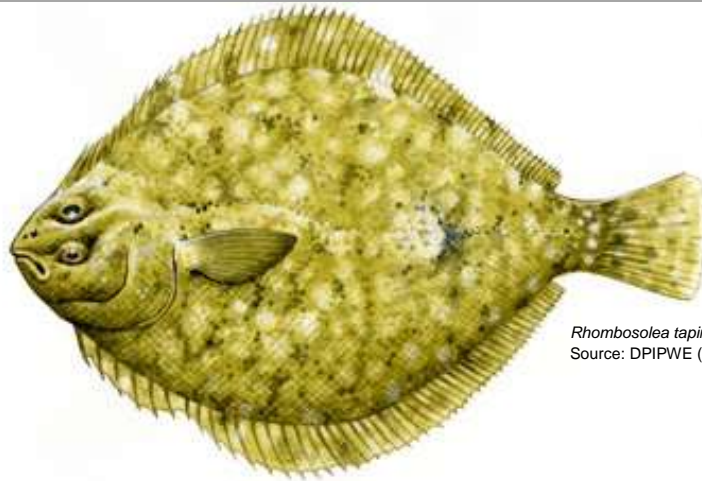
In late 2015, various management changes were introduced to improve the sustainability of this species including: an increase in the minimum size limit from 300 mm to 320 mm and introduction of a daily bag limit of 20 per fisher with a possession limit of 30. Increases in fisheries-independent catch rates in 2017 and 2018 suggest these changes may be having a positive effect on the stock, particularly at heavily fished areas, however continued monitoring is required

to examine this hypothesis. For females, fishing mortality was estimated to greatly exceed natural mortality at all regions sampled. As the current fishing pressure on the stock remains high and likely to cause the stock to become recruitment impaired, Sand Flathead is classified as depleting.

# 11. Flounder

## *Pleuronectidae* family

<b>STOCK STATUS</b>	UNDEFINED
Greenback Flounder ( <i>Rhombosolea tapirina</i> ) constitute the majority of the commercial catch, which remains low due to the ban on overnight gillnetting and limited market demand. Catch and catch rates are considered unreliable estimators of abundance and the status of the stock remains uncertain.	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Rhombosolea tapirina*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Sheltered sand, silt and mud substrates in estuaries and coastal waters. Between 0 and 100 m depth.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>Southern Western Australia to southern New South Wales, and around Tasmania. Also in New Zealand.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Polychaetes and benthic crustaceans.</li> </ul>	Edgar (2008) Ferguson (2006)
Movement and stock structure	<ul style="list-style-type: none"> <li>Several genetically distinct stocks in Australia: west coast of Tasmania, east coast of Tasmania, Victoria.</li> </ul>	van den Enden (2000)
Natural mortality	<ul style="list-style-type: none"> <li>Estimated at <math>M = 0.85</math> (for populations in New Zealand).</li> </ul>	Sutton et al. (2010)
Maximum age	<ul style="list-style-type: none"> <li>10 years</li> </ul>	Sutton et al. (2010)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 45 cm</li> <li>Maximum weight: 0.6 kg</li> </ul>	Edgar (2008) Kailola et al. (1993)

	<ul style="list-style-type: none"> <li>Growth (for populations in New Zealand) described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math> where <math>L</math> is the fork length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</li> </ul> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_{0\infty}</math></th> </tr> </thead> <tbody> <tr> <td>Females</td> <td>55.82</td> <td>0.26</td> <td>-1.06</td> </tr> <tr> <td>Males</td> <td>52.21</td> <td>0.24</td> <td>-1.32</td> </tr> </tbody> </table> <ul style="list-style-type: none"> <li>Length-weight relationship was estimated at <math>W = 0.036 L^{2.7}</math> for females and <math>W = 0.039 L^{2.64}</math> for males where <math>W</math> is weight (g) and <math>L</math> is the tail length (cm).</li> </ul>	Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Females	55.82	0.26	-1.06	Males	52.21	0.24	-1.32	Sutton et al. (2010)
Sex	$L_{\infty}$	$k$	$t_{0\infty}$											
Females	55.82	0.26	-1.06											
Males	52.21	0.24	-1.32											
Maturity	<ul style="list-style-type: none"> <li>Sexual maturity at about 218.6 mm TL for females and 190 mm TL for males.</li> </ul>	Crawford (1984)												
Spawning	<ul style="list-style-type: none"> <li>From June to October.</li> <li>Females are serial spawners and move from the shallows in deeper areas of tidal rivers and estuaries, and offshore for spawning.</li> <li>The relationship between batch fecundity and fork length is linear between 24.7 and 34.3 cm with <math>F = -1053.65 + 85.85L</math>, where <math>F</math> is the fecundity (in number of eggs) and <math>L</math> is the tail length (cm).</li> <li>Pelagic eggs, 0.7–1.0 mm in diameter.</li> </ul>	Crawford (1984)												
Early life history	<ul style="list-style-type: none"> <li>Incubation of 82–93 hours.</li> <li>Larvae hatch at 1.9 mm between May and November. Larvae remain in the plankton for over 30 days until they reach 6 mm, and then undergo metamorphosis which finishes 65 days post-hatching.</li> <li>Settlement inshore occurs during late winter to early summer.</li> <li>Juveniles live on sand flats in water less than a meter deep.</li> </ul>	Edgar (2008) Crawford (1984, 1986) Jenkins (1986)												
Recruitment	<ul style="list-style-type: none"> <li>No-stock recruitment relationship established.</li> </ul>													
Gillnet post release survival	<ul style="list-style-type: none"> <li>High: 96.1%</li> </ul>	Lyle et al. (2014b)												

## Background

While the various species of flounder are not always differentiated in logbooks, Greenback Flounder (*Rhombosolea tapirina*) constitute the vast majority of the commercial catch. Long-nouted Flounder (*Ammotretis rostratus*) are also taken but in small quantities. The main fishing methods used to target flounder are spear and gillnet, however, a 2010 ban on the overnight setting of gillnets in most waters has resulted in a marked reduction in gillnet fishing for flounder.

<b>FISHING METHODS</b>	Spear, gillnet, some beach seine.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence).</li> <li>• Recreational gear licence (graball and/or mullet net licence).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Possession limit of 30 and bag limit of 15 individuals for recreational fishers.</li> <li>• Minimum size: 25 cm.</li> </ul>
<b>MAIN MARKET</b>	Local (Tasmania).

## Current assessment

### Catch, effort and CPUE

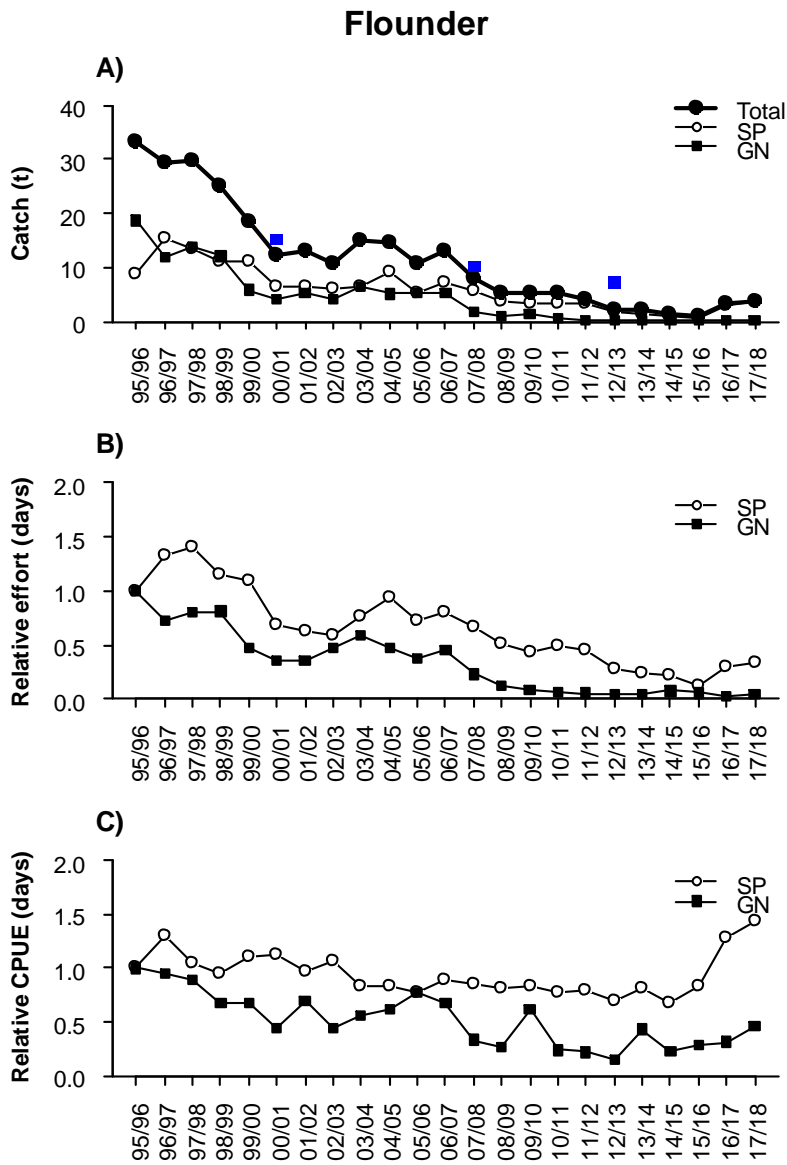
State-wide Flounder landings have declined steadily since the mid-1990s, reaching a historical low of 1.0 t in 2015/16 (Fig. 11.1A). Catches increased slightly in 2017/18 to 3.9 t, the highest level in the last six years. Since the ban on night netting, Flounder have been caught almost exclusively using spear, with 3.6 t taken using this method in 2017/18. Commercial catches have contracted spatially over recent years to Norfolk Bay and the Tamar estuary (Fig. 11.2). In 2017/18, fishing was concentrated around these areas as well as within Macquarie Harbour (Fig. 11.2).

Consistent with the trend in catches, effort for both methods has been declining steadily since the mid-1990s (Fig. 11.1B). Gillnet and spear fishing effort increased slightly relative to previous years. Catch rates for both gillnet and spear continued the increasing trend observed since 2014/15 (Fig. 11.1B & C).

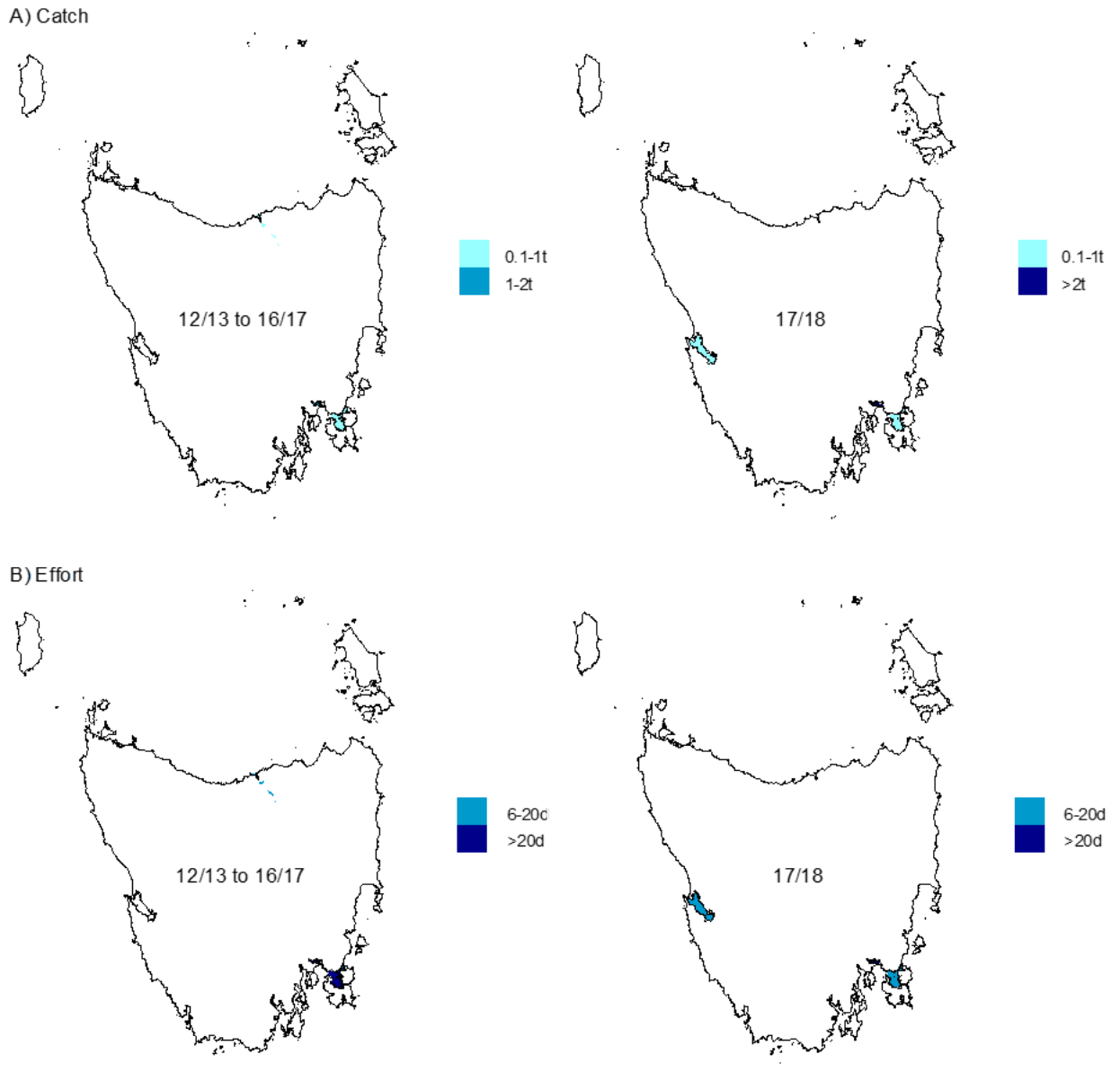
Flounder is a relatively important recreational species, and in recent years catches for the recreational sector have exceeded those of the commercial sector (Fig. 11.1A). Recreational catches have declined progressively over recent years: recreational catches were estimated at 15.2 t in 2000/01 (Lyle 2005), 10.1 t in 2007/08 (Lyle et al. 2009) and 7.2 t in 2012/13 (Lyle et al. 2014a).

### Ecological Risk Assessment

In the 2012/13 ERA of the Tasmanian scalefish fishery, spearing was considered a very low risk to flounder populations due to the negligible fishing effort directed at this species in recent years. All other ecosystem components were considered negligible risk because spearing has rarely been used in recent years and is highly selective, with fish sighted and captured individually (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.



**Figure 11.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. SP=spear, GN=gillnet.



**Figure 11.2** (A) Flounder catches (t) and (B) effort (days) by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	<ul style="list-style-type: none"> <li>Catch &gt; 3<sup>rd</sup> highest catch value from the reference period (29.4 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Catch &lt; 3<sup>rd</sup> lowest catch value from the reference period (12.3 t)</li> </ul>	Yes	↓ 8.4 t (68.1%)
	<ul style="list-style-type: none"> <li>Latest recreational catch estimate &gt; recreational catch estimate from the reference period (15.2 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Proportion of recreational catch to total catch &gt; previous proportion estimate (56.3% in 2007/08)</li> </ul>	Yes	Latest estimate (2012/13): ↑ 77.4%
<b>Biomass</b>	<ul style="list-style-type: none"> <li>Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0017)</li> </ul>	No	
<b>Stock stress</b>	<ul style="list-style-type: none"> <li>Significant change in the size/age composition of commercial catches</li> </ul>	Not assessed	
	<ul style="list-style-type: none"> <li>Significant numbers of unhealthy fish landed</li> </ul>	No	

## Stock status

UNDEFINED

The declining Flounder catch is likely to be related to market demand but has also been impacted by the ban on overnight gillnetting. The Tasmanian catch is sold locally and demand for Flounder has decreased over the last two decades to the extent that catch and catch rates are considered unreliable estimators of trends in abundance. There is insufficient information to confidently classify this stock.

# 12. Gould's Squid

## *Nototodarus gouldi*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
<p>This is a predominately Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES in 2017 (Patterson et al. 2018). Dual-licensed vessels fish in Tasmanian waters especially in years of peak abundance. The species is characterised by high inter-annual variability in abundance in State waters with low catches in recent years.</p>	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery/Southern Squid-jig Fishery (Commonwealth)
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Nototodarus gouldi*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Open water. Inhabits coastal, inner shelf and shelf break waters down to 600 m depth.</li> </ul>	Stark (2008)
Distribution	<ul style="list-style-type: none"> <li>From southern Queensland to mid-Western Australia, and around Tasmania.</li> </ul>	Dunning (1998) Dunning and Förch (1998)
Diet	<ul style="list-style-type: none"> <li>Small planktonic crustaceans, fish and squids.</li> </ul>	Machida (1983) O'Sullivan and Cullen (1983) Smith (1983) Uozumi (1998)
Movement and stock structure	<ul style="list-style-type: none"> <li>Move widely within a 300 km<sup>2</sup> area but no large-scale migration between feeding and spawning areas as for other similar squid species.</li> <li>Stock structure uncertain but appears to be only one population in southern Australia.</li> </ul>	Jackson et al. (2005a) Triantafillos et al. (2004)

Natural mortality	<ul style="list-style-type: none"> <li>No estimates available.</li> </ul>	
Maximum age	<ul style="list-style-type: none"> <li>Around 1 year.</li> </ul>	Jackson et al. (2005b)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 40 cm</li> <li>Maximum weight: 1.6 kg</li> <li>Size-at-age highly variable between individuals, years and locations.</li> <li>Growth rate rapid: between 2.559 and 5.596 g.d<sup>-1</sup> for females, and between 1.622 and 5.307 g.d<sup>-1</sup> for males.</li> </ul>	Norman and Reid (2000) Jackson et al. (2005b) Jackson et al. (2003)
Maturity	<ul style="list-style-type: none"> <li>Size-at-50% maturity: between 30.6 to 31.4 cm mantle length (ML) for females, and 20.5 to 21.5 cm ML for males.</li> </ul>	Stark (2008)
Spawning	<ul style="list-style-type: none"> <li>Spawns once and then die.</li> <li>Spawning all year-round.</li> <li>Egg mass are free-floating gelatinous sphere of at least 1.5 m in diameter and contains several thousands of eggs.</li> </ul>	Jackson et al. (2005b) Uozumi (1998) O'Shea et al. (2004)
Early life history	<ul style="list-style-type: none"> <li>Hatching throughout the year.</li> </ul>	Jackson et al. (2005b) Uozumi (1998)
Recruitment	<ul style="list-style-type: none"> <li>Highly variable. No-stock recruitment relationship established.</li> </ul>	

## **Background**

Gould's Squid, like most cephalopod species, has a very brief life cycle and can vary significantly in abundance between years. Environmental conditions are acknowledged as influences on larval and juvenile survival (Flood et al. 2012). It is likely that there is only one biological stock throughout southern Australian waters.

The Commonwealth Southern Squid-jig Fishery operates in Bass Strait waters using automatic squid-jigs and Gould's Squid are a regular by-product in the South East Trawl Fishery. Occasionally, Gould's Squid become available in high numbers in Tasmanian State waters, particularly around southeast Tasmania. Consequently, dual licensed vessels tend to fish in State waters during summer before moving back to traditional fishing grounds in Bass Strait.

Gould's Squid are processed into 'tubes' and frozen. Given the unpredictable occurrence of the species in Tasmanian waters there is limited local processing capacity, which has limited the development of the fishery. There is also limited market demand with consumers preferring Southern Calamari. Due to the abovementioned factors, catch does not necessarily reflect biomass.

<b>FISHING METHODS</b>	Automated squid jig.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence, Automated squid-jig licence).</li> <li>• Temporal and spatial closures (October-November) of some east coast waters.</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Possession limit of 30 and bag limit of 15 individuals for recreational fishers.</li> </ul>
<b>MAIN MARKET</b>	Interstate.

## Current assessment

### Catch, effort and CPUE

Gould's Squid availability in Tasmanian waters is highly variable as reflected in the catch history (Fig. 12.1A). Since 1995/96, there have been a few peaks of abundance, notably in 1999/2000, 2011/12, 2012/13 and again in 2015/16. The Gould's Squid catch for 2012/13 was the highest since 1995/96 (~1000 t) with the Australia-wide catch predominantly coming from Tasmanian waters (Flood et al. 2014). In 2017/18 a total of 528.0 t of Gould's Squid were taken from Tasmanian waters, all but 1.0 t of which was caught by automatic jig. Historically, in between the peaks, catches of the species have fluctuated between 0.1 and 50 t. The majority of the catch in 2017/18 was taken from the southeast coast of Tasmania and around King Island (Fig.12.2).

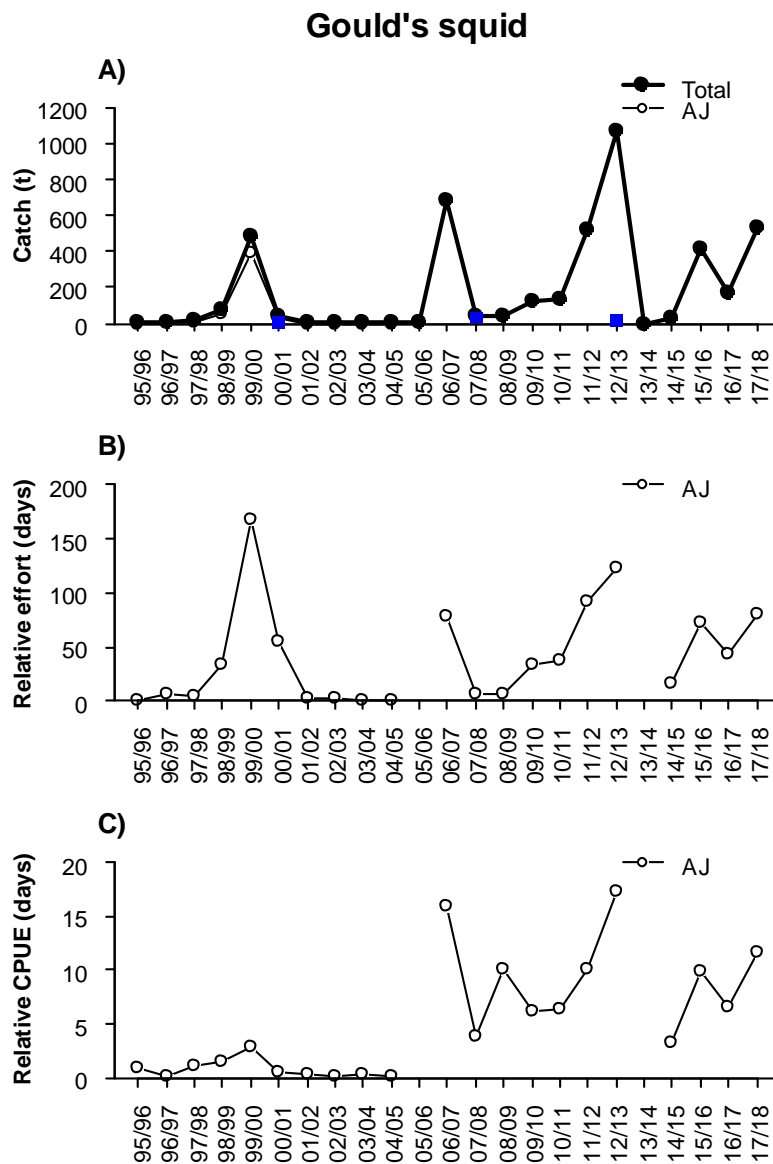
Gould's Squid catches from the recreational sector (Fig. 12.1A) were estimated at 5 t in 2000/01 (Lyle 2005), 36.6 t in 2007/08 (Lyle et al. 2009) and 21.4 t in 2012/13 (Lyle et al. 2014a).

Effort tends to follow the catch pattern closely and matches the availability of the species. In recent years high catches have been achieved with relatively low effort with a record catch obtained in 2012/13 (Fig. 12.1B).

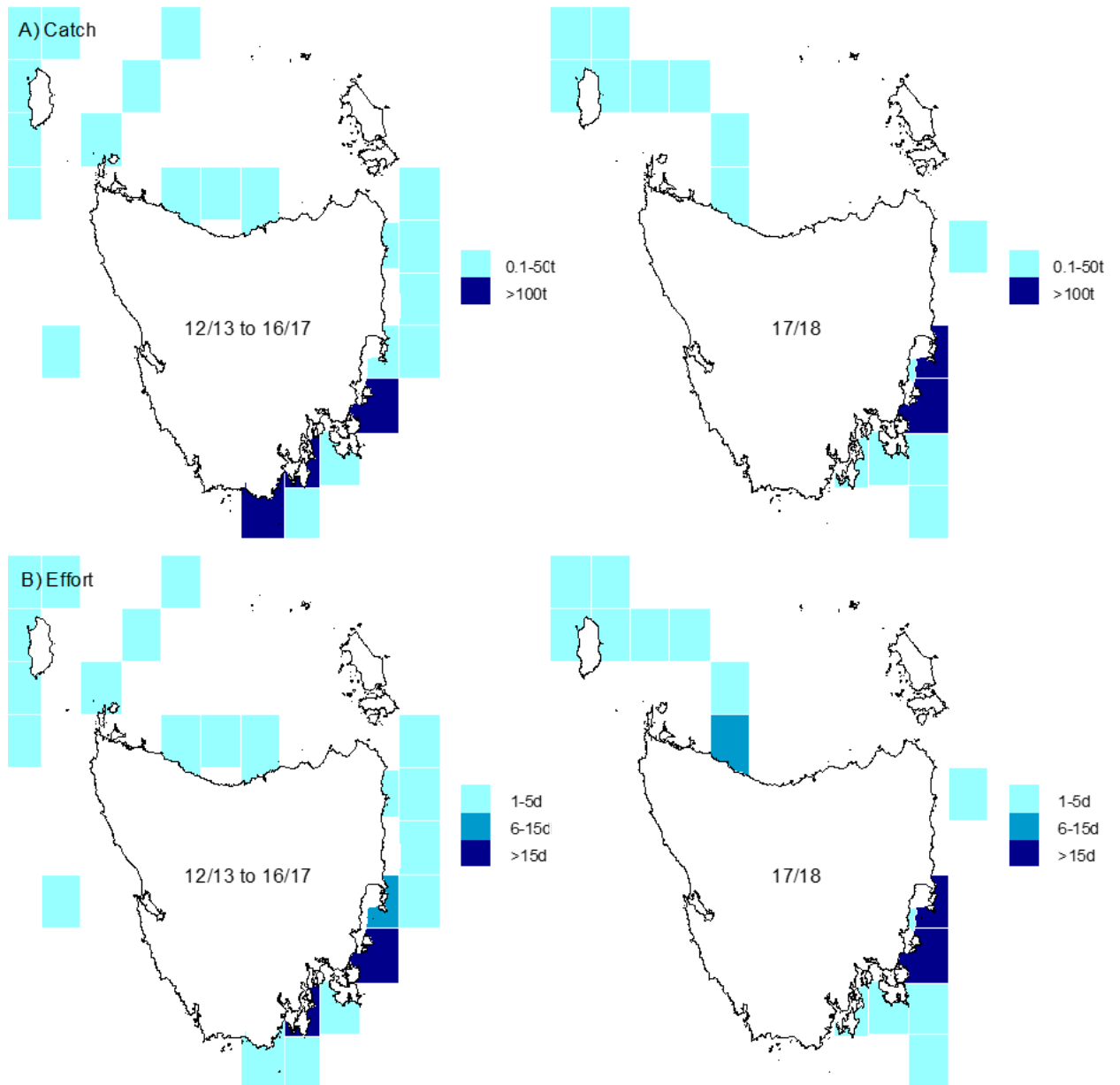
Overall, catch rates remained relatively low until 2008/09. Prior to 2013/14 there was an increase in catch rate, both in gear units and by days fished (Fig. 12.1C).

### Ecological Risk Assessment

In the 2012/13 ERA of the Tasmanian scalefish fishery automatic squid jig fishing methods were considered a very low risk activity with regard to Gould's Squid, non-retained species and the general environment (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.



**Figure 12.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on days fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished (right) relative to 1995/96. AJ=automatic squid jig. Data includes Australian Fisheries Management Authority (AFMA) catch in State waters. Note: no catch or effort using Automatic squid jig was recorded for 2005/06, 2006/07 or 2013/14.



**Figure 12.2** (A) Gould's Squid catches (t) and (B) effort (days) by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right). No notable catch of Gould's Squid was recorded in 2013/14. Data includes Australian Fisheries Management Authority (AFMA) catch in State waters.

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (79.7 t)	Yes	↑ 448.4 t (562.8%)
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (2.1 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (5 t)	Yes	16.4 t ↑ (328%)
	• Proportion of recreational catch to total catch > previous proportion estimate (44.4% in 2007/08)	No	
<b>Biomass</b>	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0657)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

**SUSTAINABLE**

The reference point for highest catch was breached, however, relative to years of high squid abundance in Tasmanian waters, 528.0 tonnes of Gould's Squid landed in 2017/18 represented a moderately high catch. Spatial variability in the distribution and abundance of Gould's Squid is a characteristic of this species and thus the strong inter-annual variability in the Tasmanian fishery needs to be considered in the context of the overall fishery and population status.

A recreational reference point was also breached, however it is unlikely that the recreational sector poses a threat to the stock as estimated catches remain small in comparison to the commercial sector when Gould's Squid are available in Tasmanian waters and even more so when the Southern Squid-jig Fishery catches are taken into consideration.

Gould's Squid are short lived, spawn year round and display highly variable growth and size/age at maturity, meaning they can rapidly increase in numbers during favourable environmental conditions. As a result, Gould's Squid may be less susceptible to overfishing than longer-lived species (Flood et al. 2012). However, their short life span (1 year) implies a reliance on a single cohort, which leaves the species very susceptible to environmental impacts on subsequent recruitment. The fishing effort in the Commonwealth Southern Squid-jig Fishery has decreased markedly since the late 1990s due to economic factors and catches are considered lower than the historical catches from the late 1970s. A biological stock depletion study found that no overfishing had occurred (Sahlqvist and Skirtun 2011). Tasmanian catches in 2017/18 represent around half of the total Commonwealth catch reported for 2017 (828 t). The species was classified as not overfished or subject to overfishing in the 2015, 2016, or 2017 Fishery Status Reports (Patterson et al. 2016, 2017, 2018), and as Sustainable in recent Status of Australian Fish Stock Reports (Flood et al. 2012, Flood et al. 2014). This assessment has therefore been applied to the Tasmanian component of the fishery.

# 13. Jack Mackerel

## *Trachurus declivis*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
<p>This is a predominately Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES for 2017 (Patterson et al. 2018). Only minor catches of this species have been taken from Tasmanian waters in recent years due to one operator leaving the fishery, and thus patterns of catch and effort do not reflect the stock status. The current level of fishing pressure in Tasmania is unlikely to cause the stock to become recruitment impaired.</p>	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery/Small Pelagic Fishery (Commonwealth)
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Trachurus declivis*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Open water. Between 0–500 m depth.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>Western Australia (Shark Bay) to southern Queensland and around Tasmania. Also in New Zealand.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Krill, planktonic crustaceans, fish.</li> </ul>	Kailola et al. (1993)
Movement and stock structure	<ul style="list-style-type: none"> <li>Schooling fish. Normally live in continental shelf waters. May move close to seabed during winter.</li> <li>Most likely two subpopulations: one eastern Australian (east Tasmania and along the eastern seaboard of Australia) and one western Australian (west Tasmania, Great Australian Bight and Western Australia).</li> </ul>	Kailola et al. (1993) Bulman et al. (2008)
Natural mortality	<ul style="list-style-type: none"> <li>Estimated between <math>M = 0.63</math> and <math>0.70</math>.</li> </ul>	Stevens and Hansfeld (1982)
Maximum age	<ul style="list-style-type: none"> <li>25 years</li> </ul>	Paul (2000)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 64 cm</li> <li>No difference between male and female growth</li> <li>Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math></li> </ul>	Paul (2000) Lyle et al. (2000)

	<p>where <math>L</math> is the length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</p> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_{0\infty}</math></th> </tr> </thead> <tbody> <tr> <td>Combined</td> <td>36.2</td> <td>0.267</td> <td>-1.21</td> </tr> </tbody> </table> <ul style="list-style-type: none"> <li>Length-weight relationship was estimated at <math>W = 1.46 \cdot 10^{-8} L^{2.982}</math> for both males and females, where <math>W</math> is weight (g) and <math>L</math> is the length (cm).</li> </ul>	Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Combined	36.2	0.267	-1.21	
Sex	$L_{\infty}$	$k$	$t_{0\infty}$							
Combined	36.2	0.267	-1.21							
Maturity	<ul style="list-style-type: none"> <li>Sexual maturity between 3 and 4 years of age, at sizes around 27 cm and weights around 250 g.</li> </ul>	Webb (1976)								
Spawning	<ul style="list-style-type: none"> <li>Occurs over a wide area in Tasmania.</li> <li>Between late December and early March.</li> <li>Pelagic eggs.</li> </ul>	Stevens and Hansfeld (1982)								
Early life history	<ul style="list-style-type: none"> <li>Larvae carried by inshore currents.</li> <li>Juveniles inhabit coastal and estuarine waters although they may sometimes be found offshore.</li> </ul>	Kailola et al. (1993) Williams and Pullen (1986)								
Recruitment	<ul style="list-style-type: none"> <li>No-stock recruitment relationship established.</li> </ul>									
Gillnet post release survival	<ul style="list-style-type: none"> <li>NA</li> </ul>									

## Background

The Jack Mackerel fishery in Tasmania started in the early 1970s, with a one year venture catching 6,300 t in 1973. In 1985 another venture aimed at fishmeal production using purse seine nets commenced and Jack Mackerel landings rose rapidly to over 40 000 t in 1986/87 (Kailola et al. 1993). By 2000, fishers were struggling to catch surface schools and the industry began mid-water trawling for the species (and redbait) in Commonwealth waters. Small quantities of Jack Mackerel are also taken inshore as by-product of beach seining and inshore purse seining.

<b>FISHING METHODS</b>	Mainly purse seine, also beach seine.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>Gear licence (Scalefish fishing licence, class seine licence).</li> <li>Species licence (Mackerel A or B).</li> <li>Recreational gear licence (beach seine).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>Possession limit of 60 and bag limit of 30 individuals for recreational fishers.</li> <li>Commercial catches taken by Mackerel licence holders (A &amp; B) are decremented against the TAC allocated to the Commonwealth Small Pelagic Fishery.</li> </ul>
<b>MAIN MARKET</b>	Local (Tasmania).

## **Current assessment**

### **Catch, effort and CPUE**

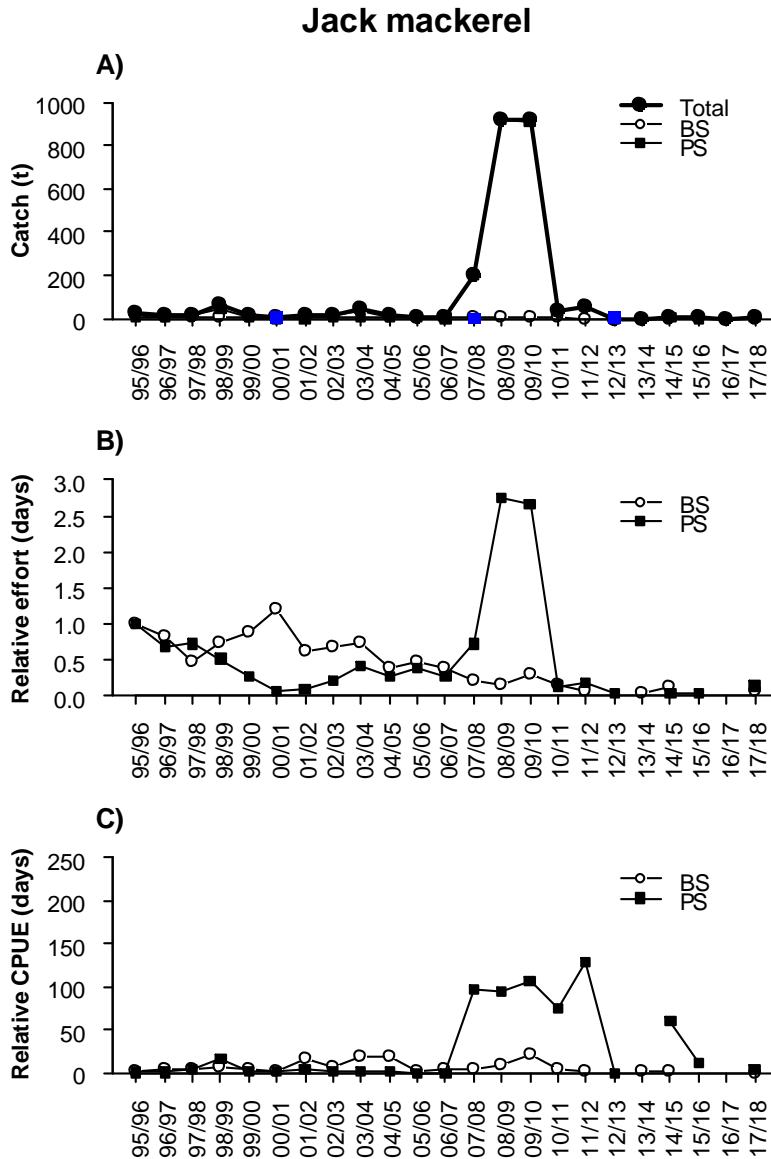
Catches of Jack Mackerel in Tasmanian waters and reported in the General Fishing Returns have been variable since 1995/96, oscillating between 2.6 and 59.8 t up until 2007/08 when there was a sharp increase in purse seine effort targeting Jack Mackerel (Fig. 13.1A). The Jack Mackerel catches peaked at 919.6 t in 2008/09, but declined sharply in 2010/11 and 2011/12 to around 60 t, the result of the major purse seiner ceasing operations. In 2017/18 2.0 t of Jack mackerel were caught in Tasmania, slightly up from the historical low of 66 kg recorded in 2016/17. Catches usually focus on the southeast coast (Fig. 13.2). It should be noted that between 1995 and 1999, purse seine catches taken as part of the then Zone A Jack Mackerel fishery, and recorded in a separate logbook, ranged from 447 t (1995/96) to 8,458 t (1997/98), averaging 4,485 t per year for that period. These data are not presented in Fig. 13.1A.

Jack Mackerel is not a significant recreational species, catches estimated at 3.2 t in 2000/01 (Lyle 2005), 1.0 t in 2007/08 (Lyle et al. 2009) and 5.2 t in 2012/13 (Lyle et al. 2014a) (Fig. 13.1A).

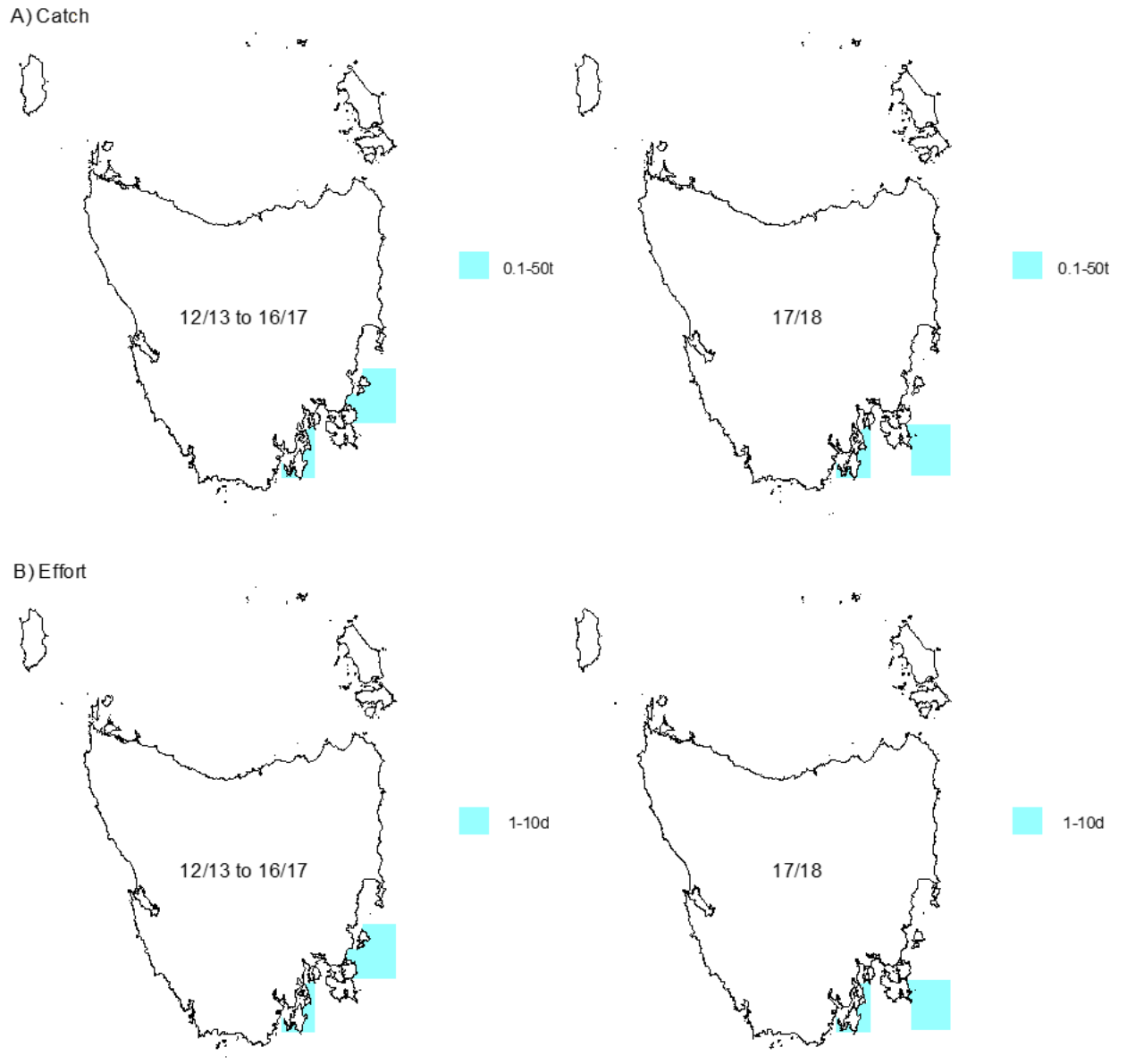
The use of purse seining by a major operator between 2008/09 to 2009/10 resulted in a spike in effort for that period of time. Beach seine effort has been declining slowly over time though Jack Mackerel represent a by-product and no meaningful catch rate trends can be drawn from these data (Fig. 13.1B). Purse seine catch rates were low until the species began being targeted in 2008/09 and remained high until 2011/12, which was the last year that the species was targeted (Fig. 13.1C). Since that time, landings have been low and there has been no targeted fishing in Tasmanian waters. By contrast, landings increased sharply in the Commonwealth Small Pelagic Fishery with the operation of a large factory trawler.

### **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian scalefish fishery, offshore purse seining was considered a negligible risk activity to populations of Jack Mackerel due to the small amount of catch currently taken in the fishery. However, it is noted that if catches increased then the risks would need to be reassessed (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.



**Figure 13.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. BS=beach seine, PS=purse seine. Note: no purse seine catch for Jack Mackerel was reported in Tasmanian waters during 2013/14, and no beach seine or purse seine catch for Jack Mackerel was recorded for 2016/17.



**Figure 13.2** (A) Jack Mackerel catches (t) and (B) effort (days) for beach seine and purse seine by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right). Note no beach seine or purse seine catch for Jack Mackerel was recorded in Tasmanian waters in 2016/17.

## Limit reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	<ul style="list-style-type: none"> <li>Catch &gt; 3<sup>rd</sup> highest catch value from the reference period (26.2 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Catch &lt; 3<sup>rd</sup> lowest catch value from the reference period (9.1 t)</li> </ul>	Yes	↓ 7.1 t (77.6%)
	<ul style="list-style-type: none"> <li>Latest recreational catch estimate &gt; recreational catch estimate from the reference period (3.2 t)</li> </ul>	Yes	↑ 2.0 t (62.5%)
	<ul style="list-style-type: none"> <li>Proportion of recreational catch to total catch &gt; previous proportion estimate (0.4% in 2007/08)</li> </ul>	Yes	Latest estimate (2012/13: ↑ 96.3%)
<b>Biomass</b>	<ul style="list-style-type: none"> <li>Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0254)</li> </ul>	No	
<b>Stock stress</b>	<ul style="list-style-type: none"> <li>Significant change in the size/age composition of commercial catches</li> </ul>	Not assessed	
	<ul style="list-style-type: none"> <li>Significant numbers of unhealthy fish landed</li> </ul>	No	

## Stock status

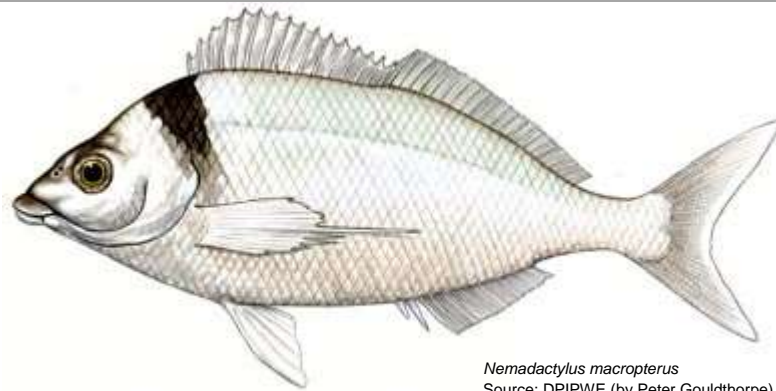
**SUSTAINABLE**

The reference point for lowest catch was breached due to minimal fishing occurring in 2017/18. The minimal commercial catch in recent years has also meant that the recreational catch is proportionally higher than historically and therefore breached the recreational reference points. Recent trends in the commercial fishery have been the response of a single operator entering and leaving the fishery and do not reflect the stock status. A 2014 study assessed the spawning stock biomass for eastern Australia to be in the order of 150,000 tonnes (Ward et al. 2015). Jack Mackerel are assessed by the Commonwealth Small Pelagic Fishery Scientific Panel and, based on current catch levels and spawning biomass, the eastern Jack Mackerel stock is assessed as not overfished and not subject to overfishing (Patterson et al. 2017, 2018). This assessment has been applied to the Tasmanian component of the fishery.

# 14. Jackass Morwong

## *Nemadactylus macropterus*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
This is a Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES for 2017 (Patterson et al. 2018), with very low commercial catches in Tasmania.	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery/Southern and Eastern Scalefish and Shark Fishery (Commonwealth)
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Nemadactylus macropterus*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Exposed sand and silt, reefs. Between 5 and 400 m depth.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>From central Queensland to southern Western Australia, and around Tasmania. Also in New Zealand, southern Africa and South America.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Polychaete worms, crustaceans, molluscs and echinoderms.</li> </ul>	Godfriaux (1974)
Movement and stock structure	<ul style="list-style-type: none"> <li>No genetic variation in southern Australia indicating larval mixing.</li> <li>Existence of at least three sub-populations: Tasmania, New South Wales/Victoria and Great Australian Bight.</li> </ul>	Richardson (1982) Grewe et al. (1994) Elliott and Ward (1994) Thresher et al. (1994)
Natural mortality	<ul style="list-style-type: none"> <li><math>M = 0.10</math> (New Zealand population).</li> </ul>	Parker and Fu (2011)
Maximum age	<ul style="list-style-type: none"> <li>50 years</li> </ul>	Edgar (2008)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 70 cm TL</li> <li>Maximum weight: 2.9 kg</li> <li>Growth varies according to location. Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math></li> </ul>	Kailola et al. (1993) Jordan (2001b)

	<p>where <math>L</math> is the length (cm FL), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</p> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_0</math></th> </tr> </thead> <tbody> <tr> <td>Females</td> <td>38.4</td> <td>0.36</td> <td>-0.07</td> </tr> <tr> <td>Males</td> <td>36.2</td> <td>0.42</td> <td>0.15</td> </tr> </tbody> </table>	Sex	$L_{\infty}$	$k$	$t_0$	Females	38.4	0.36	-0.07	Males	36.2	0.42	0.15	)
Sex	$L_{\infty}$	$k$	$t_0$											
Females	38.4	0.36	-0.07											
Males	36.2	0.42	0.15											
Maturity	<ul style="list-style-type: none"> <li>Sexual maturity at about 25 cm TL and ca. 3 years of age.</li> </ul>	Edgar (2008)												
Spawning	<ul style="list-style-type: none"> <li>Between February and June.</li> <li>At least two spawning areas: northern one (probably southern New South Wales and eastern Victoria) and a southern one (probably western and southern Tasmania).</li> </ul>	Lyle and Ford (1993) Bruce et al. (2001b)												
Early life history	<ul style="list-style-type: none"> <li>Planktonic larval stage of 7–10 months.</li> <li>Larvae up to 30 mm drift with current on the surface up to 250 km east of Tasmania.</li> <li>Settlement at 7–9 cm long.</li> <li>Juveniles live near shallow reefs.</li> </ul>	Francis (2001) Bruce et al. (2001b) Kailola et al. (1993)												
Recruitment	<ul style="list-style-type: none"> <li>No-stock recruitment relationship established.</li> </ul>													
Gillnet post release survival	<ul style="list-style-type: none"> <li>Moderate: 52%</li> </ul>	Lyle et al (2014b)												

## Background

Jackass Morwong is a Commonwealth-managed species. While there is a good market for Jackass Morwong, the species is not available in large numbers in Tasmanian waters and is mainly a by-product of gillnetting rather than a target species. Tasmanian commercial catches reached a maximum of around 250 t in the late 1980s due to the inshore demersal trawl fishery, but this fishery ceased operations following the introduction of a state-waters ban on otter board trawling in 2001. Most of the Jackass Morwong catch originates from trawling outside Tasmanian waters and while the stocks were assessed to be overfished from 2008 to 2010, stocks have been assessed as sustainable in Commonwealth Fishery Status Reports since 2011 (that is this species as not overfished nor subject to overfishing) (Patterson et al. 2016, 2017, 2018).

<b>FISHING METHODS</b>	Mainly gillnet, also hand-line and drop-line.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>Gear licence (Scalefish fishing licence).</li> <li>Recreational licence (graball and/or mullet net).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>Possession limit of 20 and bag limit of 10 individuals for recreational fishers.</li> <li>Minimum size: 25 cm TL.</li> </ul>
<b>MAIN MARKET</b>	Local.

## **Current assessment**

### **Catch, effort and CPUE**

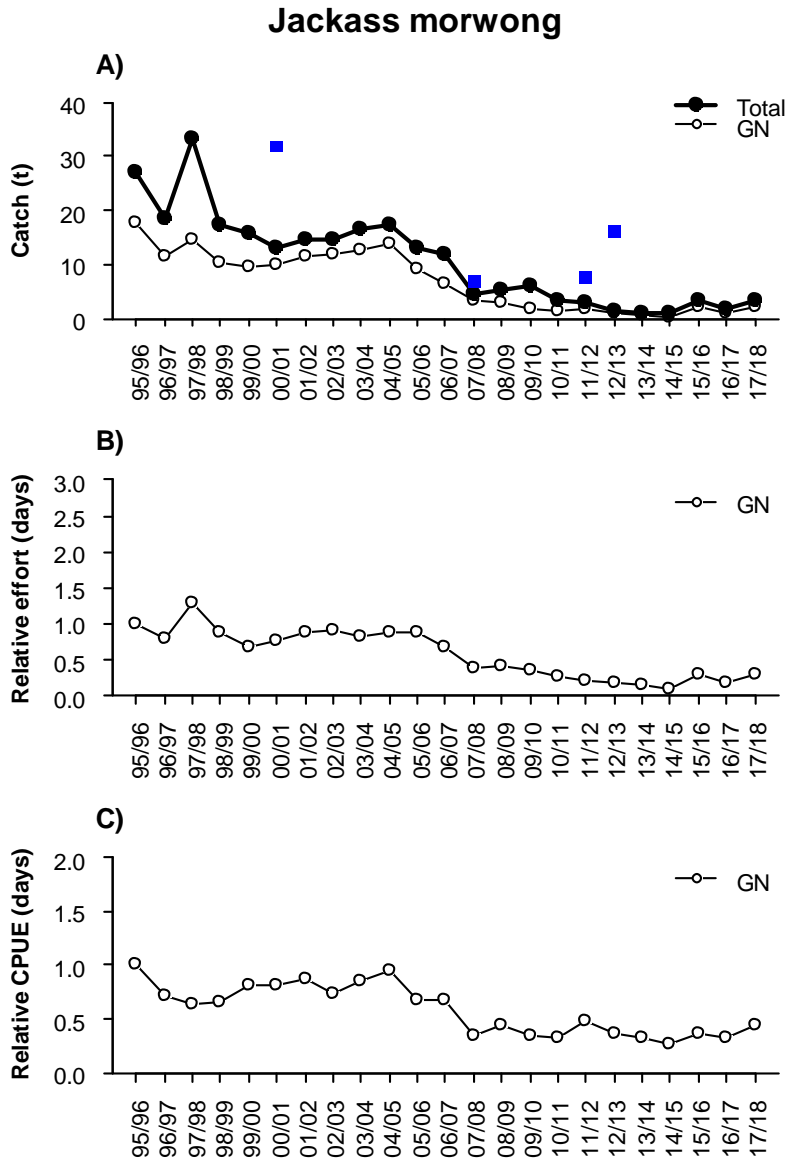
Total commercial catch of Jackass Morwong was 3.3 t in 2017/18, representing the largest catch since 2009/10 (Fig 14.1A). Commercially, Jackass Morwong is mainly caught by gillnet and landings have declined steadily since 1995/96. The majority of the catch is taken from the east and southeast coast (Fig. 14.2).

Jackass Morwong is an important recreational species, with catches estimated at higher levels than those of the commercial fishery (Fig. 14.1A). Estimates were 31.9 t in 2000/01 (Lyle 2005), 6.8 t in 2007/08 (Lyle et al. 2009), 7.7 t in 2011/12 (Tracey et al. 2013) and 16.1 t in 2012/13 (Lyle et al. 2014a). In addition to gillnet catches, Jackass Morwong are commonly caught recreationally by line fishing, often associated with targeted fishing for Striped Trumpeter.

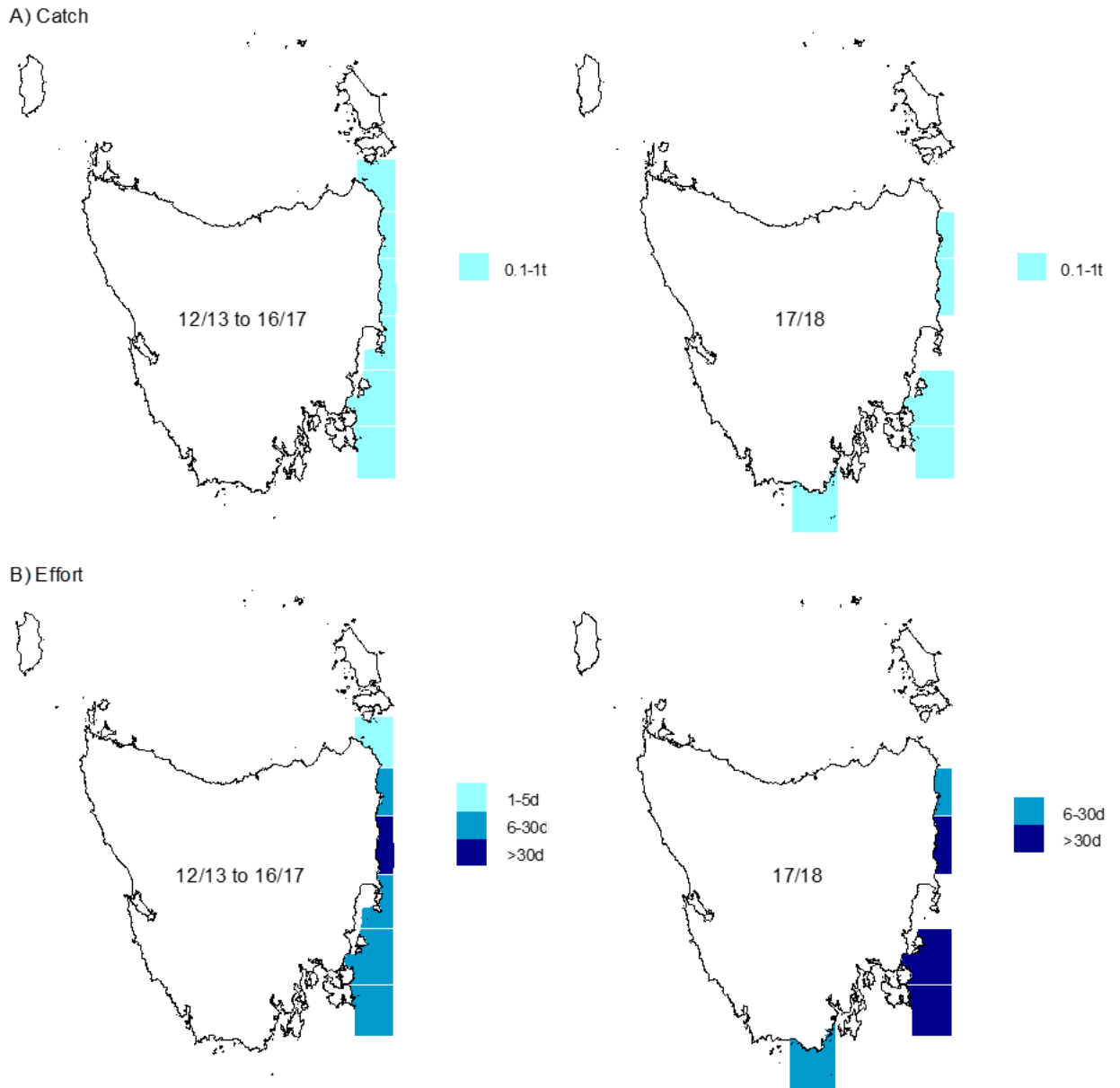
A decline in fishing effort has mirrored the declining catches for the species (Fig. 14.1B). Catch rates have declined since the early 2000s but remained stable at a reduced level since 2007/08 (Fig. 14.1C).

### **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian scalefish fishery, gillnetting was considered a medium risk activity with regard to Jackass Morwong (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.



**Figure 14.1** A) Annual commercial catch (t) by gear (left) and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. GN=gillnet.



**Figure 14.2** (A) Jackass Morwong catches (t) and (B) effort (days) for gillnet, hand-line and drop-line by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	<ul style="list-style-type: none"> <li>Catch &gt; 3<sup>rd</sup> highest catch value from the reference period (18.7 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Catch &lt; 3<sup>rd</sup> lowest catch value from the reference period (13.1 t)</li> </ul>	Yes	↓ 9.8 t (74.7%)
	<ul style="list-style-type: none"> <li>Latest recreational catch estimate &gt; recreational catch estimate from the reference period (31.9 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Proportion of recreational catch to total catch &gt; previous proportion estimate (64.2% in 2007/08)</li> </ul>	Yes	Latest estimate (2012/13): ↑ 88.5%
<b>Biomass</b>	<ul style="list-style-type: none"> <li>Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period</li> <li>(-0.0017)</li> </ul>	No	
<b>Stock stress</b>	<ul style="list-style-type: none"> <li>Significant change in the size/age composition of commercial catches</li> </ul>	Not assessed	
	<ul style="list-style-type: none"> <li>Significant numbers of unhealthy fish landed</li> </ul>	No	

## Stock status

**SUSTAINABLE**

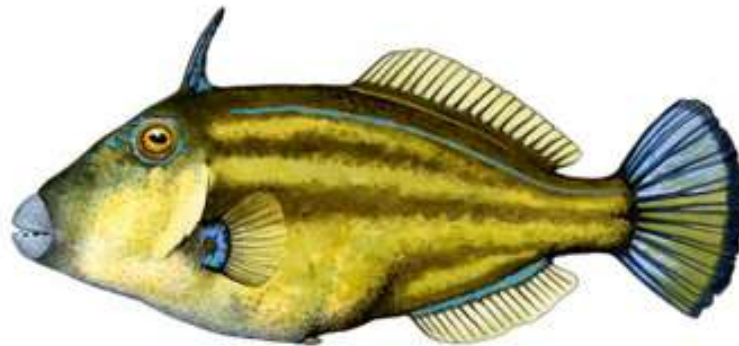
Despite commercial catches being at their highest level since 2009/10, the lowest catch reference point was breached, which reflects the ongoing declining commercial catches relative to historical levels in Tasmanian waters. Linked to this decline, one of the recreational indicators was breached, indicating that the proportion of recreational catch to total catch had increased in 2012/13 since the last recreational survey (2007/08).

A single east Australian stock of Jackass Morwong is shared between the Commonwealth and Tasmania with catch and catch rates declining in Commonwealth fisheries in a fashion similar to that observed in the Tasmanian fishery. This was driven by a prolonged period of reduced recruitment that is believed to be a result of climate induced changes to ocean currents in eastern Tasmania (Wayte 2013). Due to the extended larval phase of Jackass Morwong, these changes have meant that larvae are often dispersed away from Tasmanian waters with successful recruitment being correlated with years where oceanographic conditions are comparable to historical averages (Wayte 2013). While the Jackass Morwong stocks were considered overfished in the late 2000s, since 2011 the stock has been classified as not overfished and not subject to overfishing (Woodhams et al. 2013, Flood et al. 2014, Patterson et al. 2015, 2016, 2017, 2018). This assessment was due to a reduction of catches for the species as a response to management action in the Commonwealth fishery and revision of the stock assessment model. The total catch (recreational and commercial) of Jackass Morwong for Tasmania (3.3 t) is low compared to the Commonwealth catch (185 t) in 2017/18, and the stocks are most likely shared. Although there has not been a formal Commonwealth assessment since 2012, unpublished Commonwealth statistics indicate the east coast stock is rebuilding under current total allowable catch levels and there has not been further reductions required. As the Fishery Status Reports describe both the stock biomass and fishing mortality as sustainable (Patterson et al. 2018), this ranking has been applied to the Tasmanian fishery which shares the same stock.

# 15. Leatherjacket

*Monacanthidae* family

<b>STOCK STATUS</b>	UNDEFINED
A number of Leatherjacket species are found inshore around Tasmania's coastline and are not differentiated in logbooks. Leatherjackets are a by-product species and are not actively targeted due to a lack of market demand. Catch is therefore not a good indicator of abundance and there is little biological information to confidently classify the status of Leatherjacket stocks.	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Leatherjacket*  
Source: DPIPW (by Peter Gouldthorpe)

## Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Seagrass and reefs. Down to 200 m depending on species.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>Australia is the centre of diversity for this family with more than half of the estimated 90 species occurring here, mainly in temperate areas.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Epiphytes attached to seagrass, algae, fish flesh, molluscs and crustaceans depending on species. Many species are omnivorous.</li> </ul>	FishBase (2013)
Movement and stock structure	<ul style="list-style-type: none"> <li>Many species are site-attached.</li> </ul>	Barrett (1995a)
Natural mortality	<ul style="list-style-type: none"> <li>Undefined for most species.</li> </ul>	
Maximum age	<ul style="list-style-type: none"> <li>No information.</li> </ul>	
Growth	<ul style="list-style-type: none"> <li>Maximum length: from 90 mm to 600 mm.</li> </ul>	Edgar (2008)
Maturity	<ul style="list-style-type: none"> <li>Little information.</li> </ul>	
Spawning	<ul style="list-style-type: none"> <li>Little information.</li> </ul>	

Early life history	<ul style="list-style-type: none"> <li>• Little information.</li> </ul>	
Gillnet post release survival	<ul style="list-style-type: none"> <li>• High: 95%</li> </ul>	Lyle et al (2014b)

## Background

Leatherjackets are a by-product of fish traps and they are also caught in netting operations but are generally discarded. While Leatherjackets are consumed on the mainland, there is little market demand for the species in Tasmania.

<b>FISHING METHODS</b>	Mainly fish trap, also gillnet and hand-line.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence).</li> <li>• Recreational gear licence (graball and/or mullet net).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Possession limit of 20 and bag limit of 10 individuals for recreational fishers.</li> <li>• Minimum size: 20 cm.</li> </ul>
<b>MAIN MARKET</b>	Local.

## Current assessment

### **Catch, effort and CPUE**

Leatherjacket catches have declined continuously since the early 2000s, reaching a minimum of 1.3 t in 2015/16 t (Fig. 15.1A). Total commercial catches in 2017/18 were 2.6 t, consistent with 2016/17 but up from recent years. While catches from gillnets have consistently remained at low levels since 1995/96, catches from fish traps have declined over time. Leatherjackets are now primarily caught on the east and southeast coasts (Fig. 15.2).

Leatherjackets are also caught by the recreational sector, with catch estimates in recent surveys at a similar level to commercial catches (Fig. 15.1A). Estimates were 8.2 t in 2000/01 (Lyle 2005), 2.6 t in 2007/08 (Lyle et al. 2009), 2.3 t in 2009/10 and 1.8 t in 2012/13 (Lyle et al. 2014a).

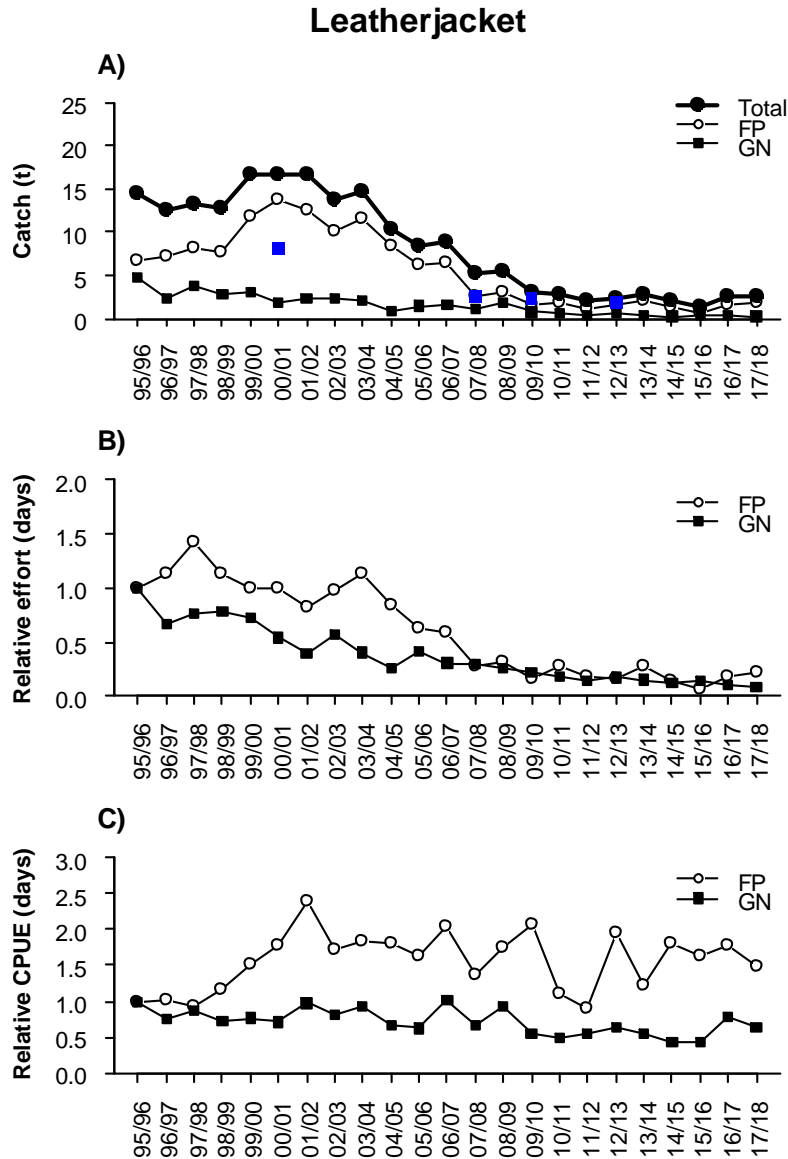
Both fish trap and gillnet fishing effort has decreased through time (Fig. 15.1B). Fish trap effort (in gear units and fishing day) dropped significantly around 2007/08. Abalone gut was the preferred bait used in fish traps; however, a ban on its use has been in place since 2008 to prevent the spread of abalone viral ganglioneuritis (AVG). Following this ban, fishing effort using fish traps declined as other baits are considered inferior.

Catch rates in days fished have remained relatively stable over time for gillnets, while fluctuating for fish traps (Fig. 15.1C).

### **Ecological Risk Assessment**

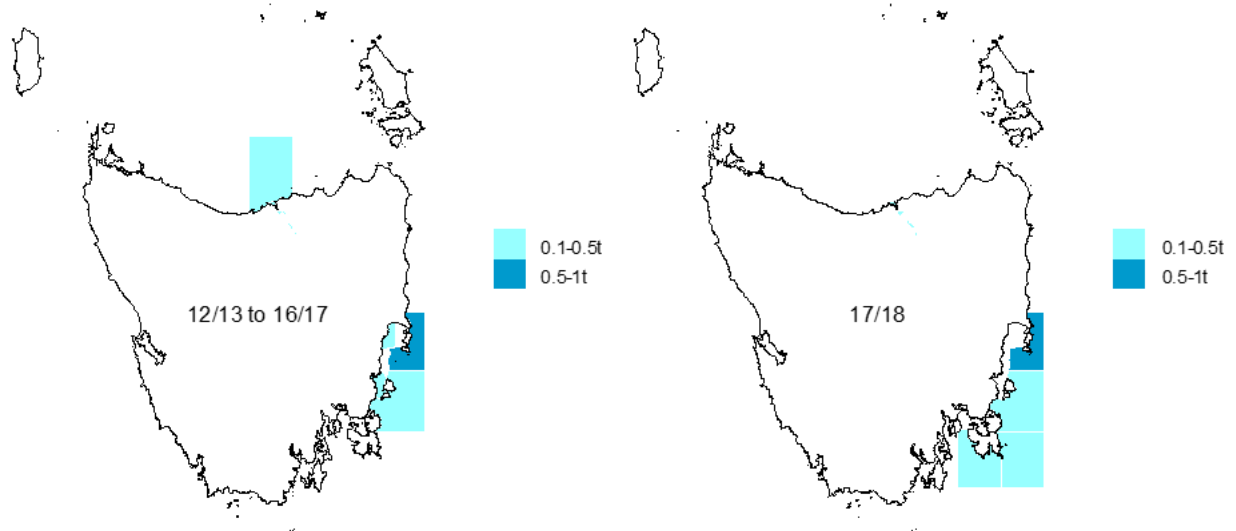
In the 2012/13 ERA of the Tasmanian scalefish fishery, fish trapping was considered a very low risk to Leatherjacket species, which is the main by-product of fishing for Wrasse. This is because

the current catch of Leatherjacket is low due to the ban on using abalone gut for bait, which significantly reduced trapping effort. Risks to non-retained species and the general ecosystem were assessed as either low or negligible (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.

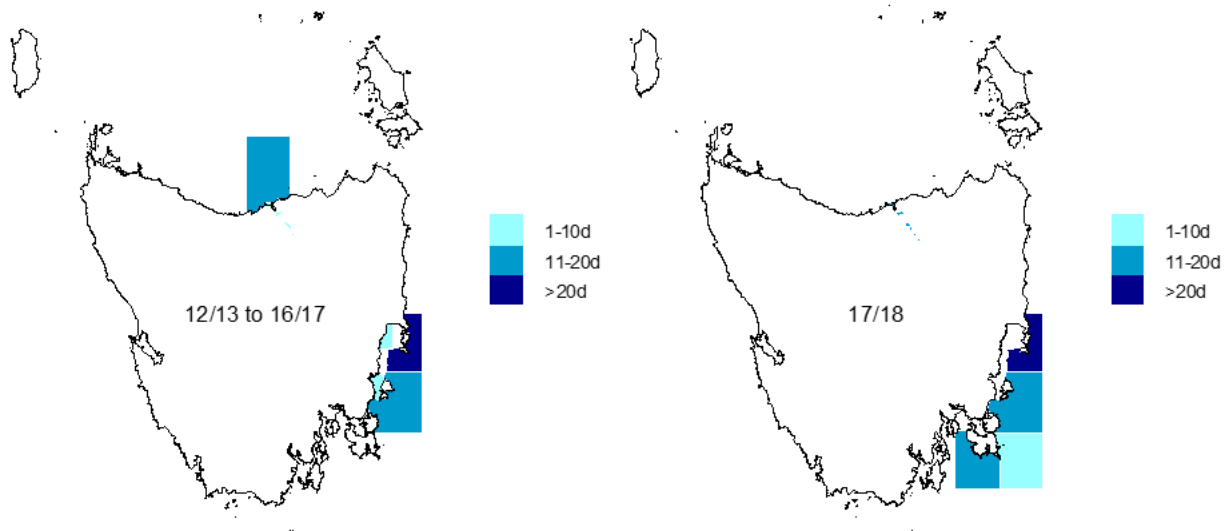


**Figure 15.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. FP=fish trap, GN=gillnet.

A) Catch



B) Effort



**Figure 15.2** (A) Leatherjacket catches (t) and (B) effort (days) for fish trap and gillnet by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (16.6 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (10.4 t)	Yes	↓ 7.8 t (75.3%)
	• Latest recreational catch estimate > recreational catch estimate from the reference period (8.2 t)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate (38.0% in 2007/08)	Yes	Latest estimate (2012/13): ↑ 41.9%
<b>Biomass</b>	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0015)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

UNDEFINED

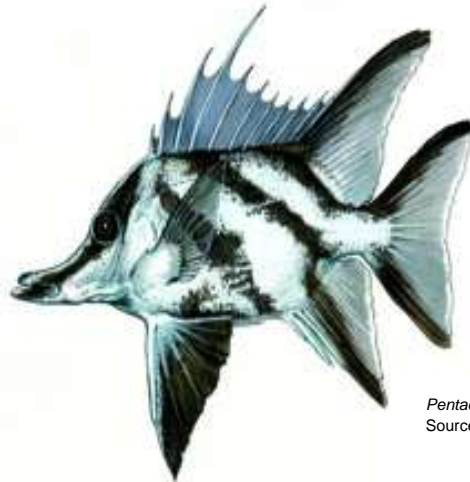
The low landings of Leatherjacket in 2017/18 meant that the lowest catch and the proportion of recreational to commercial catch reference points were breached. These decreased catches are, however, the result of a decline in the use of fish traps and a lack of demand and not an indication of relative abundance.

Leatherjackets tend to be site-attached and have limited home ranges. Two decades of monitoring eastern Tasmanian Marine Protected Areas (MPAs) show that there is no significant difference in abundances inside or outside of MPAs for several Leatherjacket species (e.g. Brown Striped Leatherjacket, Toothbrush Leatherjacket). Leatherjackets also have a very high post release survival following capture in gillnets (Lyle et al. 2014b). These results, along with the low landings, suggest that fishing is unlikely to have a significant impact on Leatherjacket populations. There is however insufficient information to confidently classify the status of Leatherjacket stocks, especially as multiple species are involved.

# 16. Longsnout Boarfish

## *Pentaceropsis recurvirostris*

<b>STOCK STATUS</b>	UNDEFINED
Boarfish are a by-product species of Banded Morwong fishing with low catches due to the large minimum legal size. There is insufficient information available to confidently classify this stock.	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Pentaceropsis recurvirostris*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Exposed reef. Between 4 and 260 m depth.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>From mid New South Wales to southern Western Australia, and around Tasmania.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Brittle stars, polychaetes and brown algae</li> </ul>	Edgar (2008) Scott et al. (1974)
Movement and stock structure	<ul style="list-style-type: none"> <li>Unknown</li> </ul>	
Natural mortality	<ul style="list-style-type: none"> <li>Unknown</li> </ul>	
Maximum age	<ul style="list-style-type: none"> <li>Unknown</li> </ul>	
Growth	<ul style="list-style-type: none"> <li>Maximum length: 61 cm</li> </ul>	Edgar (2008)
Maturity	<ul style="list-style-type: none"> <li>Unknown</li> </ul>	
Spawning	<ul style="list-style-type: none"> <li>Unknown</li> </ul>	
Early life history	<ul style="list-style-type: none"> <li>Unknown</li> </ul>	

Gillnet post release survival	<ul style="list-style-type: none"> <li>High: 99.7%</li> </ul>	Lyle et al (2014b)
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## Background

Boarfish are a by-product of gillnetting, primarily for Banded Morwong. The main Boarfish species in catches is the Longsnout Boarfish (*Pentaceropsis recurvirostris*). Due to the large minimum legal size and the requirement to release undersized fish, Longsnout Boarfish are regularly discarded. Longsnout Boarfish are considered good eating and there is a small local market for them. Shark nets also catch Boarfish as a by-product in State waters; these catches have been reported to the Commonwealth since 2000/01.

<b>FISHING METHODS</b>	Gillnet, shark net (in the past).
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>Gear licence (Scalefish fishing licence).</li> <li>Recreational gear licence (graball and/or mullet net).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>Trip limit of 50 kg for commercial fishers.</li> <li>Possession limit of 4 and bag limit of 2 individuals for recreational fishers.</li> <li>Minimum size: 45 cm.</li> </ul>
<b>MAIN MARKET</b>	Mainly local.

## Current assessment

### **Catch, effort and CPUE**

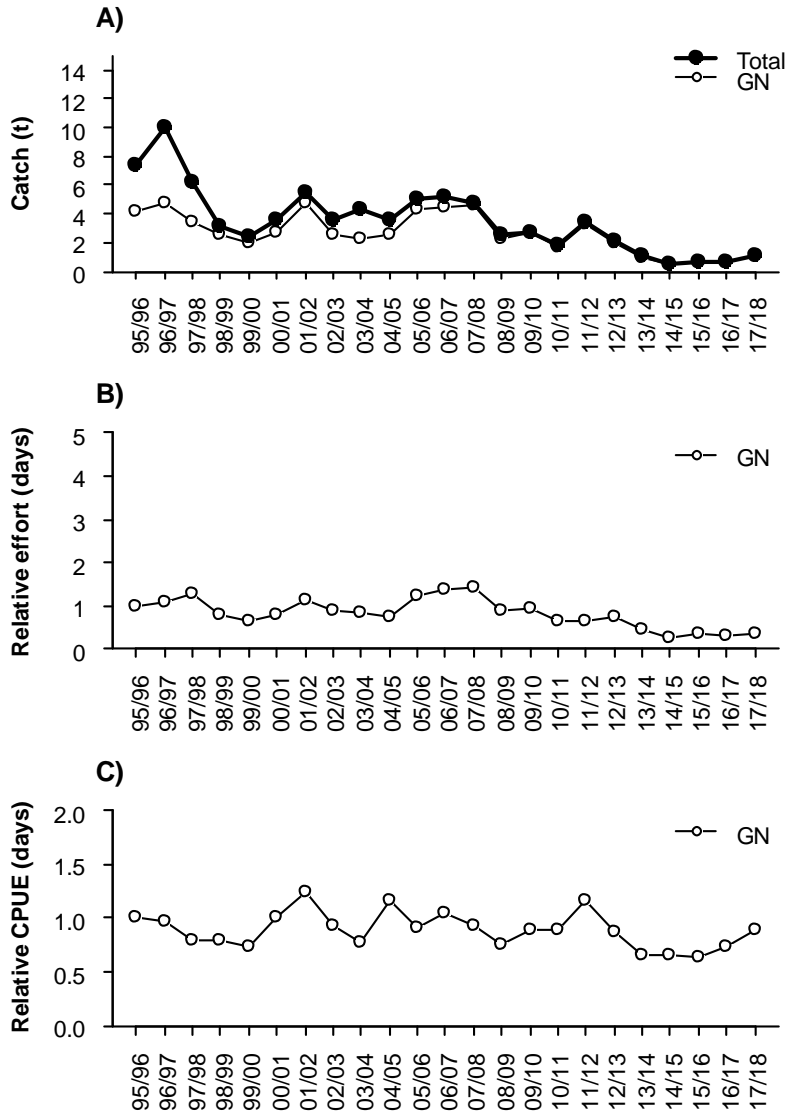
In Tasmania, Boarfish catches are now primarily derived from gillnet (Fig. 16.1A). Catches have been declining through time but appear to have stabilised since 2008/09 with landings of 1.1 t in 2017/18 (Fig. 16.1A). Catches are taken exclusively from the east coast (Fig. 16.2). Boarfish are not caught by rod and line and no recreational catch estimates are available for gillnet and spearfishing catch of this species. However, fewer than 1000 individuals were recorded (both kept and released) in the 2012/13 survey (Lyle et al. 2014a) indicating that Boarfish are not a common recreational species.

After peaking in 1997/98, commercial gillnet effort has remained stable (Fig. 16.1B). Catch rates have remained relatively stable over time (Fig. 16.1C).

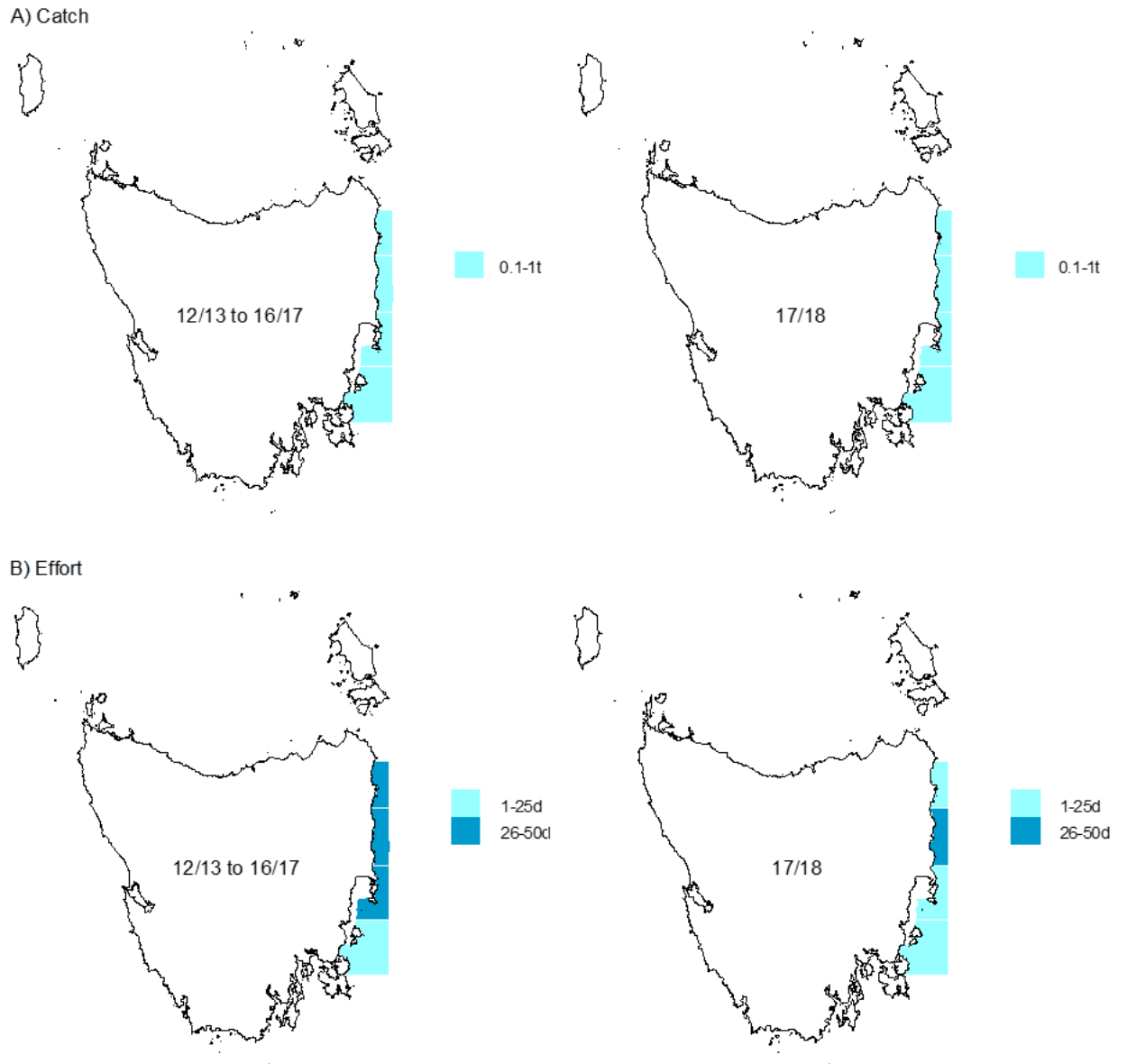
### **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian scalefish fishery, gillnetting was considered a medium risk activity with regard to Boarfish, which is a by-product of Banded Morwong fishing (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.

### Boarfish



**Figure 16.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. GN=gillnet.



**Figure 16.2** (A) Boarfish catches (t) and (B) effort (days) for gillnet fishing by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (6.2 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (3.6 t)	Yes	↓ 2.5 t (69.6%)
	• Latest recreational catch estimate > recreational catch estimate from the reference period	Not estimated	
	• Proportion of recreational catch to total catch > previous proportion estimate	Not estimated	
<b>Biomass</b>	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0009)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

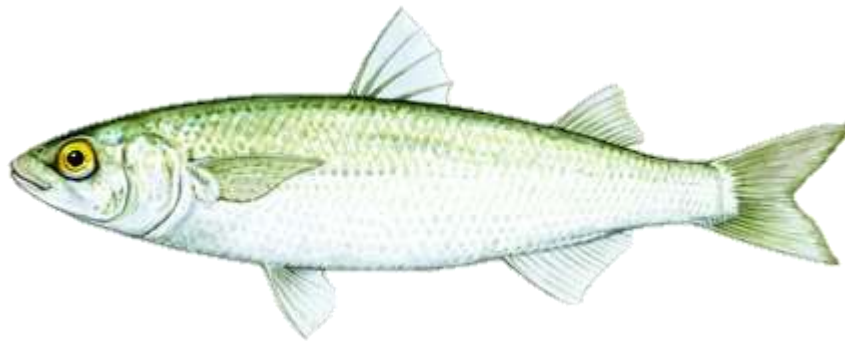
UNDEFINED

The reference point associated with low catch was breached in 2017/18. Catches are lower now due to reduced gillnetting effort compared with during in the reference period, noting that catch rates have remained relatively stable over time. Boarfish are a by-product that is taken in very small quantities. In addition to catches taken in State waters, there is also a by-product fishery from Commonwealth shark netting activity. The high size limit means that many individuals are released and the species has very high post-release survival (Lyle et al. 2014b). There is, however, insufficient information available to confidently classify this stock.

# 17. Yelloweye Mullet

*Aldrichetta forsteri*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
<p>Catches are at low levels but this is unlikely to be indicative of abundance. Yelloweye Mullet are most abundant in estuarine habitats, where netting is prohibited or restricted, thereby providing a high degree of protection throughout most of their range. It is therefore unlikely that the stock is recruitment impaired or that current fishing pressure is too high as to cause the stock to become recruitment impaired.</p>	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Aldrichetta forsteri*  
Source: DPIPW (by Peter Gouldthorpe)

## Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Sheltered sand, seagrass, up to 20 m depth. May ascend rivers into freshwaters.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>Western Australia (Shark Bay) to New South Wales, and around Tasmania, New Zealand.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Planktonic animals for juveniles, benthic crustaceans and molluscs for medium-sized fish and almost exclusively algae for larger fish.</li> </ul>	Edgar (2008)
Movement and stock structure	<ul style="list-style-type: none"> <li>Schooling fish.</li> <li>No genetic studies but there appears to be two populations (eastern Australia and western Australia).</li> </ul>	Kailola et al. (1993)
Natural mortality	<ul style="list-style-type: none"> <li><i>M</i> estimated at 0.66 (New Zealand).</li> </ul>	Paul and Taylor (1998)
Maximum age	<ul style="list-style-type: none"> <li>7 years</li> </ul>	Curtis and Shima (2005)
Growth	<ul style="list-style-type: none"> <li>Maximum length of 50 cm.</li> <li>Maximum weight: 950 g</li> <li>Differential growth between sexes and locations.</li> </ul>	Edgar (2008) Curtis and Shima (2005)

	<ul style="list-style-type: none"> <li>Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math> where <math>L</math> is the fork length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</li> </ul> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_{0\infty}</math></th> </tr> </thead> <tbody> <tr> <td>Combined</td> <td>40</td> <td>0.51</td> <td>-0.03</td> </tr> </tbody> </table> <ul style="list-style-type: none"> <li>Length-weight relationship was estimated at <math>W = 0.000239 L^{3.2}</math> for females and males combined where <math>W</math> is weight (g) and <math>L</math> is the fork length (cm).</li> </ul>	Sex	$L_{\infty}$	$k$	$t_{0\infty}$	Combined	40	0.51	-0.03	<p>Gorman (1962) Last et al. (1983) Chubb et al. (1981)</p>
Sex	$L_{\infty}$	$k$	$t_{0\infty}$							
Combined	40	0.51	-0.03							
Maturity	<ul style="list-style-type: none"> <li>2–3 years</li> </ul>	Kailola et al. (1993)								
Spawning	<ul style="list-style-type: none"> <li>Form large aggregations prior to spawning.</li> <li>Spawn in coastal waters in summer and autumn, probably in estuaries.</li> <li>Fecundity between 125,000 and 630,000 eggs.</li> <li>Pelagic eggs.</li> </ul>	<p>Chubb et al. (1981) Kailola et al. (1993)</p>								
Early life history	<ul style="list-style-type: none"> <li>Juveniles enter estuaries and sheltered bays when they are 3–4 cm long, and remain there until they reach 25–30 cm tail length.</li> <li>As they grow older, animals gradually move to more open coastal waters.</li> </ul>	Kailola et al. (1993)								
Gillnet post release survival	<ul style="list-style-type: none"> <li>Low: 10%</li> </ul>	Lyle et al (2014b)								

## Background

Mullet are occasionally targeted by netting, specifically beach and purse seining as well as small mesh net. The vast majority of catch is considered to be Yelloweye Mullet, but it is possible that some of the catch includes Sea Mullet (*Mugil cephalus*). Mullet are also targeted by recreational fishers using rod and line or small mesh gillnets called ‘mullet nets’.

<b>FISHING METHODS</b>	Mostly beach seine, also small mesh net (mullet net for recreational) and purse seine
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>Gear licence (Scalefish fishing licence, small mesh gillnet licence)</li> <li>Recreational gear licence (graball and/or mullet net)</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>Possession limit of 30 and bag limit of 15 individuals for recreational fishers – all mullet species combined.</li> <li>Minimum size: 25 cm</li> </ul>
<b>MAIN MARKET</b>	Mostly local

## **Current assessment**

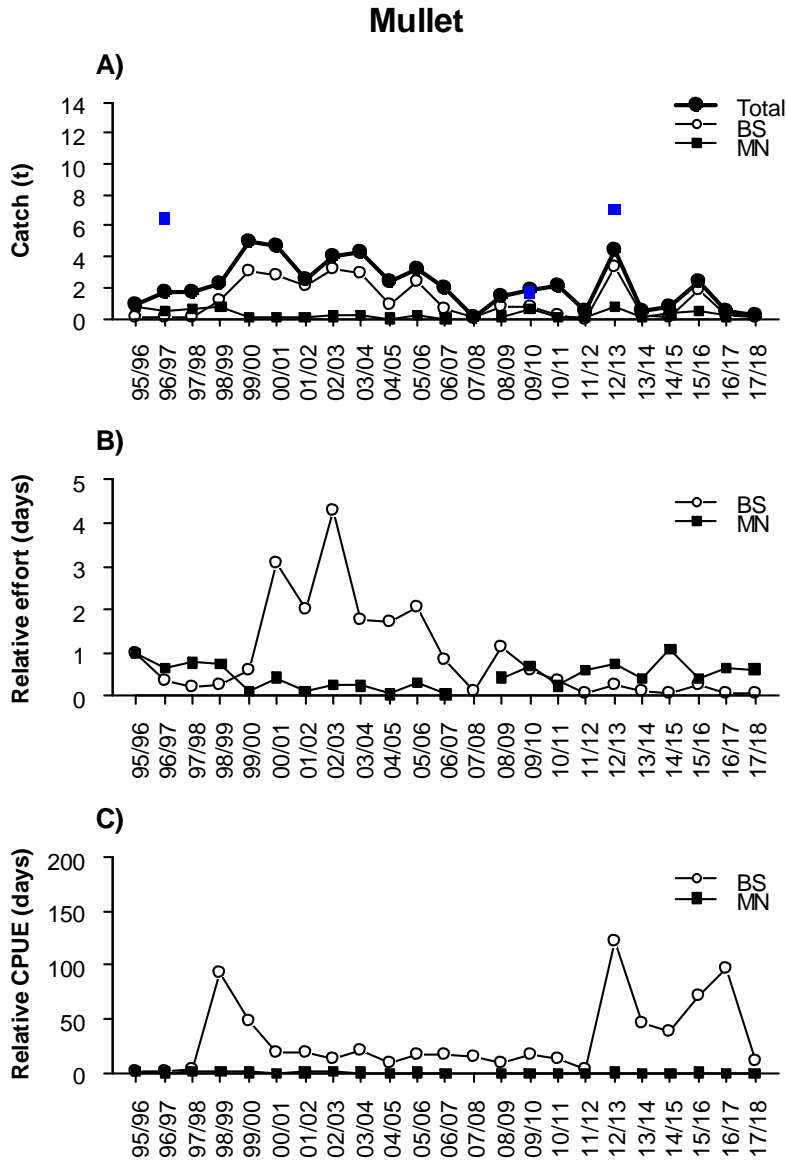
### **Catch, effort and CPUE**

After peaking in 1999/2000, commercial mullet catches have decreased and stabilised at less than 5 t since 2007/08. The commercial catch in 2017/18 was 0.3 t, a level comparable to that of other recent years (with the exception of 2012/13, when the catch was relatively large), although near the historical low of 0.1 t in 2007/08 (Fig. 17.1A). While beach seine has historically been the dominant fishing method used to harvest mullet, small mesh net was the dominant method in 2016/17, with 0.2 t harvested using this method. Fishing activity was traditionally focused in the northwest and off the east and southeast coasts, but in 2017/18 most of the catch and effort occurred in the Tamar Estuary (Fig. 17.2). Recreational catches of mullet were estimated at 6.5 t in 1996/97, 30 t in 2000/01, 1.7 t in 2009/10 and 7.1 t in 2012/13 (Lyle et al. 2014a) (Fig. 17.1A).

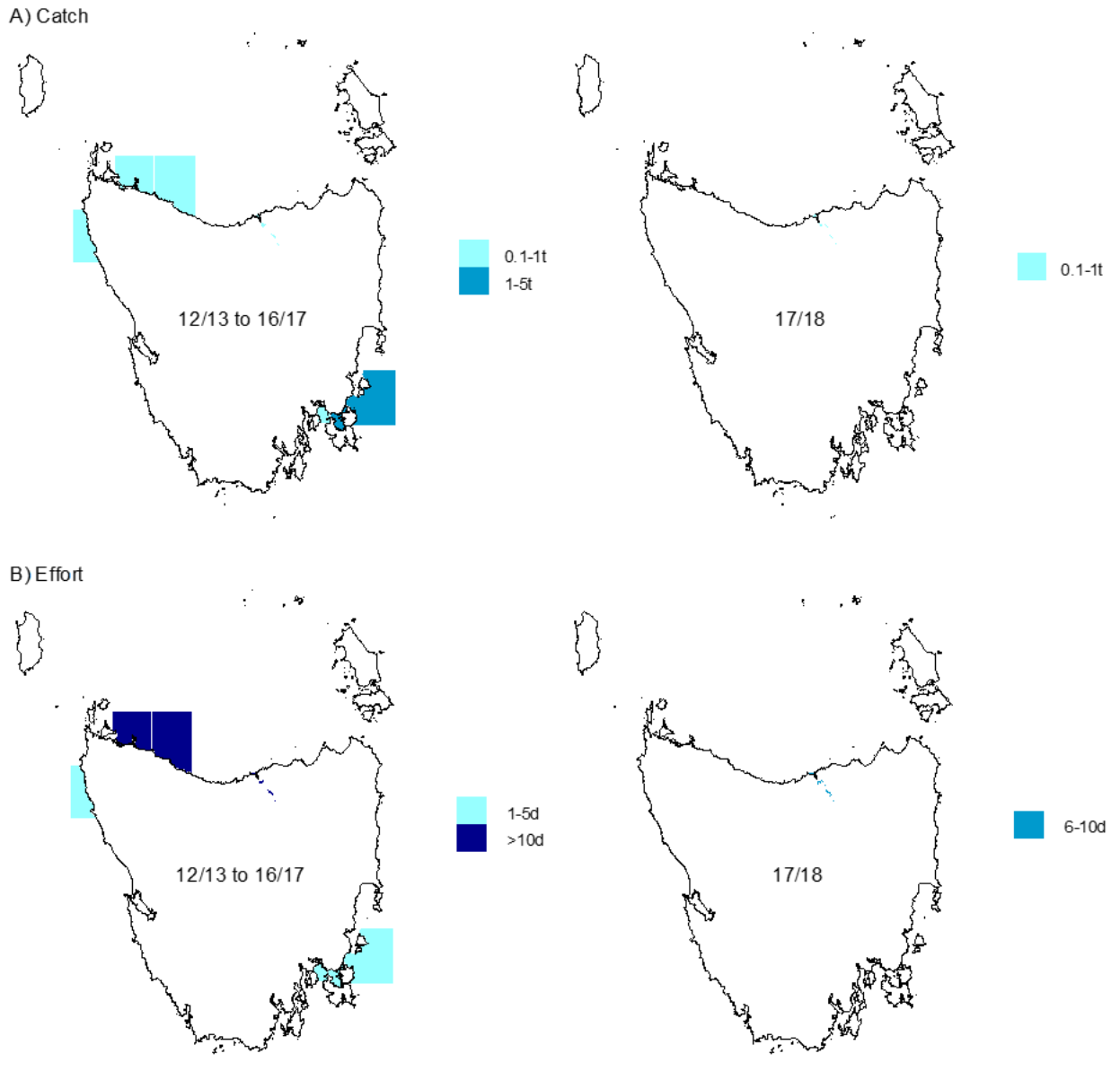
Beach seine effort was relatively stable until 2005/06 after which it declined rapidly, remaining at low levels until 2015/16 (Fig. 17.1B). Catch rates for beach seining have remained relatively constant over time with a sharp increase in 2012/13, 2015/16, and 2016/17. Catch rates for beach seining in 2017/18 were comparable to those observed between 2000/01 and 2011/12 (Fig. 17.1C).

### **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian scalefish fishery, beach seining was considered a negligible risk activity with regard to Yelloweye Mullet due to the low annual catch and the fact they are widespread and not targeted in estuarine habitats (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.



**Figure 17.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. BS=beach seine, MN=small mesh net.



**Figure 17.2** (A) Mullet catches (t) and (B) effort (days) for by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (4.3 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (1.7 t)	Yes	↓ 1.5 t (85.8%)
	• Latest recreational catch estimate > recreational catch estimate from the reference period (30.0 t)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate (73.3% in 2007/08)	No	
<b>Biomass</b>	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.013)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

**SUSTAINABLE**

With the exception of 2012/13, catches of mullet (predominantly Yelloweye Mullet) have been stable at low levels for the past six years, following a decrease in effort in the traditional fishing grounds in northern Tasmania. Yelloweye Mullet are by far the most abundant mullet species in southern Australia and are extremely abundant in estuaries (Edgar 2008). Limited commercial fishing or no recreational gillnetting occurs in most Tasmanian estuaries meaning Yelloweye Mullet are afforded a high degree of protection throughout much of their range. Further, low recreational catches indicate that they are not often targeted with rod and line in estuaries and as such current catches are highly unlikely to result in recruitment impairment. Given the low commercial and recreational catches and high degree of protection afforded to the habitat of Yelloweye Mullet the stock is classified as sustainable in Tasmanian waters.

# 18. Snook

## *Sphyraena novaehollandiae*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
<p>Snook catches declined slightly in 2017/18 relative to 2016/17 but were consistent with long-term averages. Catch rates are an unreliable estimate of abundance due to the species not being actively targeted. Recent biological analyses indicate that the current level of fishing mortality is unlikely to cause the stock to become recruitment impaired.</p>	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Exposed reef, sand, seagrass, offshore waters. Down to depth of 20 m.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li>Western Australia to southern Queensland, and northern Tasmania.</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Fish.</li> </ul>	Coleman and Mobley (1984) Scott et al. (1974)
Movement and stock structure	<ul style="list-style-type: none"> <li>Highly migratory pelagic species that often occurs in shoals of 50 or more individuals.</li> </ul>	Kailola et al. (1993)
Natural mortality	<ul style="list-style-type: none"> <li><math>M = 0.24</math>.</li> </ul>	Webb (2017)
Maximum age	<ul style="list-style-type: none"> <li>19 years.</li> </ul>	Webb (2017)
Growth	<ul style="list-style-type: none"> <li>Maximum length of 1.1 m, maximum weight of 5.6 kg, maximum age of about 20 years.</li> </ul>	Edgar (2008) Kailola et al. (1993)
Maturity	<ul style="list-style-type: none"> <li>Around 42 cm in length.</li> </ul>	Bertoni (1995)
Spawning	<ul style="list-style-type: none"> <li>Assumed to take place from October to January.</li> </ul>	Kailola et al. (1993)
Early life history	<ul style="list-style-type: none"> <li>No information.</li> </ul>	
Gillnet post release survival	<ul style="list-style-type: none"> <li>NA</li> </ul>	

## Background

Two separate species of 'Pike' are caught in Tasmania, the Longfin Pike (*Dinolestes lewini*, Dinolestidae) and Snook (*Sphyræna novaehollandiae*, Sphyrænidae). Both species are mainly targeted by trolling and small-mesh net (north coast only), and are also a by-product of beach seining and gillnetting. While there is a local and interstate market for Snook, Longfin Pike are of lesser demand. There are some uncertainties about the correct reporting of the two species in logbooks. The vast majority of 'Pike' catches are likely to be Snook, which is confirmed by anecdotal reports from the industry and thus only this species is assessed.

<b>FISHING METHODS</b>	Troll, also beach seine, gillnet and small mesh net.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence, small mesh gillnet licence).</li> <li>• Recreational gear licence (graball and/or mullet net).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• No possession or bag limits for recreational fishers.</li> </ul>
<b>MAIN MARKET</b>	Local and interstate (Victoria).

## Current assessment

### Catch, effort and CPUE

Snook catches were variable but followed a relatively stable trend from 1998/99 to 2002/03, before declining in the mid-2000s to less than 5 tonnes. Since then catches averaged around 6 to 9 tonnes before declining to a historical low in 2015/16 of 2.4 t (Fig. 18.1A). Catches in 2017/18 were 5.9 t, down slightly from 9.4 t landed in 2016/17 but consistent with catches for 2006/07 to 2011/12. In recent years, Snook catch and effort has been concentrated on the north coast and around Flinders Island (Fig. 18.2).

There are no estimates of recreational landings (by weight) but evidence suggests that neither Pike species are a major recreational target (Lyle et al. 2009, Lyle and Tracey 2012) and that around 57% of all Pike caught by recreational fishers are released (Lyle et al. 2009). Nevertheless, in 2012/13, 3,895 Pike were landed by recreational fishers (Lyle et al. 2014a) and if an average weight of 1 kg per fish is assumed, recreational landings in 2012/13 would have been approximately 4 t.

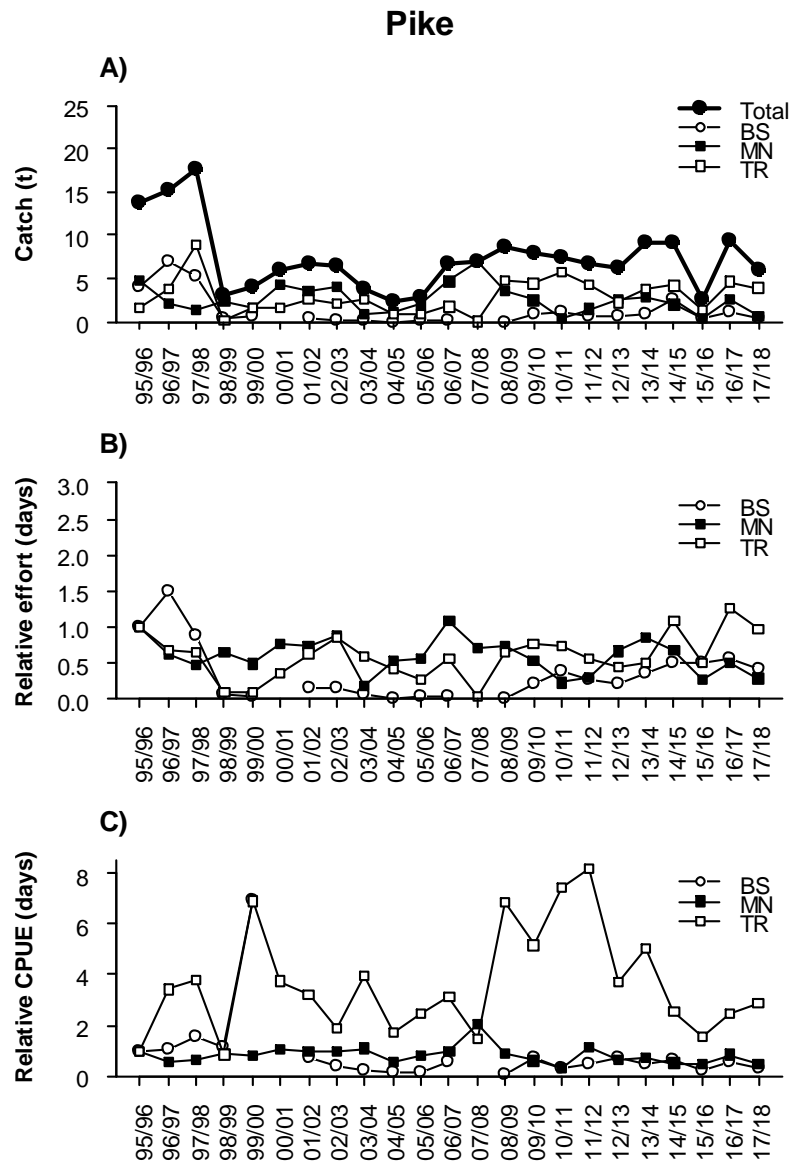
Commercial troll effort for Snook has been variable through time. Since peaking in 2010/11 effort slowly declined, however in 2014/15 there was a sharp increase in days fished before a return to average levels in 2015/16 (Fig. 18.1B). Catch rates increased again in 2016/17 to levels comparable to 2014/15. Beach seine effort has remained stable over time but Snook represent a by-product of this fishing method rather than a target.

Catch rates for troll have been highly variable through time, which is a factor of availability and targeting practices, whereas catch rates for both beach seine and mesh net have been relatively stable (Fig. 18.1C).

### Ecological Risk Assessment

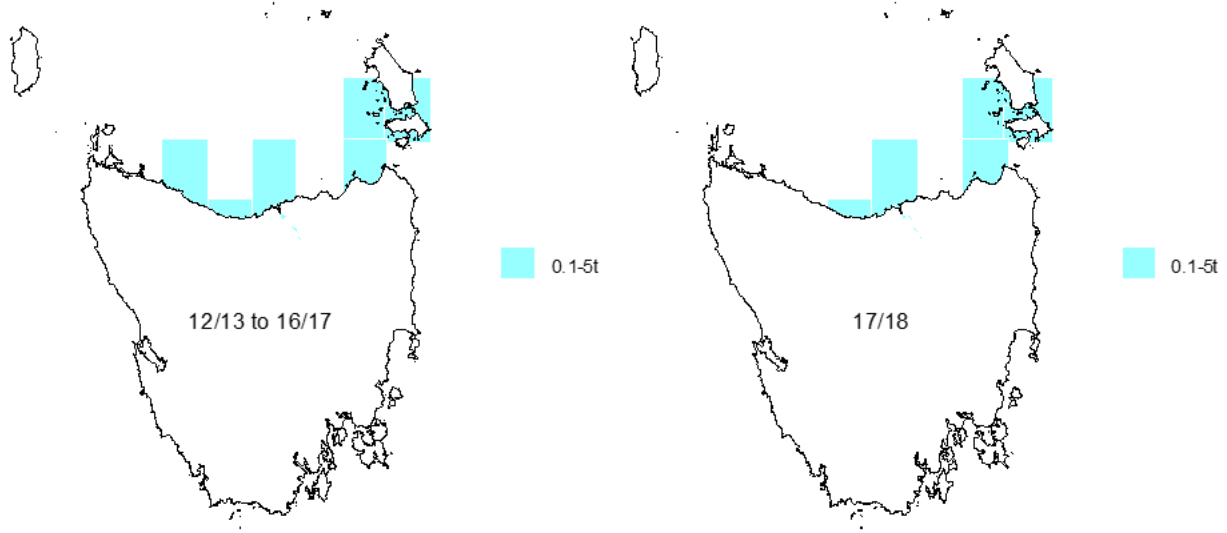
In the 2012/13 ERA of the Tasmanian scalefish fishery, trolling was considered a low risk activity with regard to Snook. While Snook are targeted throughout most of its range by multiple fishing methods the low combined catch is likely to be within sustainable levels. Trolling was also

considered a negligible risk to all other ecosystem components (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.

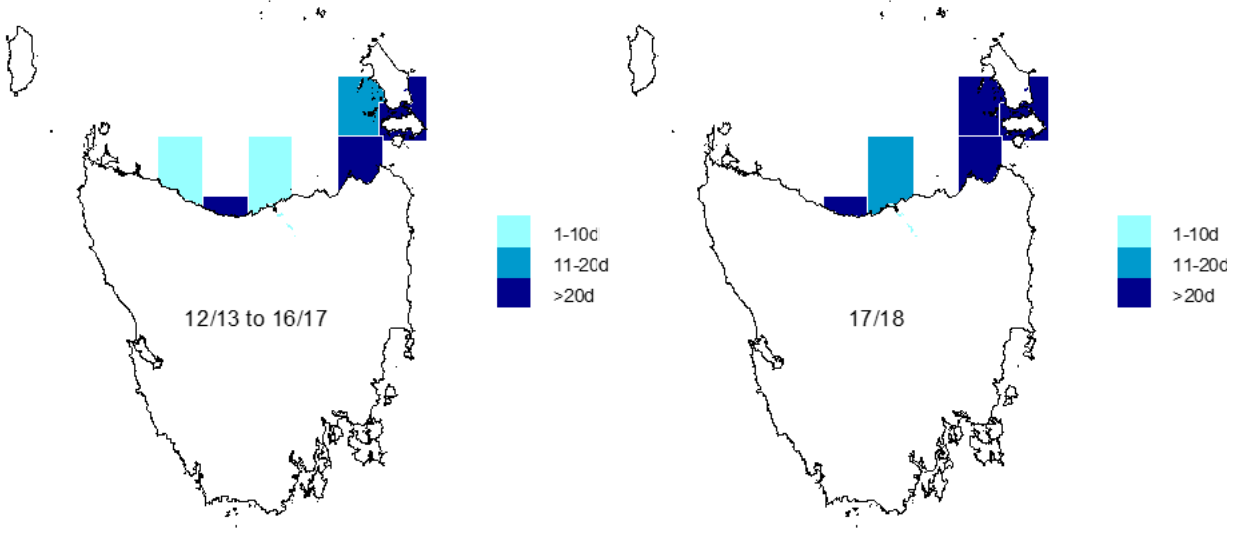


**Figure 18.1** A) Annual commercial catch (t) by gear. B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. BS=beach seine, TR=troll, MN=small mesh gillnet.

A) Catch



B) Effort



**Figure 18.2** (A) Snook catches (t) and (B) effort (days) for troll, beach seine and small mesh net by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (13.7 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (3.2 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (based on numbers)	No	
	• Proportion of recreational catch to total catch > previous proportion estimate	Not estimated	
<b>Biomass</b>	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0035)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

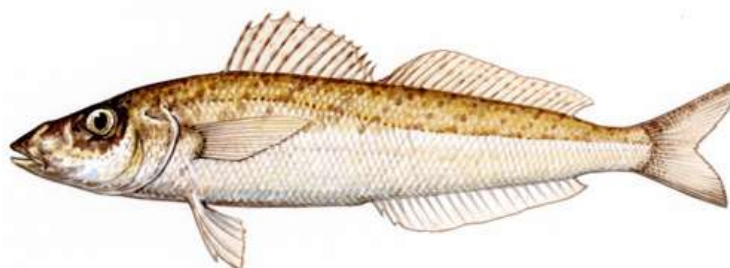
**SUSTAINABLE**

The commercial fishery for Snook is relatively small and mainly limited to the northern part of Tasmania. Moreover, the species is rarely targeted recreationally. A recent fishery-dependent sampling program conducted in the north of the state estimated fishing mortality ( $F$ ) to be low, with  $F$  estimated one quarter of natural mortality ( $M$ ) ( $F=0.06$  yr<sup>-1</sup>,  $M=0.24$  yr<sup>-1</sup>) (Webb 2017). The above evidence indicates that the current level of fishing mortality is unlikely to cause the stock to become recruitment impaired.

# 19. Eastern School Whiting

## *Sillago flindersi*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
<p>This is a predominately Commonwealth-managed species that is classified as not overfished nor subject to overfishing by ABARES for 2017 (Patterson et al. 2018). Tasmanian catches fluctuate due to market demand and species targeting (Tiger Flathead or Eastern School Whiting) by the primary operator. Tasmanian commercial catches represent a small fraction of the Commonwealth commercial catch.</p>	
<b>IMPORTANCE</b>	Minor
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery/Southern and Eastern Scalefish and Shark Fishery (Commonwealth)
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Sillago flindersi*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Coastal lakes, estuaries and along outer coast. Down to 170 m depth.</li> </ul>	Gomon et al. (2008)
Distribution	<ul style="list-style-type: none"> <li>Endemic to southeastern Australia, from southern Queensland to western Victoria, and around Tasmania.</li> </ul>	Gomon et al. (2008)
Diet	<ul style="list-style-type: none"> <li>Feed mainly on crustaceans, amphipods, decapods, mysids and copepods. Juveniles consume mostly copepods.</li> </ul>	Burchmore et al. (1988)
Movement and stock structure	<ul style="list-style-type: none"> <li>There is some evidence of four genetically distinct stocks (two in New South Wales, one in Tasmania and one in Victoria).</li> <li>Commonwealth assessments state that the evidence for separate stocks is weak and manage the species as a single stock.</li> </ul>	Dixon (1987) Morison et al. (2012)
Natural mortality	<ul style="list-style-type: none"> <li>No information but likely to be around <math>M = 0.7</math> based on related species.</li> </ul>	Butcher and Hagedoorn (2003)
Maximum age	<ul style="list-style-type: none"> <li>7 years</li> </ul>	Kailola et al. (1993)
Growth	<ul style="list-style-type: none"> <li>Maximum length: 33 cm SL</li> <li>Growth described by von Bertalanffy growth function <math>L = L_{\infty}(1 - e^{-k(t-t_0)})</math></li> </ul>	Gomon et al. (2008) Tilzey (1994)

	<p>where <math>L</math> is the length (cm), <math>t</math> is the age (years), <math>L_{\infty}</math> is the average maximum length for the species, <math>k</math> is a constant and <math>t_0</math> is the (theoretical) age where length equals zero.</p> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th>Sex</th> <th><math>L_{\infty}</math></th> <th><math>k</math></th> <th><math>t_0</math></th> </tr> </thead> <tbody> <tr> <td>Combined</td> <td>23.9</td> <td>0.46</td> <td>-0.50</td> </tr> </tbody> </table>	Sex	$L_{\infty}$	$k$	$t_0$	Combined	23.9	0.46	-0.50	
Sex	$L_{\infty}$	$k$	$t_0$							
Combined	23.9	0.46	-0.50							
Maturity	<ul style="list-style-type: none"> <li>Reached at 2 years and a size of 14–16 cm FL.</li> </ul>	<p>Hobday and Wankowski (1987)</p> <p>Burchmore et al. (1988)</p>								
Spawning	<ul style="list-style-type: none"> <li>Spring to late summer.</li> <li>Females release between 30,000 and 110,000 eggs in total during the season.</li> </ul>	Hobday and Wankowski (1987)								
Early life history	<ul style="list-style-type: none"> <li>Juveniles inhabit inshore waters.</li> </ul>	FishBase (2013)								
Gillnet post release survival	<ul style="list-style-type: none"> <li>NA</li> </ul>									

## Background

There is a large Commonwealth fishery for School Whiting that has landed 1000–2500 t for the last 30 years with about 75% of the catch taken by Danish seine vessels operating out of Lakes Entrance, Victoria (Morison et al. 2012). The New South Wales state-managed fishery has been increasing in recent years and now accounts for around 60% of the total catch, which has led to some equity disagreements as their catch is deducted from the Commonwealth total allowable catch (Morison et al. 2012). School Whiting have been exploited in Tasmania since the mid-1970s with catches ranging from 20–175 t throughout the 1980s (Kailola et al. 1993). The vast majority of the catch is taken by Danish seine in the south of the State. Danish seine fishing operations target either School Whiting (with Flathead as a by-product) or Flathead (with School Whiting as by-product), which accounts for opposing trends in catches for the two species. School Whiting are mainly marketed and processed in Melbourne.

In recent years, small catches of King George Whiting (*Sillaginodes punctatus*) have been recorded from small mesh netting operations in the north of the State and this species is increasingly becoming a target of both commercial and recreational fishers. King George Whiting tend to be large in Tasmanian waters and there is evidence that northern Tasmania may be an important spawning location for the species – King George Whiting in Victoria are comprised entirely of juveniles with the seed stock previously believed to come from South Australia.

<b>FISHING METHODS</b>	Danish seine.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence, small mesh gillnet).</li> <li>• Danish seine licence (with whiting cod-end endorsement).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Possession limit of 30 and bag limit of 15 individuals for recreational fishers.</li> </ul>
<b>MAIN MARKET</b>	Interstate.

## Current assessment

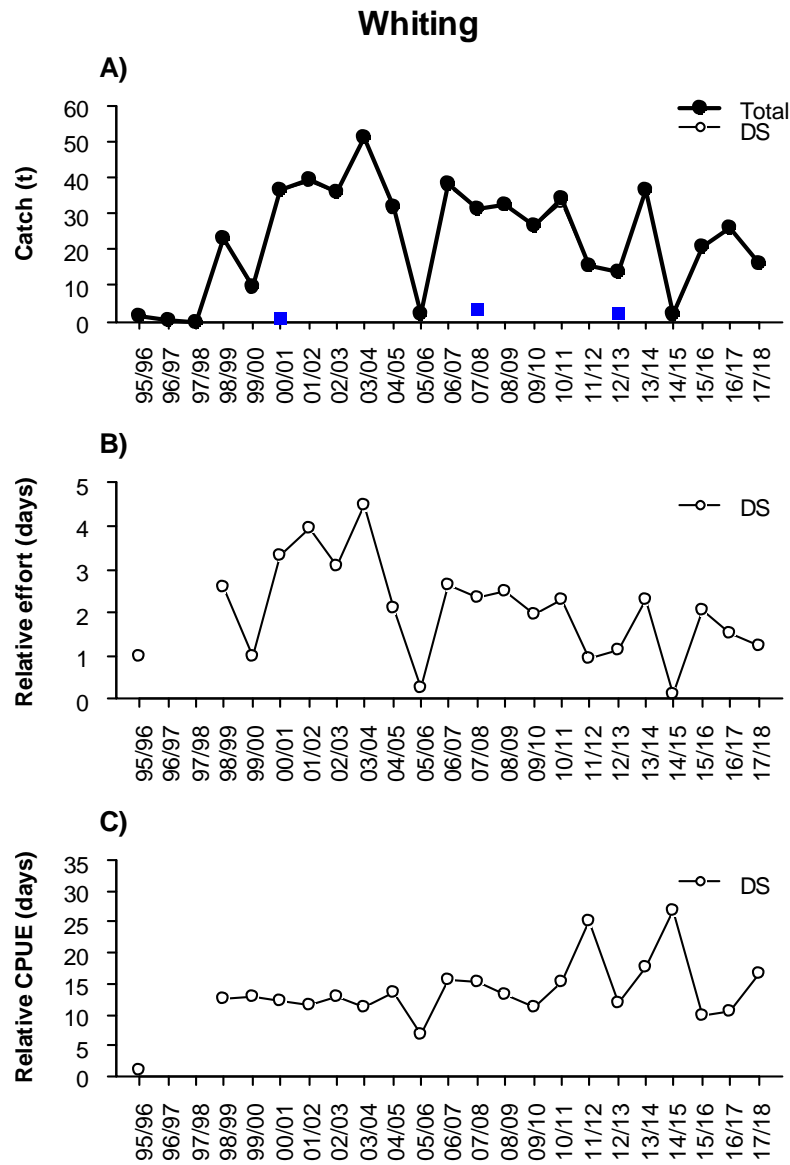
### **Catch, effort and CPUE**

Landings have generally fluctuated between 20 and 50 t per annum since 1998/99, with 16.1 t reported in 2017/18 (Fig. 19.1A). Catches are influenced largely by the targeting practices of a small number of operators. Catches in 2017/18 were concentrated on the southeast coast (in particular the Derwent Estuary) as has historically been the case (Fig. 19.2). Recreational catches on average are low compared to commercial catches, and were estimated at 0.8 t in 2000/01 (Lyle 2005), 3.4 t in 2007/08 (Lyle et al. 2009) and 2.1 t in 2012/13 (Lyle et al. 2014a) (Fig. 19.1A).

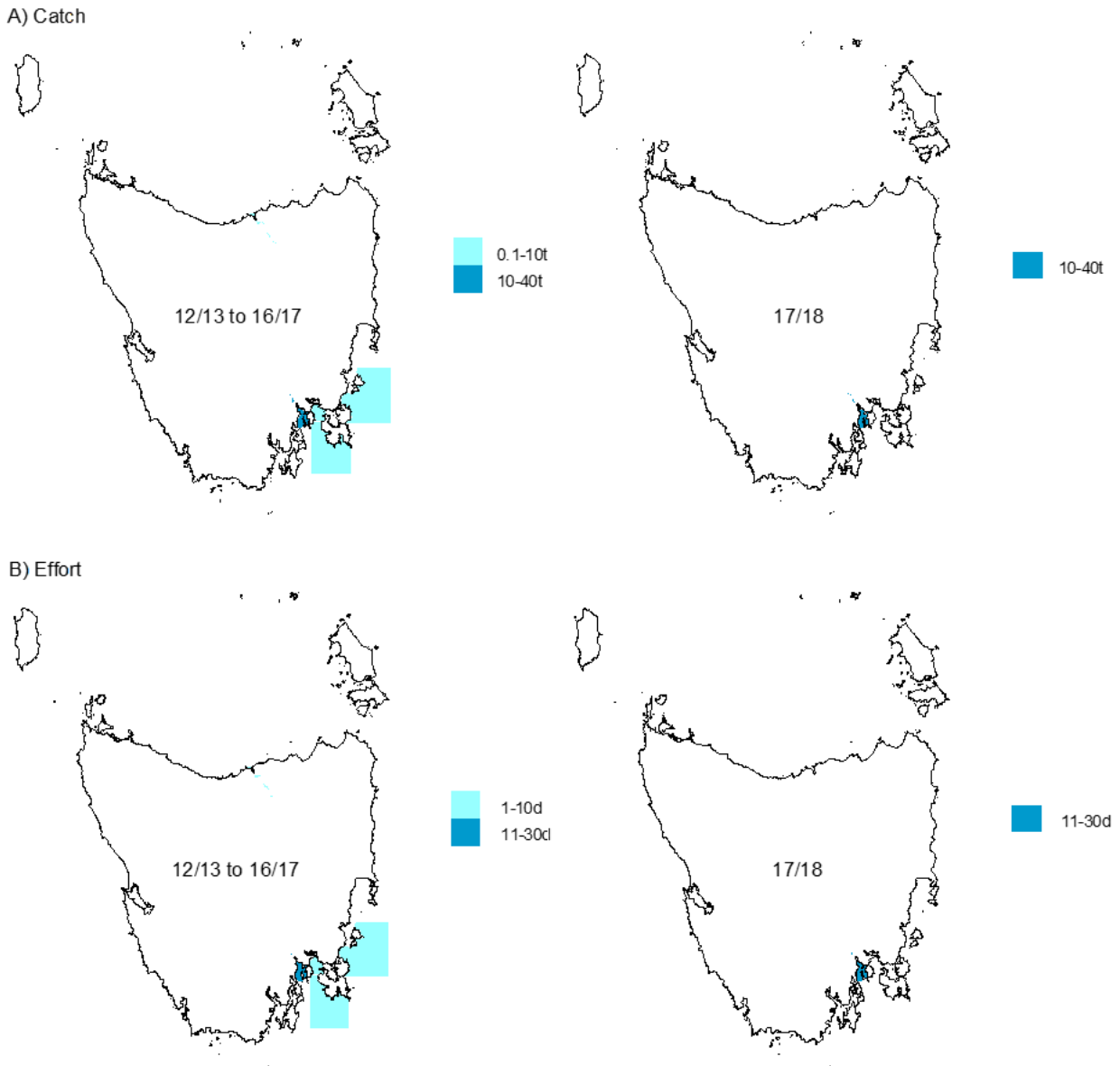
While fishing effort has declined slowly over time there was a small increase in 2015/16 back to 2012/13 levels (Fig. 19.1B). Effort decreased slightly in 2017/18. Catch rates by days fished generally increased from 1995/96 to 2014/15, though increased in 2017/18 relative to the last two fishing years (Fig. 19.1C).

### **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian scalefish fishery, Danish seining was considered a low risk activity with regard to School Whiting due to low effort and fishing activities occurring within only a very small fraction of the species range (Bell et al. 2016). There is no new information to suggest that this has changed otherwise since then.



**Figure 19.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. DS=Danish seine.



**Figure 19.2** (A) School Whiting catches (t) and (B) effort (days) for Danish seine by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (38.1 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (1.4 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (0.8 t)	Yes	↑ 1.3 t (162%)
	• Proportion of recreational catch to total catch > previous proportion estimate (8.7% in 2007/08)	No	
<b>Biomass</b>	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0954)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

**SUSTAINABLE**

Catch, effort and catch rate patterns for School Whiting have been dictated to a large extent by the level of targeting, with the primary fisher switching between targeting of Tiger Flathead or School Whiting depending upon factors such as market demand. Furthermore, the commercial fishery that is dependent on the interstate market and is (generally) based on only a small number of operators, fishing at a low level of effort (days fished).

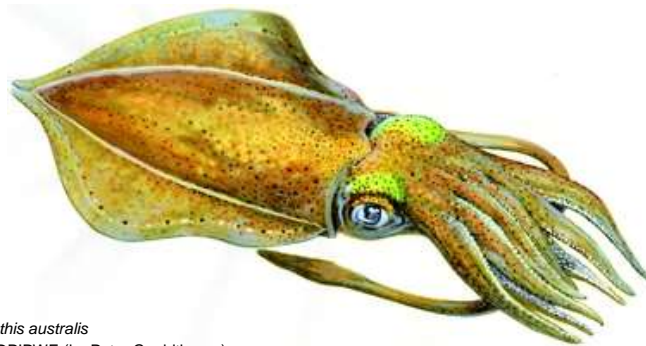
While the most recent recreational catch was higher than during the reference period, catches by the recreational sector remain low and are inconsequential at the scale of the stock.

The Tasmanian fishery harvests only a small proportion of the overall catch when compared with Commonwealth landings (736 t in 2017/18). The latest Fishery Status Report (Patterson et al. 2018) classified School Whiting as sustainable for both biomass and fishing mortality. In Tasmania, current levels of fishing effort are unlikely to cause this species to become depleted or for recruitment to become impaired. Given this, and the sustainable classification of the stock nationally, the stock is classified as sustainable in Tasmania.

# 20. Southern Calamari

## *Sepioteuthis australis*

<b>STOCK STATUS</b>	<b>DEPLETING</b>
<p>State-wide commercial catches in 2017/18 declined by more than 50% from 2016/17 when record high levels were reported, largely due to a decline in catch in the northern areas of the State. Fishing effort similarly declined, particularly on the north coast, though remained high relative to historic levels. Catch rates fell markedly along the north coast and in Great Oyster Bay relative to 2016/17, though were relatively stable for Mercury Passage. Preliminary information from dive surveys suggest low levels of egg-laying activity during the peak spawning season in late 2017, implying low abundance of adults. Catches in 2015/16 and 2016/17 were above estimates of MSY. While there is uncertainty as to the current status of the biomass, recent high catches and effort, particularly off the north coast, suggest that fishing mortality has been excessive and is likely to cause the stock to become recruitment impaired.</p>	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch and CPUE trends



*Sepioteuthis australis*  
Source: DPIPW (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Shallow inshore water.</li> </ul>	Gomon et al. (2008)
Distribution	<ul style="list-style-type: none"> <li>Endemic to southern Australia and northern New Zealand waters.</li> </ul>	Gomon et al. (2008)
Diet	<ul style="list-style-type: none"> <li>Various crustaceans and fishes.</li> </ul>	Norman (2000)
Movement and stock structure	<ul style="list-style-type: none"> <li>Highly mobile. Undergoes migration between feeding grounds and spawning grounds.</li> <li>Preliminary genetic studies showed there are a minimum of 5 genetically distinct stocks in Australia. 98% of the Tasmania population belongs to a single genetic stock which is also found in various proportions in the South Australian, New South Wales and Western Australian populations,</li> </ul>	Triantafillos and Adams (2001) Triantafillos (2004) Smith et al. (2015b)

	<p>suggesting some degree of interbreeding among populations in these areas.</p> <ul style="list-style-type: none"> <li>• A recent and more thorough genetic study showed a single stock across southern Australia with Tasmania being particularly important in terms of reproduction.</li> </ul>	
Natural mortality	<ul style="list-style-type: none"> <li>• High.</li> <li>• Embryo mortality rate between 5% and 25%.</li> </ul>	Steer et al. (2004)
Maximum age	<ul style="list-style-type: none"> <li>• Short-lived (&lt;1 year). Maximum recorded ages are 275 days for males and 263 days for females.</li> </ul>	Pecl et al. (2004) Pecl and Moltschaniwsky 2006
Growth	<ul style="list-style-type: none"> <li>• Rapid growth at 7–8% body weight/day in individuals &lt;100days old, decreasing to 4–5% BW/day in squids older than 200 days.</li> <li>• No gender difference in growth. Growth from 80 days (when recruitment to inshore waters occurs) best described by a power function <math>L = 2e^{-6}t^{3.5332}</math> where <math>L</math> is the mantle length (mm) and <math>t</math> is the age (days).</li> <li>• Extremely variable growth. Some variability may be explained by temperature and food availability (individuals hatched in warmer season/year generally grow faster) but there is also a genetic component.</li> <li>• Length-weight relationship was set at <math>W = 0.00081L^{2.427}</math> where <math>W</math> is weight (g) and <math>L</math> is the dorsal mantel length (mm).</li> </ul>	Pecl et al. (2004) Triantafillos (2004) Data from Pecl (2004)
Maturity	<ul style="list-style-type: none"> <li>• Size-at-50% maturity estimated at 184.5 mm for females.</li> </ul>	Data from Pecl et al. (2006)
Spawning	<ul style="list-style-type: none"> <li>• Major spawning period in spring/summer (September to February) in Tasmania, with low levels of spawning occurring all year round.</li> <li>• Great Oyster Bay (east coast Tasmania) is a known spawning ground. Spawning aggregations are male-biased.</li> <li>• Multiple spawners with individual spawning activity occurring over several months (up to 3.5 months).</li> <li>• Females deposit eggs together in collective egg masses, attaching the finger-like capsules to the substrate by small stalks.</li> </ul>	Moltschaniwskyj and Pecl (2003) Pecl et al. (2004)  Pecl et al. (2006)
Early life history	<ul style="list-style-type: none"> <li>• Incubation time estimated at 4 to 8 weeks depending on water temperature. Hatchlings (2.4-7 mm) swim to the surface and can be found near spawning grounds for 20–30 days.</li> <li>• Habitat and ecology between 20–80 days unknown.</li> <li>• From 80–150 days, juveniles are found in deeper water adjacent to spawning ground.</li> <li>• Individuals become available to the fishery between 90–120 days of age.</li> </ul>	Steer et al. (2002)  Pecl (2000)  Pecl (2004)

## Background

The commercial fishery for Southern Calamari in Tasmania initially developed in the mid-1990s in Great Oyster Bay (GOB) and then expanded rapidly to the southeast (including Mercury Passage, Maria Island and Tasman Peninsula) during the latter half of the 1990s. Annual catches rose from less than 20 t prior to 1995/96 to around 90 t in 1998/99. Since then, catches have fluctuated between 40 and 110 t. The expansion of the fishery was accompanied by a massive increase in effort, particularly squid-jig, which has become the primary capture method in recent years. Calamari are taken in lesser quantities by purse seine, beach seine, spear and dipnet. Although some night fishing occurs, Calamari are mainly targeted during the day over shallow areas of seagrass and macro-algae where they aggregate to spawn.

<b>FISHING METHODS</b>	Squid jig (main), purse seine, beach seine, spear, dipnet.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence).</li> <li>• Species licence (Southern Calamari licence) for the Southeast region.</li> <li>• Class seine licences.</li> <li>• Temporal and spatial closures (mid-October to mid-November) of some east coast waters.</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Daily bag limit of 10 and possession limit of 20 individuals for recreational fishers.</li> <li>• Trip limit of 10 calamari in 24-hour period for SE waters for holders of a fishing licence (personal) not the holder/operator of a calamari, seine or Danish seine licence.</li> <li>• Trip limit of 15 calamari outside SE waters for holders of a fishing licence (rock lobster).</li> </ul>
<b>MAIN MARKET</b>	Local and interstate.

## Current assessment

### **Catch, effort and CPUE**

The total commercial catch of Southern Calamari for 2017/18 was 60.6 t, representing a decrease of 62.0 t from 2016/17 when a record 122.6 t was caught, and the lowest value since 2011/12, when 50.8 t were caught (Fig. 20.1A). Catches declined in all areas of the State, but in particular on the northeast and northwest coasts, where catches have been increasing since 2009/10 and 2010/11, respectively (Fig. 20.1B). Current production levels for these areas (17.3 t and 11.4 t for the northeast and northwest coasts, respectively), are similar to those of the Mercury Passage (14.0 t in 2017/18), but higher than GOB and the southeast coast (3.4 t and 10.2 t in 2017/18; Fig. 20.2B and 20.3). Effort in all areas declined relative to the 2016/17 fishing season, when effort was at an all-time high for the northeast and northwest coasts, and the highest level observed in the last 8 years for the GOB area and Mercury Passage. Preliminary data for 2018/19 indicate catches and catch rates along the north coast continue to be depressed relative to recent years.

The shift in catch and effort to the northeast and northwest coasts has in recent years been accompanied by a general trend of increasing catch rates in both areas (Fig. 20.1D). In 2017/18

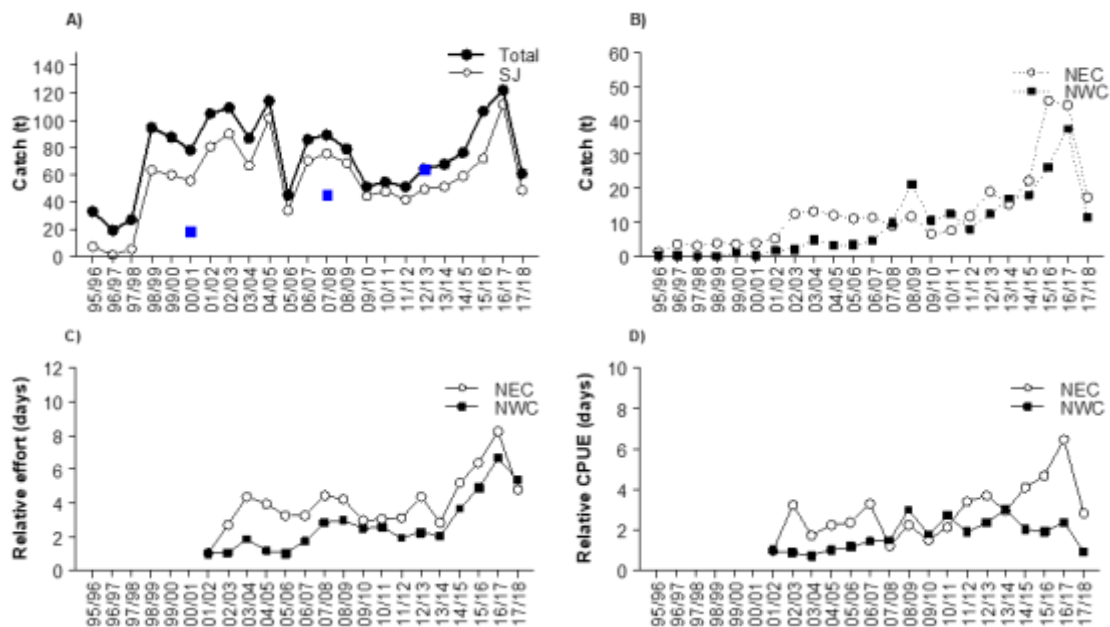
catch rates in both the northeast and northwest coasts declined significantly from those observed in 2016/17, with catch rates in the northwest falling to levels observed in the mid-2000s. Catch rates in the GOB area declined slightly to 2016/17, following a consistent trend observed for the last 3 fishing seasons, while catch rates in Mercury Passage have remained relatively stable since 2010/11 (Fig. 20.2C and 20.2D).

The most recent recreational catch estimate (2012/13) indicated that landings were 63.5 t (Lyle et al. 2014b) and similar to commercial landings. This is a considerable increase from previous years with landings from the recreational fishery growing consistently since 2000 (Fig. 20.1A), indicating high interest by recreational fishers in the species.

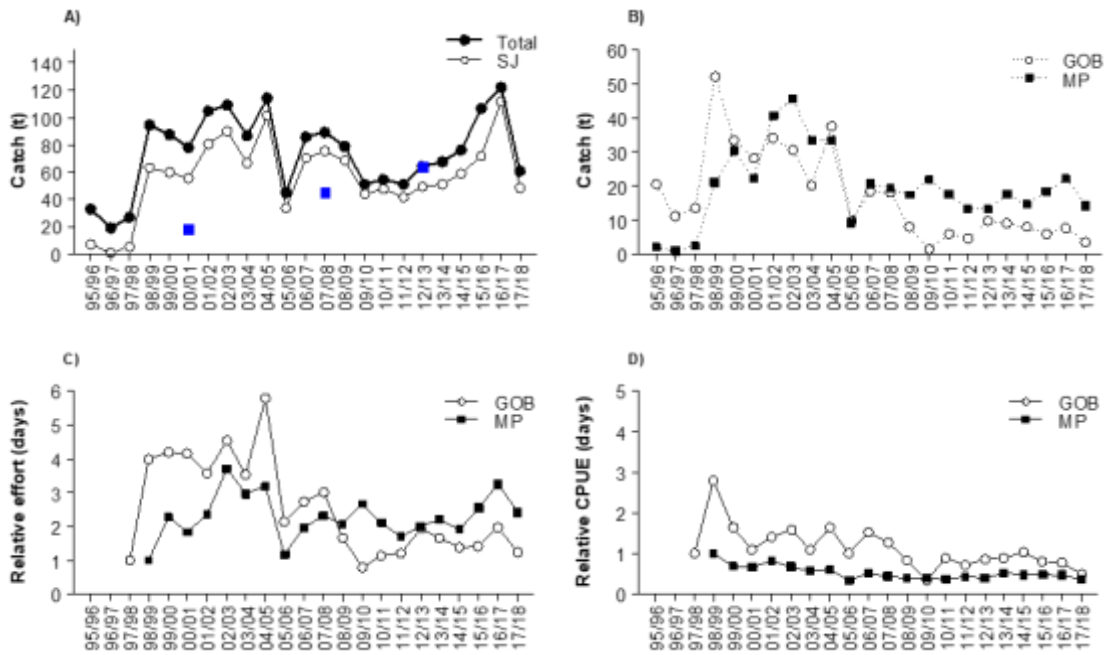
In 2018, the Maximum Sustainable Yield (MSY) of Southern Calamari in Tasmania (State-wide and north coast) was estimated using the catch-only approach of the *simpleSA* R package of Haddon and Punt (2018). The catch-only MSY uses the Schaefer production model to calculate annual biomasses for a given set of *r* and *k* parameters, draw from a uniform prior distribution. Estimates of MSY, derived from commercial catch data only, were 75 t (95% CI = 64–84 t) for all State waters and 33 t (95% CI = 23–48 t) for the north coast.

### Ecological Risk Assessment

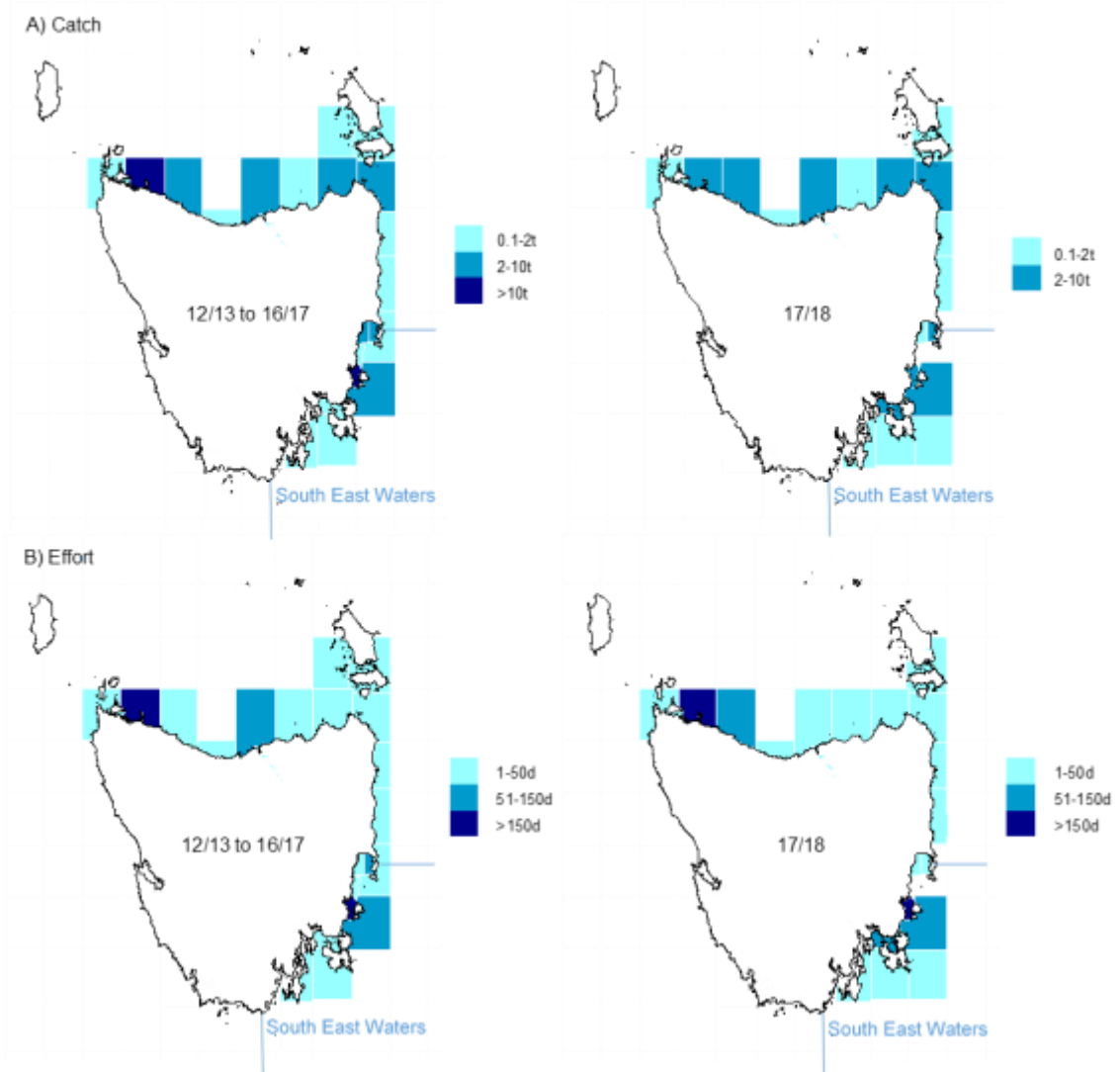
In 2012/13, an environmental risk assessment of the Tasmanian Scalefish Fishery considered the risks to bycatch, threatened endangered and protected species and habitats from fishing for Southern Calamari using squid jigs as negligible. Changes to the ecosystem and community structure from fishing for Southern Calamari using squid jigs was considered a low risk as opposed to negligible, as they are an important predator and their removal could lead to negative effects on the ecosystem and community structure (Bell et al. 2016). As the commercial and recreational catches of Southern Calamari have continued to rise since 2012/13, it is likely that the overall risk has increased in the northern areas of the State.



**Figure 20.1** Annual commercial catch (t) by gear (A) and by region (B), and best estimates of recreational catches (single squares); Commercial squid-jig effort based on days fished relative to relative to 2001/02 for NEC and NWC (C); Commercial squid-jig catch per unit effort (CPUE) based on weight per day (D); SJ=squid jig, NWC=northwest coast, NEC=northeast coast.



**Figure 20.2** Annual commercial catch (t) by gear (A) and by region (B), and best estimates of recreational catches (single squares); Commercial squid-jig effort based on days fished relative to relative to 1998/99 for MP and 1997/98 for GOB (C); Commercial squid-jig catch per unit effort (CPUE) based on weight per day (D); SJ=squid jig, GOB=Great Oyster Bay, MP=Mercury Passage.



**Figure 20.3** (A) Calamari catches (t) and (B) effort (days) for squid jig and purse seine by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right). South East Waters management boundaries are also displayed.

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (105.2 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (33.0 t)	No	
	• Catch variation from the previous year above the greatest inter-annual increase from the reference period (67.9 t)	No	
	• Catch variation from the previous year above the greatest inter-annual decrease from the reference period (-69.6 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (17.7 t)	Yes	↑ 47.6 t (269%)
	• Proportion of recreational catch to total catch > previous proportion estimate (30.3% in 2007/08)	Yes	Latest estimate (2012/13): ↑ 51.3%
<b>Biomass</b>	• CPUE < 3 <sup>rd</sup> lowest CPUE value from the reference period (0.0198 t/days fished)	No	
	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0030)	Yes	↓ 0.0009 t/day fished (29.9%)
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	Not assessed	
	• Significant numbers of unhealthy fish landed	No	

## Stock status

**DEPLETING**

Commercial catches fell markedly relative to the last two fishing years when the commercial catch reference point was breached due to ongoing high landings of Southern Calamari from the northern areas of the State. The recreational reference points were also breached in 2012/13 as a result of growing interest in the species from this sector, with the recreational catch now possibly of a similar magnitude to that of the commercial sector. Thus, overall fishing pressure on Southern Calamari from commercial and potentially recreational fishers has increased substantially since 2012/13. The next state-wide recreational fishing survey is current underway and will provide insight into recent developments within the recreational sector, with data being available in 2019.

Vulnerability of Calamari to fishing pressure is unclear but probably high because spawning aggregations are targeted and the species has an annual or sub-annual life span that renders the stock susceptible to spawning and/or recruitment failure. In light of such susceptibility, spatial and temporal closures have been implemented to reduce fishing pressure on spawning aggregations. The species-specific licence has effectively capped commercial effort in the southeast but fishing effort has increased outside of the South East Waters (defined as waters

between Whale Head to Lemon Rock for calamari management) in the northeast and northwest coasts.

In 2017/18, fishing effort fell in all areas of the State relative to 2016/17, particularly on the north coast. Fishing effort remained high relative to historic levels, particularly along the north coast and Mercury Passage, with anecdotal evidence also suggesting high recreational fishing effort in these regions. Along the north coast, catch rates fell significantly in 2017/18 relative to 2016/17, with catch rates for the northwest coast reaching the lowest level recorded since 2003/04. Catch rates fell slightly in the Great Oyster Bay area, following a trend observed since 2015/16, though were relatively stable for Mercury Passage.

The sharp decline in catch and catch rates for the north coast raises uncertainty about the sustainability of increased effort from both the commercial and recreational sectors on populations in that region. Egg mop surveys conducted in late 2017 provided evidence of limited spawning activity on the main north coast fishing grounds (IMAS unpublished data), implying low abundance of spawning adults, although the role of local environmental factors at that time on spawning behaviour remain unclear. Nevertheless, catches in both 2015/16 and 2016/17 exceeded recent estimates of maximum sustainable yield (MSY) based on catch-only estimation approach of Haddon and Punt (2018) (estimated to be 75 t (95% CI = 64–84 t) State-wide, and 33 t (95% CI = 23–48 t) for the north coast). While there is uncertainty as to the current status of the biomass, recent high catches and effort, particularly off the north coast, suggest that fishing mortality has been excessive and is likely to cause the stock to become recruitment impaired. On this basis, Southern Calamari in Tasmania is classified as a depleting stock.

# 21. Southern Garfish

## *Hyporhamphus melanochir*

<b>STOCK STATUS</b>	<b>DEPLETED</b>
<p>After a strong decline in catches in 2006/07 and 2007/08 coupled with changes in population structure, which prompted management actions, this species showed signs of recovery. However, over the last five fishing seasons there have been significant reductions in total catch and in catch rates, particularly beach seine (the main catching method), with both total catch and beach seine catch rates falling to record low levels. Despite declining catches, catch curve analysis indicates fishing mortality and exploitation rates to be higher in 2017 and 2018 than the late 2000s, when the stock was considered to be in a depleted state. These results suggest current low levels of fishing pressure may be too high to allow stock to recover.</p>	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends; changes in size/age composition



*Hyporhamphus melanochir*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Shallow inshore water (&lt;20m depth) in association with seagrass beds.</li> </ul>	Gomon et al. (2008)
Distribution	<ul style="list-style-type: none"> <li>Eden (New South Wales) to Perth (Western Australia), including Bass Strait and Tasmanian waters. Endemic to Australia.</li> </ul>	Gomon et al. (2008)
Diet	<ul style="list-style-type: none"> <li>Predominantly herbivores (seagrass, algal filaments). Also consume planktonic crustaceans, worms, diatoms and stray insects landing on the surface.</li> </ul>	Edgar (2008) Klumpp and Nichols (1983)
Movement and stock structure	<ul style="list-style-type: none"> <li>Schooling fish, highly mobile.</li> <li>School near the surface at night and close to bottom during day.</li> <li>There are 4 genetically distinct populations distributed in Western Australia, western South Australia, eastern South Australia/Victoria and Tasmania.</li> <li>It is likely that at least two Garfish subpopulations exist in Tasmania, as suggested by the different size and age characteristics exhibited through sampling of the north and east coasts. This has not been confirmed genetically.</li> </ul>	Grant (1991) St Hill (1996)  Jones et al. (2002)

Natural mortality	<ul style="list-style-type: none"> <li>High. Estimated at 55% for adults of four years and over for the east coast population.</li> </ul>	Jones (1990)
Maximum age	<ul style="list-style-type: none"> <li>Up to 9 years.</li> </ul>	Jordan et al. (1998)
Growth	<ul style="list-style-type: none"> <li>From 6 month onwards, growth follows a von Bertalanffy growth function with <math>L_{\infty} = 34.3</math>, <math>k = -0.54</math> and <math>t_0 = 0.23</math></li> <li>Length-weight relationship: <math>W = 0.0011L^{3.4403}</math> where <math>W</math> is weight (g) and <math>L</math> is the fork length (cm).</li> </ul>	Jordan et al. (1998) Hartmann and Lyle (2011)
Maturity	<ul style="list-style-type: none"> <li>Size-at-50% maturity estimated at 19.9 cm for females and 17.1 cm for males.</li> <li>The relationship between batch fecundity and fork length is linear with <math>F = 188.75L - 3585.8</math>, where <math>F</math> is the fecundity (in number of eggs) and <math>L</math> is the fork length (cm).</li> </ul>	Hartmann and Lyle (2011)
Spawning	<ul style="list-style-type: none"> <li>Spawning is concentrated in shallow (&lt;5 m deep) over beds of drift algae in eastern Tasmania. It occurs over an extended period of at least five months from October to February, with peak activity occurring between October and December.</li> <li>Eggs are around 2.93 mm in diameter and are negatively buoyant, sinking to the bottom after fertilisation and becoming attached to drift algae.</li> </ul>	Jordan et al. (1998)
Early life history	<ul style="list-style-type: none"> <li>Incubation time estimated at one month. Large hatchlings (7.8-8.5mm) are assumed to stay in shallow waters as evidenced by the presence of small juvenile (0+ cohort) in the coastal sheltered waters of east Tasmania.</li> </ul>	Jordan et al. (1998)
Recruitment	<ul style="list-style-type: none"> <li>Variable. No-stock recruitment relationship established.</li> </ul>	
Gillnet post release survival	<ul style="list-style-type: none"> <li>NA</li> </ul>	

## **Background**

Traditionally a winter beach seine fishery, catches were initially centred off the northeast coast, including Flinders Island. More recently, the fishery has extended to the east and southeast coasts and with the introduction of dip-nets, catches have increasingly been taken over the summer months. Nowadays, Garfish are caught almost exclusively by beach seine on the northeast coast and mainly by dip-nets off the southeast and east coasts.

## **Current assessment**

### **Biological sampling**

The sharp and unexpected decline in catches in 2006/07 and 2007/08 prompted the need to better understand stock status. Dedicated sampling took place between 2008–2012, and again between 2017 and 2018 as part of a student research project at the University of Tasmania (Reid 2018). These data have been reanalysed for this assessment.

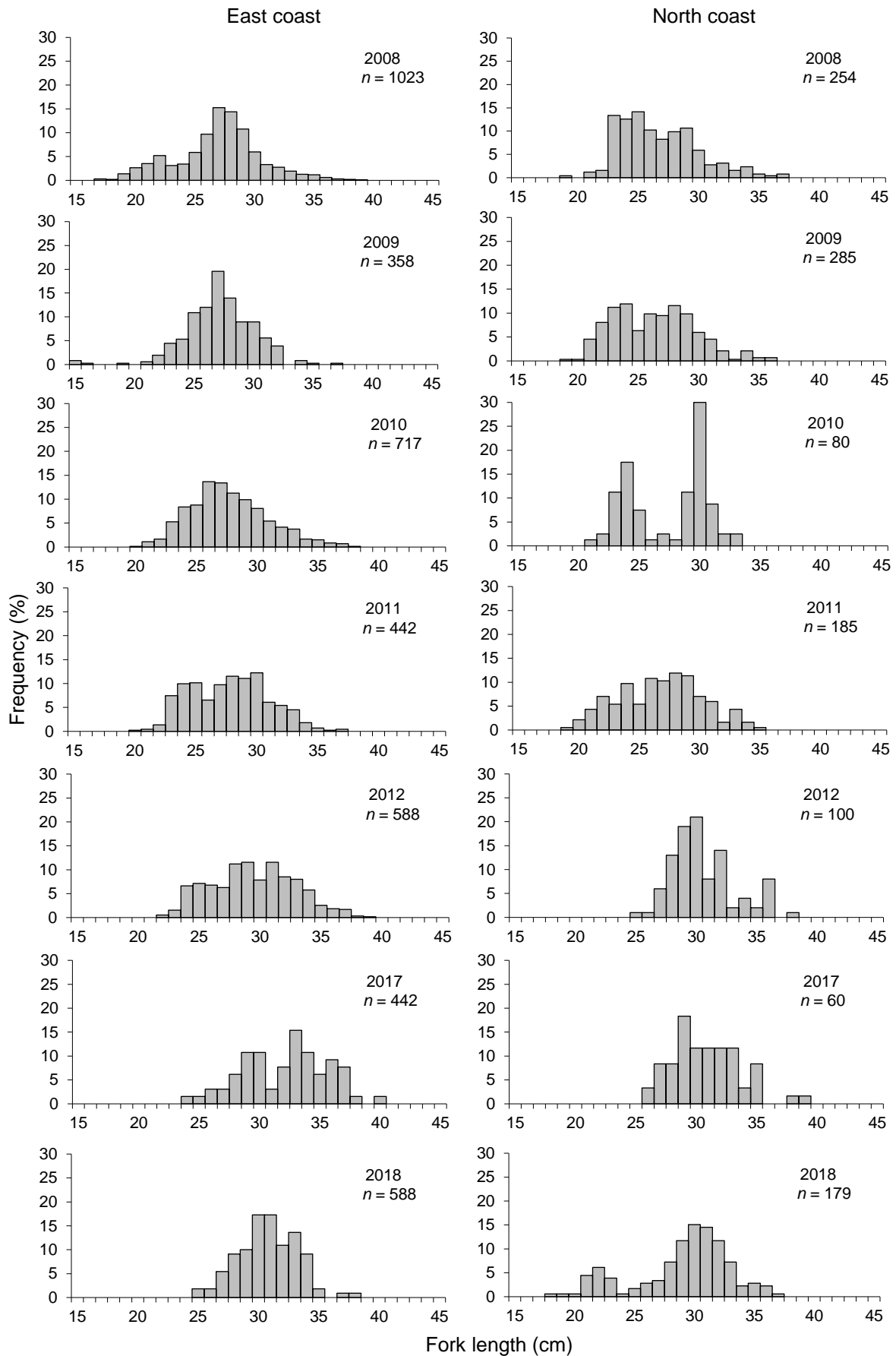
<b>FISHING METHODS</b>	Mainly dip-net and beach seine.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Temporal closure during spawning: mid-Nov to mid-Dec for southern waters, mid-Jan to mid-Feb for northern waters (Northern and southern waters are delineated by a line following the north coast of Tasmania, joined and bounded in the west by a line of latitude through Cape Grim and in the east by a line of latitude through Cape Naturaliste).</li> <li>• Gear restriction (Scalefish fishing licence, Purse seine licence, Beach seine licence).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Legal size: 25 cm (upper jaw to end of tail).</li> <li>• Possession limit of 30 and bag limit of 15 individuals for recreational fishers.</li> </ul>
<b>MAIN MARKET</b>	<ul style="list-style-type: none"> <li>• Local and interstate.</li> </ul>

### **Size composition**

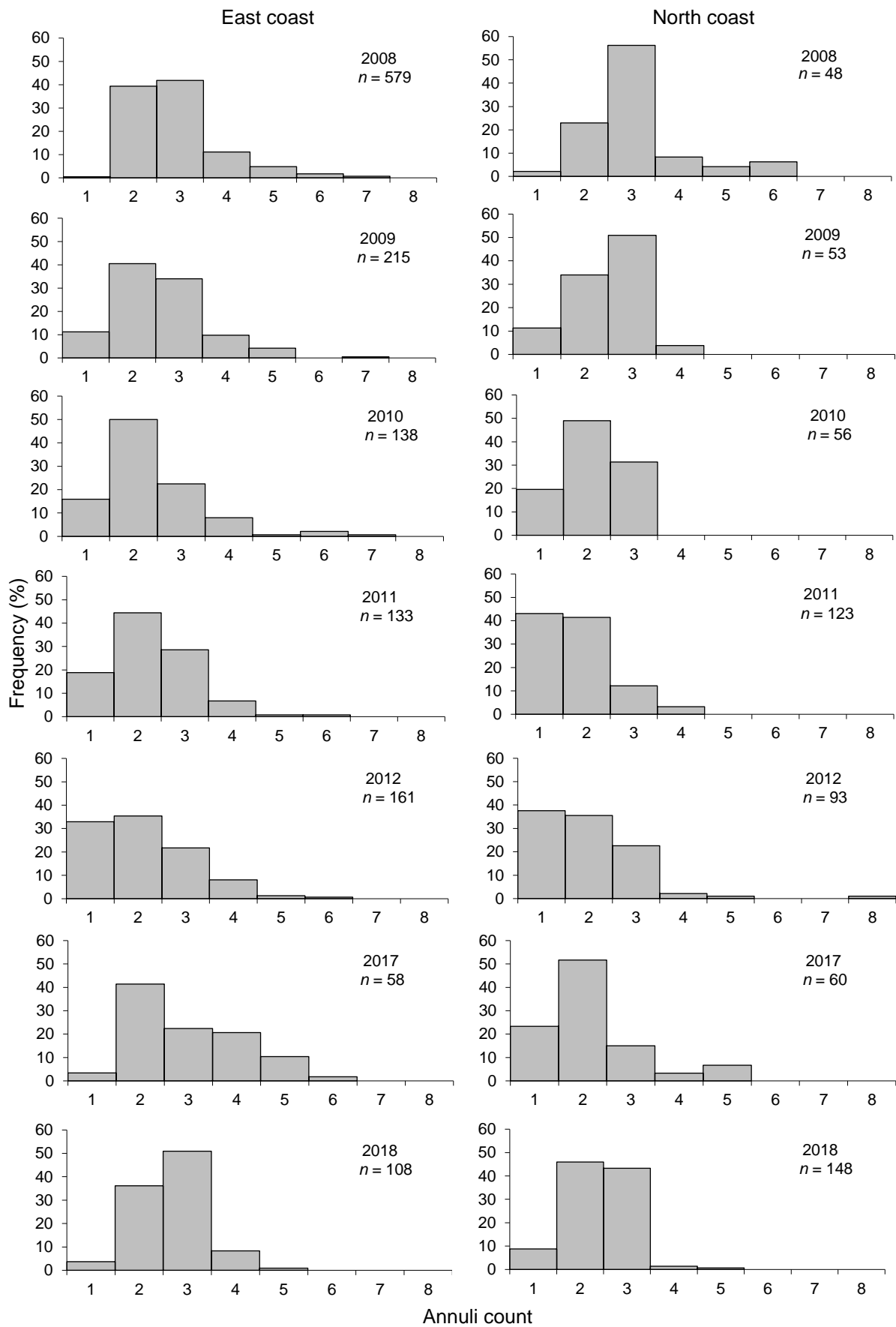
Annual length frequency distributions of east coast catches tended to be unimodal with a peak between 27–28 cm during the late 2000s, falling from around 30 cm in the 1990s (Emery et al. 2015). These reductions were associated with a sharp decline in commercial catches and may be indicative of a more general change in population structure. Since 2012 however, there appeared to be a recovery with average size in eastern Tasmania moving towards 30 cm, which Emery et al. (2015) considered as evidence for population recovery. However, Reid (2018) showed that increases were largely due to increased size-at-age in these later years, and not related to changes in age structure. A similar pattern was apparent on the north coast. During the mid-1990s individuals between 35 and 40 cm were relatively common in the catch with modal peaks >30 cm but in subsequent years fish larger than about 35 cm were rare and modal peaks were generally <30 cm (Fig. 20.3). Since 2012, there has been an increase in the average size of Garfish, and in the catch of larger fish (>35 cm).

### **Annuli counts (as proxy for age composition)**

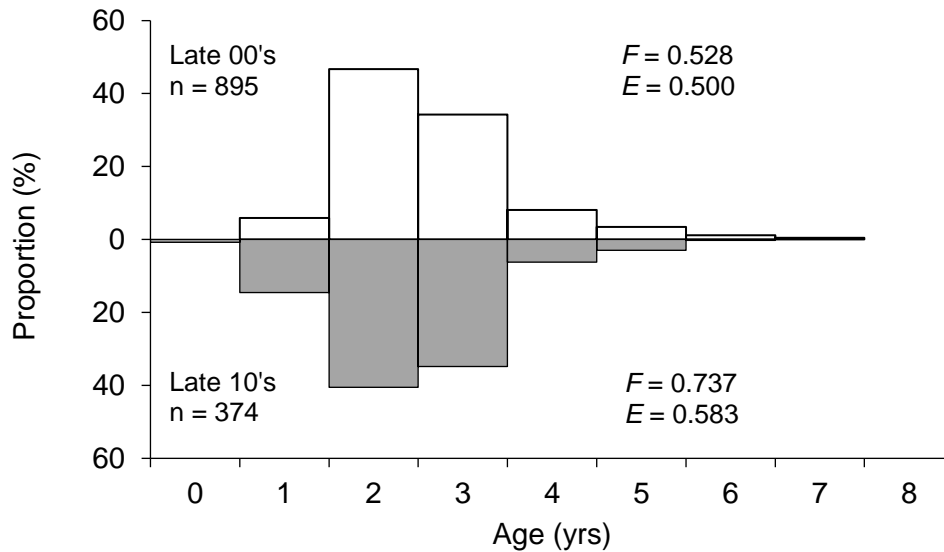
As noted by Reid (2018), precise ages were not able to be calculated for Garfish collected prior to 2017 due to a suspected confusion between either the date of capture and/or edge classifications not conforming to the established biological trend of increment formation within the species for these samples. Accordingly, comparisons of age structures and mortality rates between collection periods were performed using the raw count of opaque zones (hereafter referred to as annuli). The age structure of east and north coast catches taken between 2008 and 2009 were largely similar, and both were dominated by fish with 2–4 annuli, with a small proportion of fish with 6 or more annuli (Fig. 21.2). In 2010–2011 however, the age distribution changed in both regions, with an apparent truncation in age structure (particularly on the north coast) fish with a single annuli becoming more prevalent in catches (Fig 21.2). In 2012, fish with 3 or more annuli were more prevalent, which Emery et al. (2015) interpreted as the start of a recovery. However, in 2017 and 2018 catches on both coasts were dominated by fish with 2–3 annuli, with few fish with 4 or more annuli. Estimates of total mortality ( $Z$ ) based on catch curve analysis were higher for samples collected in the late 2010s (2017–2018;  $Z=1.264$ ) compared to those collected in the late 2000s ( $Z=1.055$ ), inferring that fishing mortality ( $F$ ) and exploitation rate ( $E$ ) was higher in 2017 and 2018 than in the late 2000s (Fig 21.3). Assuming samples are representative, and annuli counts indicative of age, this suggests that despite declines in catch and effort, the population is subject to greater relative fishing pressure than previously estimated.



**Figure 21.1** Garfish length composition 2008–2018. *n* = sample size. Note the 2018 sample included several undersized fish for growth studies of Reid (2018).



**Figure 21.2** Garfish annuli counts 2008–2018. *n* = sample size. Note undersized fish (i.e. those < 25 cm) have been excluded.



**Figure 21.3** Age class frequency plots for Garfish collected in the late 2000s (2008–2009; top) and late 2010s (2017–2018; bottom) (samples from all regions and fishing methods combined). Note undersized fish (i.e. those < 25 cm) have been excluded from this analysis.

### Catch, effort and CPUE

The total commercial catch of Garfish for 2017/18 was 8.9 t, the lowest catch on record and representing the continuation of a general decline in landings that commenced in 2009/10 (Fig. 21.4A). After many years of relative stability in Garfish catches at 80–90 t per annum, catches fell sharply in 2006/07 and 2007/08 before recovering to around 60 t before the current general decline. Catches, both historically and in 2017/18, were concentrated off the northeast coast (Fig. 21.4A and Fig. 21.5A). Historically, there were also moderate landings in the east and southeast, however these have been greatly reduced in recent years (Fig. 21.5A).

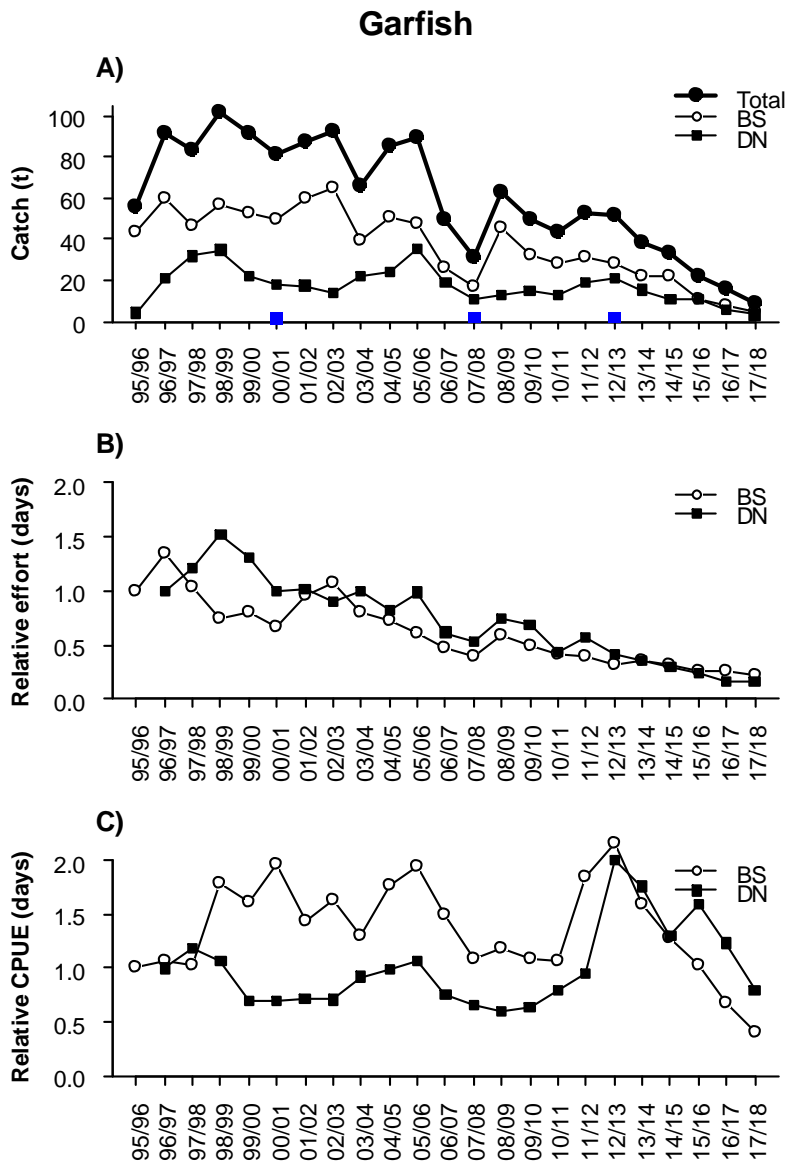
Recreational Garfish catches are low compared to commercial catches and were estimated around 2 t in 2000/01, 2007/08 and 2012/13 (Henry and Lyle 2003, Lyle et al. 2009, Lyle et al. 2014a).

Effort of both major commercial gear types has been slowly decreasing through time, reaching historic lows in the current year (Fig. 21.4B).

Catch rates have fluctuated much more strongly over time for beach seine but did peak for both methods in 2012/13 (Fig. 21.4C). Beach seine catch rates have declined steadily since then such that by 2017/18 they were at a historical low, and were well below those during the late 2000s, when there were strong concerns about the status of the Southern Garfish stocks. Dip-net catch rates also declined after 2012/13 but have remained generally around the long-term average for the method.

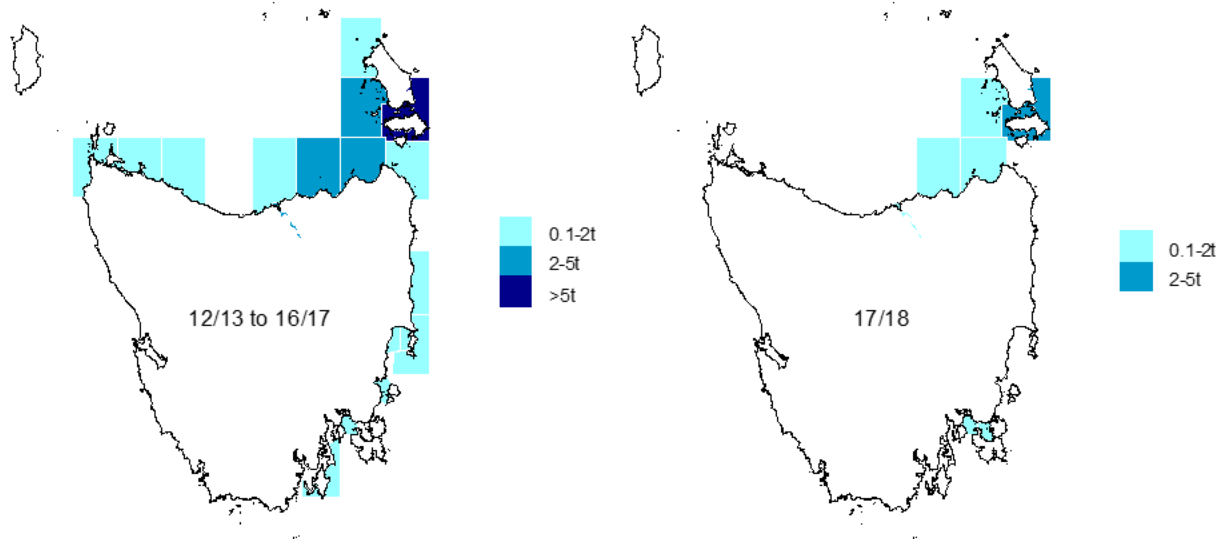
### Ecological Risk Assessment

In the 2012/13 ERA of the Tasmanian scalefish fishery, beach seining was considered a low risk activity with regard to Southern Garfish due to low catches and signs of population recovery at the time. Beach seining was also considered a low risk activity with regards to non-retained species as bycatch is usually released alive and “herded” not “meshed/gilled” and a very low risk in regards to the general ecosystem (Bell et al. 2016).

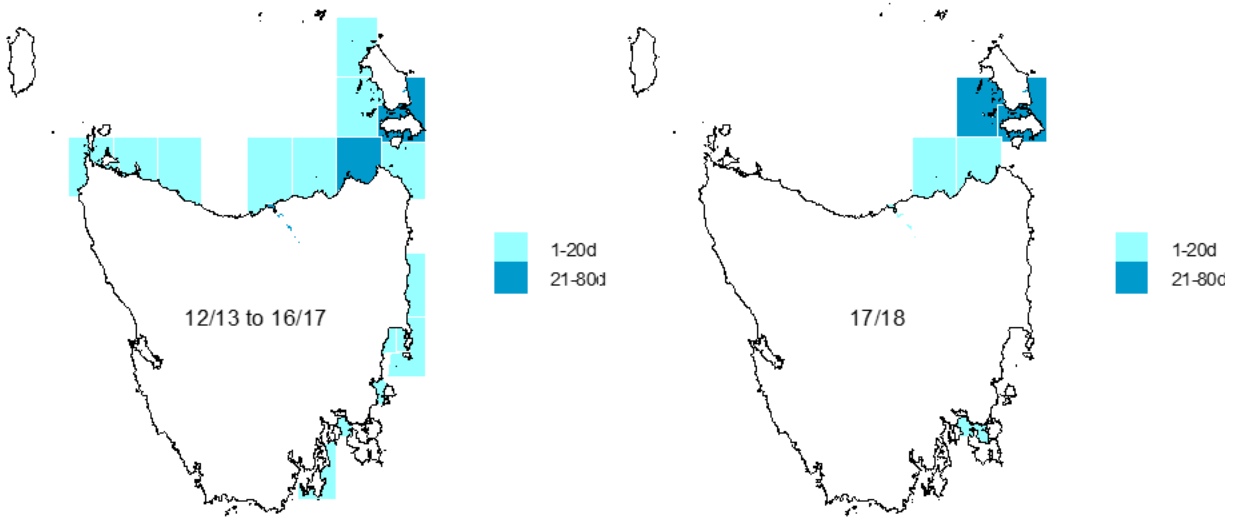


**Figure 21.4** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. BS=beach seine, DN=dip-net.

A) Catch



B) Effort



**Figure 21.5** (A) Garfish catches (t) and (B) effort (days) for beach seine and dipnet by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (91.7 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (66.2 t)	Yes	↓ 57.4 t (86.7%)
	• Catch variation from the previous year above the greatest inter-annual increase from the reference period (35.5 t)	No	
	• Catch variation from the previous year above the greatest inter-annual decrease from the reference period (-39.4 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (1.9 t)	Yes	0.1 t ↑ (5.3%)
	• Proportion of recreational catch to total catch > previous proportion estimate (3.7% in 2007/08)	Yes	Latest estimate (2012/13): ↑ 3.8%
<b>Biomass</b>	• CPUE < 3 <sup>rd</sup> lowest CPUE value from the reference period (0.05 t/days fished)	Yes	0.0288 t ↓ (57.2%)
	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0073)	Yes	0.0094 t ↓ (128.7%)
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	No	
	• Significant numbers of unhealthy fish landed	No	

### Stock status

**DEPLETED**

Catch and fishing mortality reference points were breached due to the low 2017/18 catch relative to levels prior to 2006/07. Catch rates over the last four fishing seasons have also declined, coupled with reductions in beach seine and dip net effort. There has also been a notable reduction in the catch from the east and southeast of the state. While the recreational reference points were also breached, catches have remained essentially unchanged among survey years and are insignificant when compared to the commercial catch.

Vulnerability of Southern Garfish to fishing pressure is unclear but is probably moderate to high due to the schooling behaviour of the species, which means they can be efficiently targeted under favourable conditions. As a relatively short-lived species, and with a population dominated by relatively few age classes, any recruitment variability is likely to have a marked impact on population size.

Up to 2011 there was evidence of a reduction in average fish size and a truncation of the age structure that was presumed indicative of a more general change in population structure resulting from heavy fishing pressure and/or poor recruitment. While there were early signs of population

recovery in 2012 (increasing size and age), which may have been linked in part to the implementation of spawning closures in 2009, subsequent declines in catches and catch rates suggest that any stock recovery was short-lived. Furthermore, fishing mortality and exploitation rates are currently estimated to exceed those observed for the late 2000s when the stock was in a depleted state, implying that stock biomass has returned to low levels. Based on the available evidence of historically low catches and catch rates coupled with high fishing mortality and exploitation rates the stock is, therefore, classified as depleted.

## 22. Striped Trumpeter

### *Latris lineata*

<b>STOCK STATUS</b>	<b>RECOVERING</b>
After a lack of recruitment for over a decade, and limited biological information from which to assess stock status in recent years, Striped Trumpeter is showing signs of recovery in the last two years, with evidence for recent recruitment.	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery, Commonwealth fisheries
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Latris lineata*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li>Exposed reefs and rocky bottom down to 300 m depth.</li> </ul>	Edgar et al. (2004) Gomon et al. (2008)
Distribution	<ul style="list-style-type: none"> <li>Sydney (New South Wales) to Albany (Western Australia), Tasmania, New Zealand, Amsterdam Islands (southern Indian Ocean) and most of temperate Southern hemisphere (excl. South Africa and South America).</li> </ul>	Edgar et al. (2004) Gomon et al. (2008)
Diet	<ul style="list-style-type: none"> <li>Small fish, cephalopods, crustaceans.</li> </ul>	Nichols et al. (1994)
Movement and stock structure	<ul style="list-style-type: none"> <li>Juveniles have limited movement, remaining around shallow reefs for several years before moving into deeper offshore reefs.</li> <li>Adults have the capacity to undergo wide-scale movements (e.g. Tasmania to St Paul Island in Indian Ocean).</li> <li>Uniform stock structure in Tasmanian waters (no significant genetic separation of populations).</li> </ul>	Tracey and Lyle (2005) Lyle and Jordan (1999) Lyle and Murphy (2001) Tracey et al. (2007a)
Natural mortality	<ul style="list-style-type: none"> <li>Estimated at <math>M = 0.096</math>.</li> </ul>	Tracey and Lyle (2005)

Maximum age	<ul style="list-style-type: none"> <li>Estimated at 43 years.</li> </ul>	Tracey and Lyle (2005)																		
Growth	<ul style="list-style-type: none"> <li>Maximum length: 1.2 m</li> <li>Maximum weight: 25 kg</li> <li>Rapid juvenile growth (mean FL = 28 cm after 2 years, 42 cm after 4 years). Slower adult growth (large range of size-at-age over 50 cm FL).</li> <li>Growth for both sex described by a two-phase von Bertalanffy growth function</li> </ul> $L = \left( 1 - \int_{t=-t_0}^{+t^\delta} \frac{1}{\sigma\sqrt{2\pi}} e^{\left(\frac{-(t-t_{01})^2}{2\sigma^2}\right)} \right) (L_{\infty 1}(1 - e^{-k_1(t-t_{01})}) + \varepsilon) + \left( 1 - \int_{t=t^\delta}^{+t_{max}} \frac{1}{\sigma\sqrt{2\pi}} e^{\left(\frac{-(t-t^\delta)^2}{2\sigma^2}\right)} \right) (L^\delta + (L_{\infty 2} - L^\delta)(1 - e^{-k_2(t-t_{02})}) + \varepsilon)$ <p>where <math>L</math> is the length (mm), <math>t</math> is the age (years), <math>L_{\infty 1}</math> and <math>L_{\infty 2}</math> are the average maximum length for the species for the 1<sup>st</sup> and 2<sup>nd</sup> growth phase respectively, <math>k_1</math> and <math>k_2</math> are constants, <math>t_{01}</math> and <math>t_{02}</math> are the (theoretical) age where length equals zero for the 1<sup>st</sup> and 2<sup>nd</sup> growth phase respectively, <math>L^\delta</math> and <math>t^\delta</math> are the length and age of transference from one growth phase to the next, <math>t_{max}</math> is the maximum age present in the sample, <math>\sigma^2</math> is the standard deviation of cumulative density function with mean <math>t^\delta</math>, and <math>\varepsilon</math> is an error term.</p> <p>Parameter estimates are:</p> <table border="1"> <thead> <tr> <th><math>L_{\infty 1}</math></th> <th><math>k_1</math></th> <th><math>t_{01}</math></th> <th><math>L^\delta</math></th> <th><math>L_{\infty 2}</math></th> <th><math>k_2</math></th> <th><math>t_{02}</math></th> <th><math>t^\delta</math></th> <th><math>\sigma^2</math></th> </tr> </thead> <tbody> <tr> <td>532.77</td> <td>0.43</td> <td>0.03</td> <td>450.1</td> <td>871.59</td> <td>0.08</td> <td>3.49</td> <td>4.4</td> <td>1.0</td> </tr> </tbody> </table> <ul style="list-style-type: none"> <li>Length-weight relationship for both sex was estimated at <math>W = 2E^{-5}L^{3.00}</math>, where <math>W</math> is weight (g) and <math>L</math> is the fork length (mm).</li> </ul>	$L_{\infty 1}$	$k_1$	$t_{01}$	$L^\delta$	$L_{\infty 2}$	$k_2$	$t_{02}$	$t^\delta$	$\sigma^2$	532.77	0.43	0.03	450.1	871.59	0.08	3.49	4.4	1.0	<p>Gomon et al. (2008)</p> <p>Murphy and Lyle (1999)</p> <p>Tracey and Lyle (2005)</p>
$L_{\infty 1}$	$k_1$	$t_{01}$	$L^\delta$	$L_{\infty 2}$	$k_2$	$t_{02}$	$t^\delta$	$\sigma^2$												
532.77	0.43	0.03	450.1	871.59	0.08	3.49	4.4	1.0												
Maturity	<ul style="list-style-type: none"> <li>Size-at-50% maturity estimated at 54 cm FL (62 cm TL; 6.8 years) for females and 53 cm FL (61 cm TL; 6.2 years) for males.</li> <li>Batch fecundity (BF) estimate: <math>BF = 4.15E^{-8}FL^{4.69}</math>, where <math>FL</math> is fork length in cm.</li> </ul>	<p>Tracey et al. (2007b)</p> <p>IMAS unpublished data</p>																		
Spawning	<ul style="list-style-type: none"> <li>July to early October depending on geographic location (early start and finish at lower latitudes).</li> <li>Multiple spawners, highly fecund (100,000 to 400,000 eggs for females weighing 3.2 kg and 5.2 kg respectively).</li> <li>Small pelagic eggs (1.3 mm diameter).</li> </ul>	<p>Ruwald et al. (1991)</p> <p>Ruwald, 1992 (1992)</p> <p>Hutchinson (1993)</p>																		
Early life history	<ul style="list-style-type: none"> <li>Complex and extended larval phase of at least 9 months before settlement.</li> <li>No information on size and timing of settlement. Juveniles of around 18 cm FL (23 cm TL) have been caught on shallow reefs off southeast coast in January.</li> </ul>	<p>Ruwald et al. (1991)</p> <p>Ruwald (1992)</p> <p>Murphy and Lyle (1999)</p>																		

Recruitment	<ul style="list-style-type: none"> <li>Highly variable. No-stock recruitment relationship established.</li> </ul>	Murphy and Lyle (1999)
Gillnet post release survival	<ul style="list-style-type: none"> <li>NA</li> </ul>	

## Background

Striped Trumpeter has a long history of commercial exploitation in Tasmania, being highly valued for its eating qualities. There is also a high level of interest in the species from recreational fishers and charter boat operators. The species is taken by a variety of fishing gears, with hand-line and drop-line the primary methods. Juvenile Striped Trumpeter are occasionally taken in gillnets in inshore waters and usually in depths <50 m, whereas adult fish are taken in deeper offshore waters by line methods and as by-product in large mesh gillnets (shark nets). Historically, catches have been concentrated off the east coast, including Flinders Island, as well as off the south and southwest coasts of Tasmania (André et al. 2015).

Responsibility for the management of Striped Trumpeter was passed to Tasmania in 1996 through an Offshore Constitutional Settlement (OCS) arrangement with the Commonwealth. A memorandum of understanding accompanied the OCS, specifying trip limits for Commonwealth only fishers. As part of the Tasmanian scalefish management plan, gear restrictions for all commercial scalefish fishers operating in State waters were introduced in 1998. This, however, enabled dual licensed operators (i.e. holders of a Tasmanian licence and a Commonwealth permit for southern shark or South East Non-Trawl fisheries) as well as rock lobster fishers to take unrestricted quantities of Striped Trumpeter in offshore waters using their gear allocations. In 2000, the State Government introduced a combined trip limit of 250 kg for Striped Trumpeter, yellowtail kingfish and snapper for all fishers (Commonwealth and State) in inshore and offshore waters to limit the potential for expansion of effort directed at these species. Over time, there have been additional management measures targeted at the species, including a spawning closure, a decrease in the recreational possession limit, introduction of a recreational boat limit and several increases in the minimum size limit for the species (currently 55 cm total length (TL), which is still below the size at maturity of 62 cm TL for females and 61 cm TL for males).

Although the seasonal closure or minimum size limit does not restrict fishing by Commonwealth vessels, in 2013 the Commonwealth reduced their Striped Trumpeter trip limit component to 150 kg (it is still a part of the 250 kg combined species trip limit, but only a maximum of 150 kg can comprise Striped Trumpeter) year round.

<b>FISHING METHODS</b>	Mainly hand-line, also gillnet and drop-line.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>Gear licence (Scalefish fishing licence, rock lobster licence).</li> <li>Temporal closure (Sept–Oct – does not apply to Commonwealth operators).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>Trip limit of 250 kg.</li> <li>Trip limit of 150 kg for Commonwealth operators.</li> <li>Bag limit of 4 fish for recreational fishers.</li> <li>Possession limit of 8 fish for recreational fishers.</li> <li>Boat limit of 20 fish for recreational vessels.</li> <li>Minimum size (550 mm TL).</li> </ul>
<b>MAIN MARKET</b>	Mainly local.

## **Current assessment**

### **Biological characteristics**

#### ***Length frequency composition***

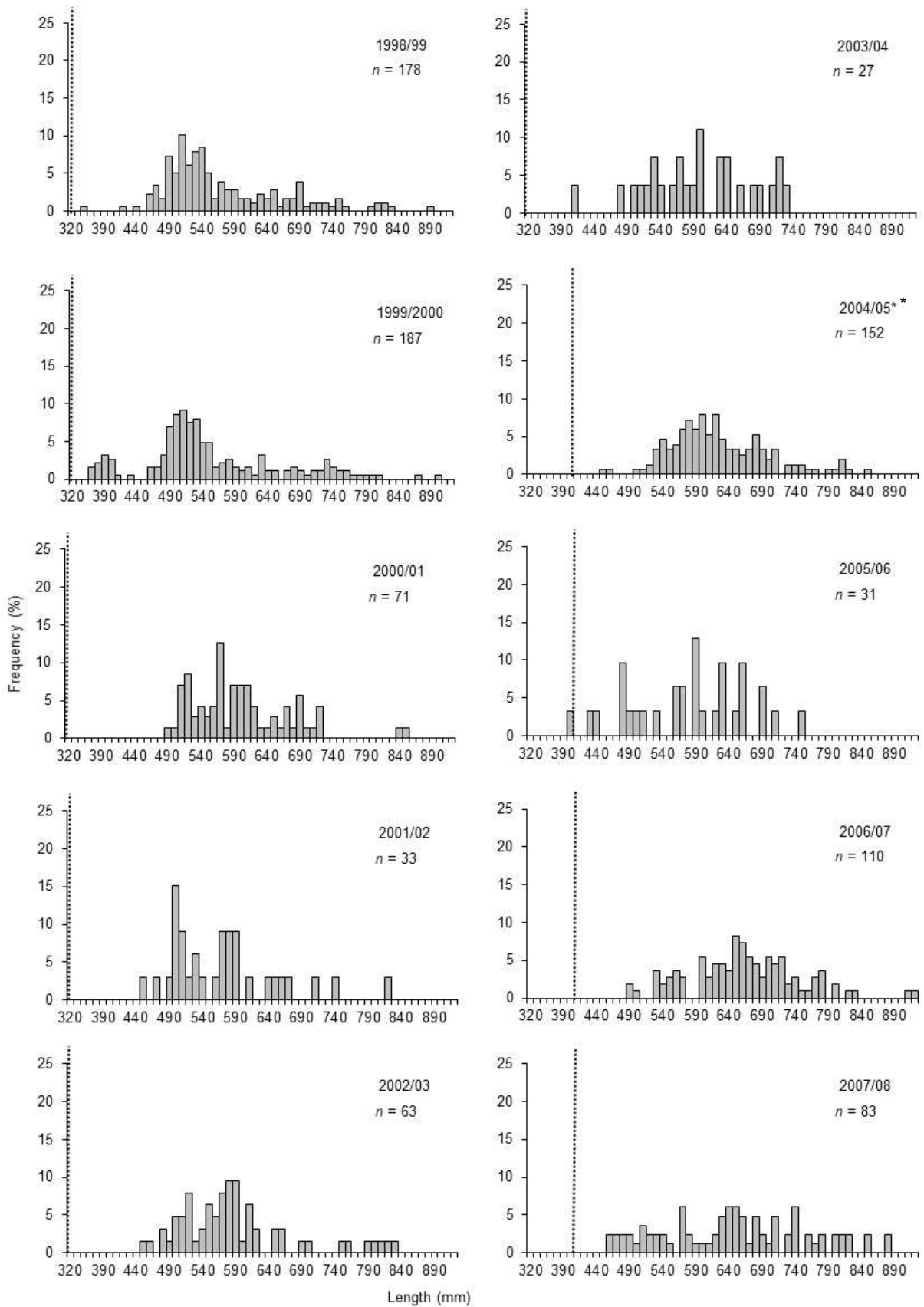
In 1999, the length distribution was unimodal and dominated by 50 cm fork length (FL) fish, with a tail of larger fish (Fig. 22.1). Many of the subsequent sample sizes were low and sampling has been opportunistic, which may not represent population length structure. Years with good sample sizes, however, suggest that no significant recruitment occurred for about a decade during the late 1990s and early 2000s following a major recruitment event in 1993. This 1993 cohort still dominated the catches in 2009 and at this time, samples were dominated by fish around 62 cm in length. In 2010, the length distribution changed and was again dominated by 50 cm fish, suggesting a new wave of recruits had entered the fishery (Fig. 22.1). The length structure appeared stable between 2010 and 2012 suggesting recruitment over at least 3 consecutive years. There was a lack of samples between 2012/13 and 2015/16, precluding a definitive assessment in these years. Relatively large sample sizes were available in 2016/17 and 2017/18, with length frequencies over these years showing a good spread of length classes, dominated by 54 cm to 59 cm individuals in 2016/17 and 64 cm individuals in 2017/18 (Fig. 22.1).

#### ***Age frequency composition***

The very strong 1993 year class (5 year olds in 1998/99, 6 year olds in 1999/2000, etc.) was prominent in samples obtained from research fishing and commercial catches up until 2008/09 (Fig. 22.2). As with length data, the low sample sizes in many years may not represent population age structure. Notwithstanding these concerns, the data demonstrated a consistent lack of recent recruits (3–5 year olds) up until 2010 which, along with catch declines, implied an extended period of low recruitment. During this time the population appears to have been sustained largely by strong year classes spawned during the 1990s.

Samples from 2009/10 to 2015/16 were dominated by 4–6 year olds (the age at which the species tends to recruit to the offshore hook fishery), providing evidence of recent recruitment success. While the relative strength of these cohorts is unknown, they provide a positive sign for the future of the stock, although caution is recommended since the number of individuals sampled between 2012/13 and 2015/16 was low and previous indicators have suggested that the adult segment of the population may still be in a depleted state (due to fishing and lack of recruitment over many years). Few individuals older than 10 years old were sampled in these years but again this is likely a factor of low sample sizes rather than a major contraction in the population age structure. Age composition data for 2016/17 and 2017/18, when relatively large sample sizes were obtained, showed evidence for recent recruitment, with a number of fish aged 3–6 years observed, and an increasing proportion of fish aged between 7 and 13 years, along with several fish > 20 years old (Fig. 22.2).

Using the GLMM estimator of Millar (2015) applied to the 2017/18 samples, total mortality ( $Z$ ) was estimated as 0.15. Assuming a maximum age of 43 years, natural mortality ( $M$ ) was estimated as 0.096 (using the equation of Hoenig 1983; Tracey and Lyle 2005). According, fishing mortality ( $F$ ) based on 2018 samples was estimated as 0.054 (i.e. around half that of  $M$ ).



**Figure 22.1** Striped Trumpeter length composition (financial year 1998/99–2017/18) from commercial and recreational catches.  $n$  = sample size. Dotted line indicates the minimum legal size limit (in FL) and \* indicate years when the minimum size limit was changed. Length is fork length.

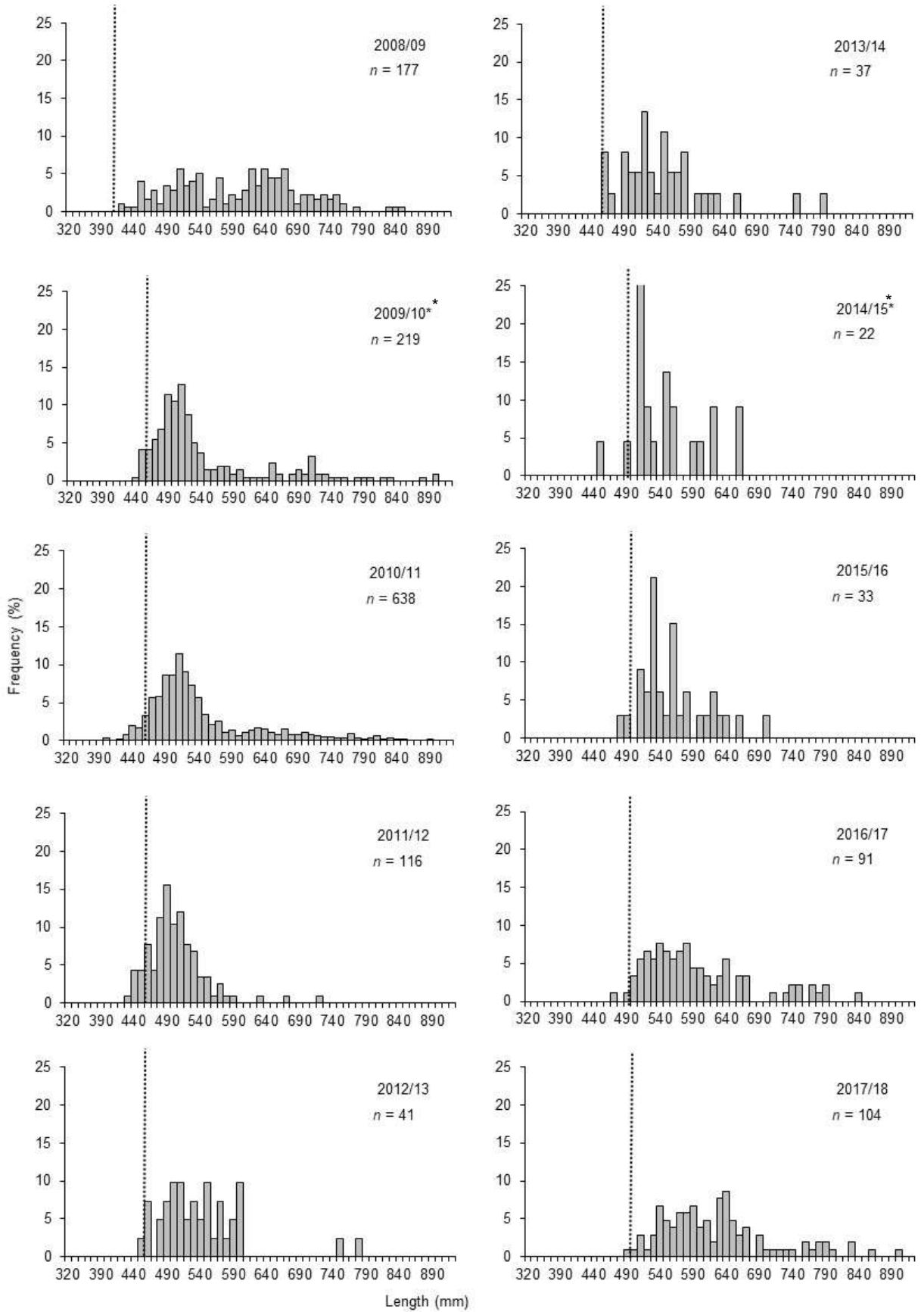
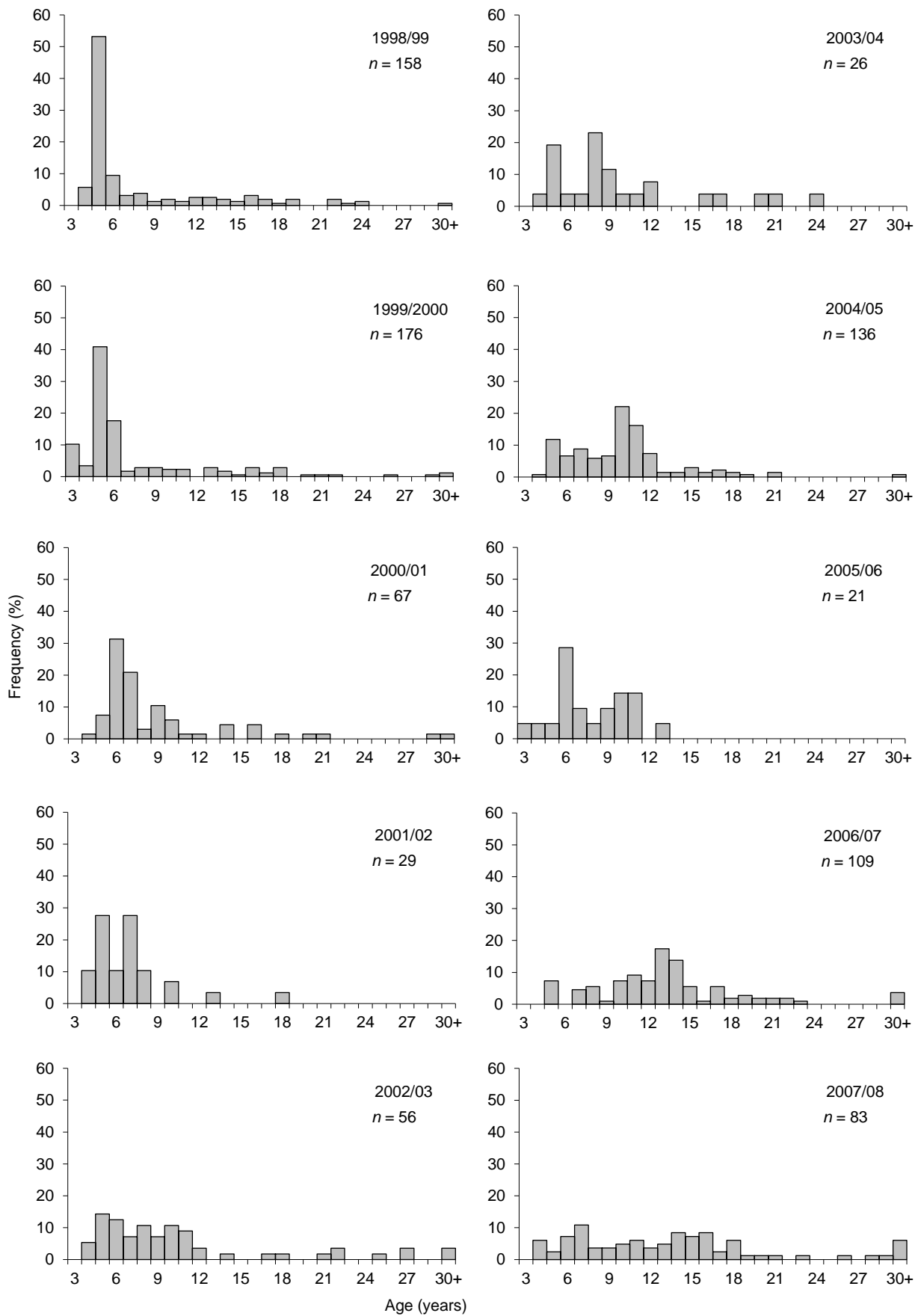


Figure 22.1 – continued



**Figure 22.2** Striped Trumpeter age composition (financial year 1998/99–2017/18) from commercial and recreational catches. *n* = sample size (years 2008/09 to 2016/17 on following page).

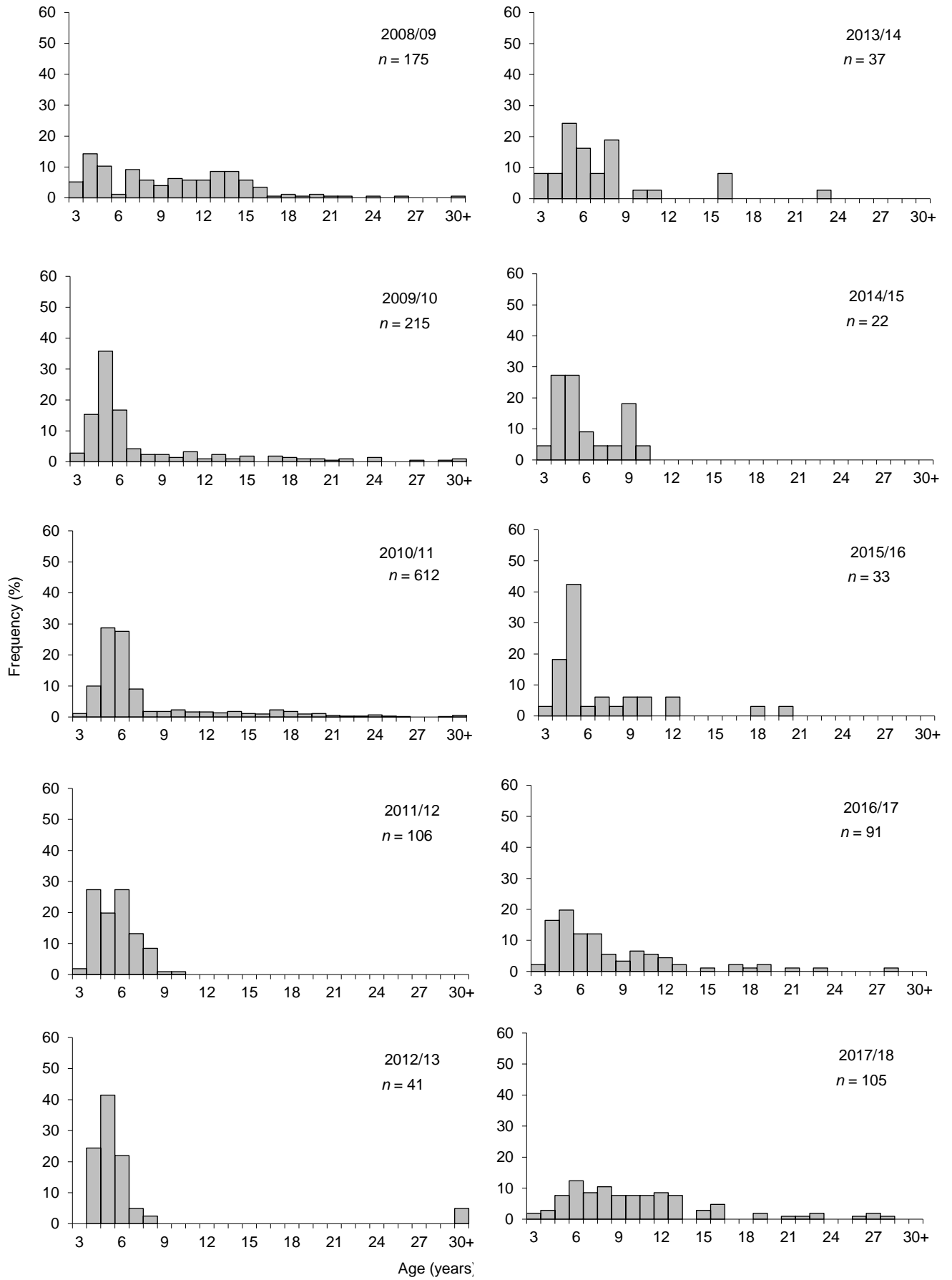


Figure 22.2 (continued)

## Catch, effort and CPUE

The recent catch history in waters south of latitude 39° 12'S (i.e. waters incorporated within the OCS agreement for Striped Trumpeter), including catches reported in the Victorian and Commonwealth logbooks, is presented in Table 22.1. In the early 1990s catches by Victorian vessels were significant, peaking at around 37 t. Since the mid-1990s, data from this sector have been unavailable, though it is assumed that subsequent catches have been reported in Commonwealth logbooks. Other than 1999/2000 when >14 t was taken, Commonwealth catches have been relatively low (generally < 5 t) (Table 22.1).

Annual production was high at over 110 t in the early 1990s, with Victorian vessels accounting for 17–39% of the total, but then fluctuated between 70–80 t through the mid-1990s before increasing again to over 100 t by the late 1990s (Table 22.1). Catches almost halved in 2000/01 to less than 50 t and have remained low since that time. This trend was observed by all methods in Tasmania (Fig. 22.1A). In 2015/16 the total catch fell to a historical low of 7.1 t caused by decreased catches reported in both State and Commonwealth logbooks. In 2017/18 catches increased to 14.1 t (Table 22.1).

**Table 22.1** Annual commercial catches of Striped Trumpeter (t) south of latitude 39° 12'S. Data based on Tasmanian (General Fishing return), Victorian and Commonwealth catch returns.

Year	Catch (t)			Combined
	Tasmania	Victoria	Commonwealth	
1990/91	74.5	37.1		111.6
1991/92	58.2	36.8		95.0
1992/93	52.7	19.8		72.5
1993/94	56.5	16.0		72.5
1994/95	72.4	14.6		87.0
1995/96	60.3			60.3
1996/97	79.7		0.7	80.4
1997/98	75.4		5.7	81.1
1998/99	98.4		8.9	107.4
1999/2000	86.3		14.5	101.8
2000/01	41.2		7.5	49.6
2001/02	40.0		4.8	44.9
2002/03	36.8		3.2	40.0
2003/04	36.8		3.7	40.5
2004/05	24.0		2.2	26.2
2005/06	19.1		4.7	23.8
2006/07	18.8		3.5	22.3
2007/08	13.1		3.0	16.1
2008/09	10.5		2.8	13.3
2009/10	10.0		2.3	12.3
2010/11	15.0		4.8	19.8
2011/12	15.9		5.4	21.3
2012/13	12.3		5.1	17.4
2013/14	8.0		2.5	10.5
2014/15	9.6		3.4	13.0
2015/16	6.0		1.1	7.1
2016/17	8.3		4.0	12.3
2017/18	7.8		6.3	14.1

Commonwealth catches are believed to have been substantially underreported in the past which, coupled with limited information regarding recreational catches, represents a major source of uncertainty in estimating total mortality. The recreational fishery has heavily targeted Striped Trumpeter in the past, with an estimated 38 t in 2000/01 (Lyle 2005). Estimates of recreational catches in 2007/08 were highly uncertain but likely to be lower, with an estimated 19 t for combined catches of Striped and Bastard Trumpeter (Lyle et al. 2009). Estimates in 2011/12 and

2012/13 were 31.9 and 15.2 t respectively for Striped Trumpeter and exceeded the commercial catch for the species in both years (Fig. 22.1A). These catch estimates do not fully represent catches by charter boats. Apart from the north coast, Striped Trumpeter catches have historically been reported from all areas around the state (Figs. 22.1A and 22.2). Since 2009/10, catches in the northwest coast have increased to match those of the southeast.

The observed catch trends reflect the influence of especially strong year classes that entered the fishery between 1995/96 and 1997/98. Increased gillnet catches in 1998/99 followed by a decline suggest that the 1996 year-class, which would have recruited to the inshore gillnet fishery in 1998/99, may also have been relatively strong. The subsequent decline in gillnet catches presumably reflects the movement of these relatively strong year-classes offshore but also suggests that there has been limited recruitment in recent years. Industry representatives also suggest that the trip limit of 250 kg introduced in 2000 provided a strong disincentive for some operators to target the species and may have contributed to the sharp falls in drop-line and hand-line effort since 2000/01.

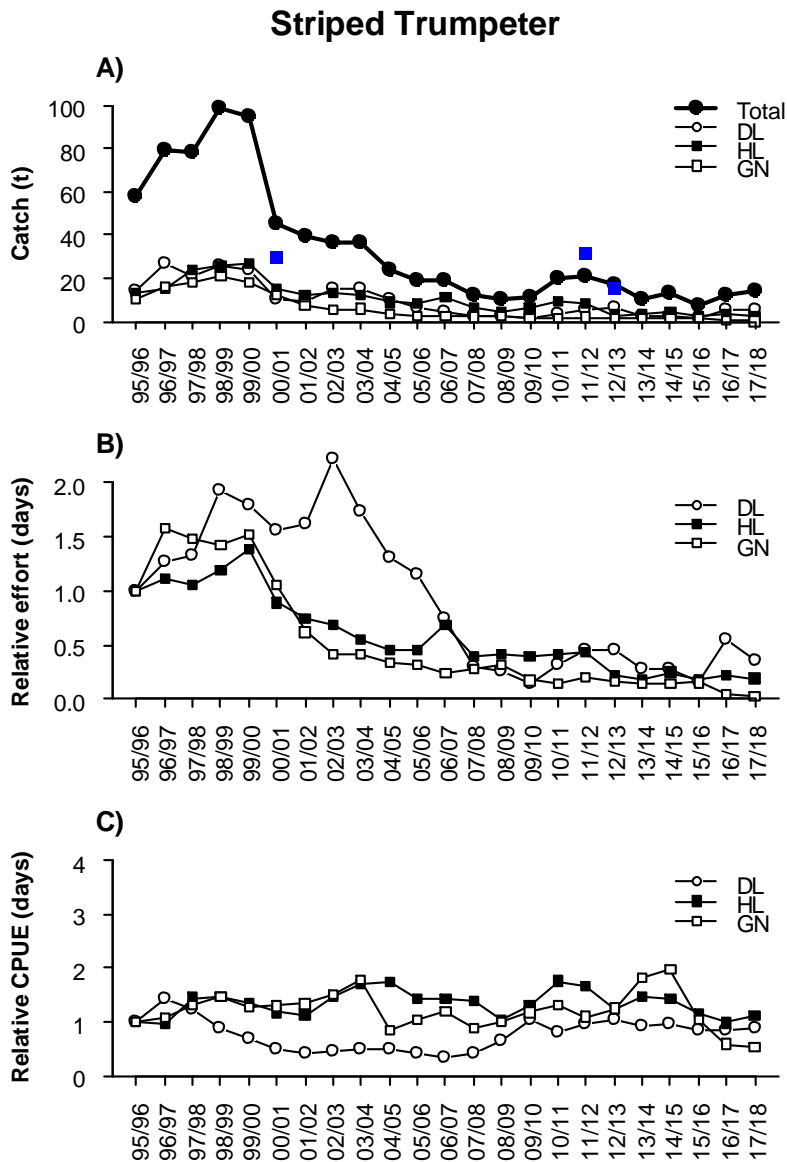
Fishing effort increased during the latter part of the 1990s, presumably linked to the increased availability of Striped Trumpeter (Fig. 22.1B). Subsequently, effort for all methods has declined, with drop-line and hand-line fishing now the dominant methods. Gillnet fishing effort reached a historical low in 2017/18. Fishing activity has been focused mainly on the southeast and west coasts and to a lesser extent off the east and northeast (Fig. 22.1B).

Drop-line catch rates have remained relatively stable since 2009/10 whereas there has been a slight decline in hand-line catch rates, although both are currently at levels comparable to the mid-1990s (Fig. 22.1C). Gillnet catch rates increased steadily up until 2003/04, despite declining catches during the latter half of the period. The sharp fall in gillnet catch rates after 2004/05 may have been influenced by the minimum size limit increase that took effect during 2004. Over the last three fishing seasons gillnet catch rates have further decreased, reaching a historical low in 2017/18 (Fig. 22.1C).

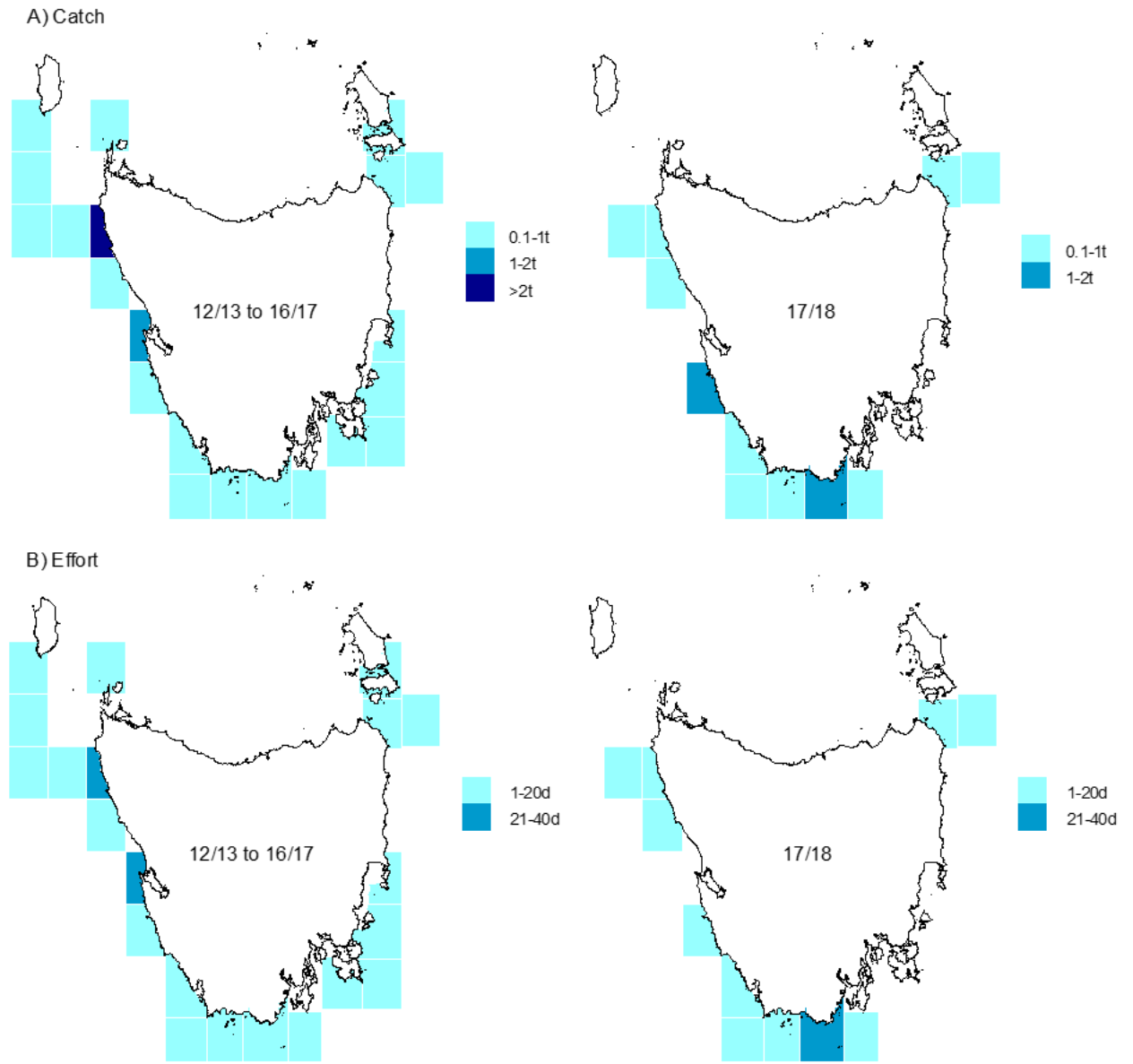
## **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian Scalefish Fishery, hand-lining was considered a medium risk with regard to Striped Trumpeter due to evidence of the population being subject to heavy fishing pressure from combined fishing methods. Hand-lining for Striped Trumpeter was also considered a medium risk to by-product mixed fish species, such as Jackass Morwong and Ocean Perch, due to the uncertainty surrounding their population status. Impacts on communities and protected species were generally low or negligible (Bell et al. 2016).

Post release survival is believed to be high in Striped Trumpeter, even when captured from relatively deep water (Lyle et al. 2014b). This is relevant given that bag and trip limits are relatively low and some catch may need to be released.



**Figure 22.1** A) Annual commercial catch (t) by gear and best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. HL=hand-line, GN=gillnet, DL=drop-line. Data includes Australian Fisheries Management Authority (AFMA) catch in State waters.



**Figure 22.2** (A) Striped Trumpeter catches (t) and (B) effort (days) for drop-line, hand-line and gillnet by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right). Data includes Australian Fisheries Management Authority (AFMA) catch in State waters.

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	• Catch > 3 <sup>rd</sup> highest catch value from the reference period (79.4 t)	No	
	• Catch < 3 <sup>rd</sup> lowest catch value from the reference period (23.9 t)	Yes	↓ 11.5 t (48.2%)
	• Catch variation from the previous year above the greatest inter-annual increase from the reference period (21.1 t)	No	
	• Catch variation from the previous year above the greatest inter-annual decrease from the reference period (49.5 t)	No	
	• Latest recreational catch estimate > recreational catch estimate from the reference period (19.6 t)	Yes	↑ 2.3 t (7.8%)
	• Proportion of recreational catch to total catch > previous proportion estimate (61.1% in 2011/12%)	No	
<b>Biomass</b>	• CPUE < 3 <sup>rd</sup> lowest CPUE value from the reference period (0.0210 t/days fished)	No	
	• Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0034)	No	
<b>Stock stress</b>	• Significant change in the size/age composition of commercial catches	No	
	• Significant numbers of unhealthy fish landed	No	

### Stock status

**RECOVERING**

The sharp decline in commercial catches since 2000/01 gave rise to concerns about the status of Striped Trumpeter stocks. As noted in previous assessments, strong recruitment variability could result in marked variation in population size, especially if there is a prolonged period of poor recruitment. Age composition data collected between 2010 and 2012 provided some encouraging signs regarding new recruits entering the fishery (4–6 year olds). The strength of these cohorts is difficult to assess, however age composition data for 2016/17 and 2017/18 appear to show that this recruitment has had a positive effect on the population, with an increasing proportion of fish aged between 7 and 13 years old evident from samples from these years. Moreover, that fishing mortality was estimated to be around half that of natural mortality suggests that fishing pressure at current levels is unlikely to cause the stock to become recruitment impaired.

Research undertaken during 2010 confirmed that the current minimum size limit was below the size at maturity (>60 cm TL cf minimum size limit of 55 cm TL) and that with the increasing proportion of small fish in the catch, due to the recent recruitment events, there was potential for growth overfishing if fishing pressure were to increase dramatically. Aligning the size limit with

size at maturity, thereby allowing fish to spawn before they become vulnerable to capture, would almost certainly increase spawning biomass and enhance recruitment potential.

Growing interest from the recreational sector suggests that recreational catches represent an increasingly significant proportion of total fishing mortality. While the implementation of bag limit of four fish and a boat limit of 20 fish for recreational fishers may help to constrain recreational harvest, the presence of small legal sized fish may provide an incentive for high grading as fishers seek to maximise the weight of their catch under the reduced bag limit. A state-wide survey currently underway will provide a new recreational catch estimate for 2017/18 (this will be presented in the 2018/19 assessment report).

A spawning season closure during September and October (not recognised by the Commonwealth managed sector), when fish are particularly vulnerable to capture, was introduced in 2009. Fishing trials have confirmed that while there may be some variability in the timing of spawning around the state, the closure encompasses the bulk of the spawning season in all areas.

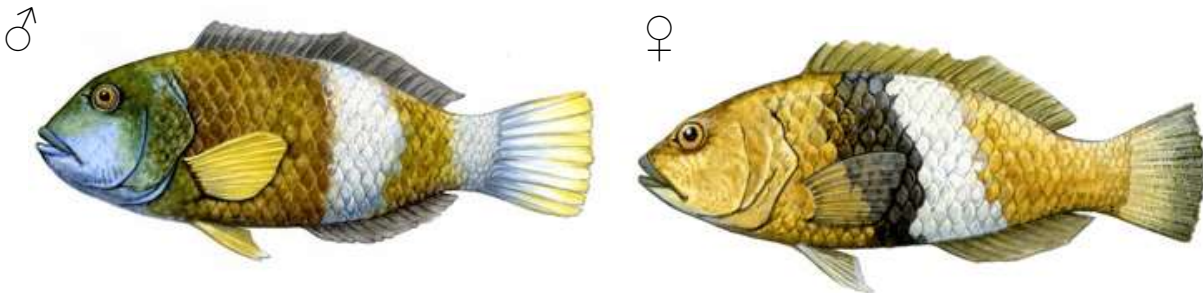
The influx of 4–6 year olds between 2010 and 2016 provided positive signs of population recovery after a prolonged period of low recruitment, which led to the previous stock status of transitional-recovering. Age composition data for 2016/17 and 2017/18 appear to show that this recruitment has had a positive effect on the population. On the basis of evidence for positive recruitment and the observed increase in older fish in age composition data in recent years, the stock is re-classified as recovering. Future recruitment and the relative strength of these year classes will need to be monitored over time to assess the extent of further population recovery.

# 23. Wrasse

## Bluethroat Wrasse *Notolabrus tetricus*

## Purple Wrasse *Notolabrus fuciola*

<b>STOCK STATUS</b>	<b>SUSTAINABLE</b>
Catches, effort and catch rates have remained relatively stable for almost a decade providing no indication that recent fishing mortality is too high. There is, however, some uncertainty over the size of the catch taken by rock lobster fishers and used for bait.	
<b>IMPORTANCE</b>	Key
<b>STOCK(S)</b>	Tasmanian Scalefish Fishery
<b>INDICATOR(S)</b>	Catch, effort and CPUE trends



*Notolabrus tetricus*  
Source: DPIPWE (by Peter Gouldthorpe)

### Species biology

Parameters	Estimates	Source
Habitat	<ul style="list-style-type: none"> <li><u>Bluethroat Wrasse</u>: Sheltered and exposed reefs, from 1 to 160 m depth.</li> <li><u>Purple Wrasse</u>: Predominantly on exposed reefs up to 90 m depth.</li> </ul>	Edgar (2008)
Distribution	<ul style="list-style-type: none"> <li><u>Bluethroat Wrasse</u>: from Sydney (New South Wales) to Ceduna (South Australia).</li> <li><u>Purple Wrasse</u>: New Zealand and Australian waters, from southern New South Wales to Kangaroo Island (South Australia).</li> </ul>	Edgar (2008)
Diet	<ul style="list-style-type: none"> <li>Both species consume a range of molluscs and crustaceans.</li> </ul>	Shepherd and Clarkson (2001) Denny and Schiel (2001)
Movement and stock structure	<ul style="list-style-type: none"> <li><u>Bluethroat Wrasse</u>: site-attached. Females have overlapping home ranges and males are territorial, at least during the reproductive season. No emigration of individuals between reefs has been noted for this species.</li> </ul>	Barrett (1995b)

	<ul style="list-style-type: none"> <li>• <u>Purple Wrasse</u>: site-attached, with no evidence of territorial behaviour. Movements between reefs is limited and likely to be restricted to a small proportion of the population.</li> <li>• No information on stock structure exists for either Bluethroat Wrasse or Purple Wrasse.</li> </ul>	
Natural mortality	<ul style="list-style-type: none"> <li>• Low adult mortality for both species. Estimated at <math>M = 0.2</math> for Bluethroat Wrasse.</li> </ul>	Smith et al. (2003) Barrett (1995a)
Maximum age	<ul style="list-style-type: none"> <li>• <u>Bluethroat Wrasse</u>: 11 years</li> <li>• <u>Purple Wrasse</u>: up to 24 years</li> </ul>	Barrett (1995a) Welsford (2003)
Growth	<ul style="list-style-type: none"> <li>• <u>Bluethroat Wrasse</u>: From 6 months onwards, growth (males and females confounded) follows a von Bertalanffy growth function with <math>L_{\infty} = 36.12, k = 0.2</math> and <math>t_0 = -0.35</math>. Length-weight relationship was set at <math>W = 0.0545L^{2.7157}</math> (both sexes) where <math>W</math> is weight (g) and <math>L</math> is the fork length (cm).</li> <li>• <u>Purple Wrasse</u>: From 2 years onwards, growth (males and females confounded) follows a Von Bertalanffy growth function with <math>L_{\infty} = 44.7, k = 0.085</math> and <math>t_0 = -3.23</math>. Length-weight relationship (both sexes) was set at <math>W = 0.0161L^{3.0407}</math> where <math>W</math> is weight (g) and <math>L</math> is the fork length (cm).</li> </ul>	Welsford (2003) Barrett (1995a) Unpublished data
Maturity	<ul style="list-style-type: none"> <li>• <u>Bluethroat Wrasse</u>: protogynous hermaphrodite (i.e. developing into female first before changing to male) with sex change happening between 27 and 32 cm in Tasmania. However, not all individuals undergo a sex inversion. Size at 50% maturity reached at 29.89 cm for females (corresponding to around 8 years old). Batch fecundity unknown.</li> <li>• <u>Purple Wrasse</u>: gonochoristic species (i.e. sex is fixed at maturity). Size at 50% maturity reached at 18.41 cm for females (corresponding to around 3 years old). Batch fecundity is estimated at <math>74,500 \pm 34,900</math> eggs/kg.</li> </ul>	Barrett (1995a) Hardwood and Lokman (2006) Unpublished data
Spawning	<ul style="list-style-type: none"> <li>• The spawning season for both species extends from August to January.</li> </ul>	Barrett (1995a)
Early life history	<ul style="list-style-type: none"> <li>• <u>Bluethroat Wrasse</u>: planktonic larval duration ranges from 44 to 66 days.</li> <li>• <u>Purple Wrasse</u>: planktonic larval duration ranges from 40 to 87 days.</li> <li>• Settlement on the reefs at around 10.4mm for both species.</li> </ul>	Welsford (2003)

## Background

Of the several species of Wrasse occurring in Tasmanian waters, Purple Wrasse (*Notolabrus fucicola*) and Bluethroat Wrasse (*N. tetricus*) are the main species taken commercially. Wrasse are targeted for live fish markets as well as being sold as dead product and utilised as bait for rock lobster (bait usage is likely to be under-reported). Fish marketed live are distinguished in the logbooks, and have accounted for over 90% of the total reported catch since 2001/02. Thus, trends in the live-fish fishery will ultimately be reflected in overall production levels. The two species of Wrasse have only been distinguished in catch returns since 2007. While there is an apparent market preference for Bluethroat Wrasse, Purple Wrasse are more robust for live handling.

<b>FISHING METHODS</b>	Fish trap and hand line.
<b>MANAGEMENT METHODS</b>	<p><b>Input control:</b></p> <ul style="list-style-type: none"> <li>• Gear licence (Scalefish fishing licence).</li> <li>• Species licence (Wrasse licence).</li> <li>• Rock lobster licence (for bait only).</li> </ul> <p><b>Output control:</b></p> <ul style="list-style-type: none"> <li>• Minimum size: 30 cm.</li> <li>• Possession limit of 10 and bag limit of 5 individuals for recreational fishers.</li> <li>• Limit of 30kg for landed dead wrasse if not the holder of a wrasse licence.</li> </ul>
<b>MAIN MARKET</b>	<ul style="list-style-type: none"> <li>• Interstate (live trade) and local (bait and food).</li> </ul>

## Current assessment

### Catch, effort and CPUE

Wrasse catches fluctuated between 75–110 t between 1995/96 and 2007/08, peaking at 113 t in 2006/07 (Fig. 23.1A). Reported catches fell in 2008/09 to 68.2 t and have ranged around 64 t to 82 t since then. In 2017/18 total commercial landings were 83.8 t (comprising 62.2 t of Bluethroat Wrasse, 21.4 t of Purple Wrasse, and 0.2 t of unspecified Wrasse). The lower catches since the late 2000s are largely due to a decline in the use of fish traps as a result of the prohibition on the use of abalone gut as bait. This prohibition was a response to the appearance of the abalone viral ganglioneuritis in Victoria and forced fishers to seek alternative but less effective baits. Catches, effort and catch rates subsequently declined for fish traps, particularly in the southeast (Fig. 23.1) where fish traps had been the dominant fishing method. By contrast, hand-line catch and effort has remained relatively stable since that time. Wrasse are targeted all around Tasmania, the west coast being the least fished area (Fig. 23.2).

It is important to note that Wrasse caught and used as bait in rock lobster pots are not included in the catches outlined above as historically they have been underreported. In 2017, the total recorded catch from the rock lobster fishery was 563 kg, up from just 126 kg reported in 2016.

With Bluethroat Wrasse being more susceptible to line methods and Purple Wrasse more vulnerable to trap capture, Bluethroat Wrasse are now taken in larger quantities in the live fishery. Gillnets account for the bulk of the remaining catch (< 4 t) but because survival in nets is poor, gillnet caught Wrasse are rarely marketed live.

Recreational catches were estimated at 13.6 t in 2000/01 (Lyle 2005) and 10.3 t in 2007/08 (Lyle et al. 2009) and 6.4 t in 2012/13 (Lyle et al. 2014b), representing around 10% of the total Tasmanian catch. Further, Bluethroat Wrasse are a reasonably common by-catch of recreational

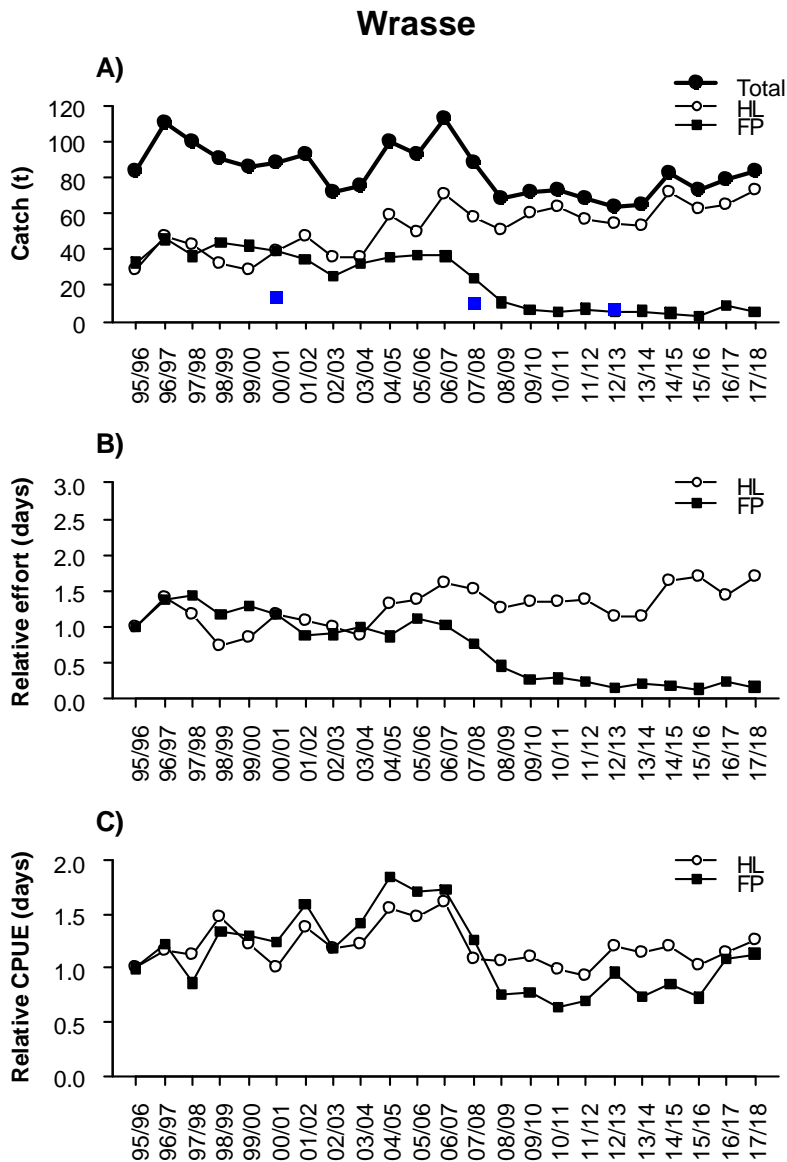
gillnet fishers and a recent study showed that they have a moderate to low post-release survival, particularly when gillnets are deployed for > 4 hours (Lyle et al. 2014a).

Catch rate trends imply that Wrasse stocks have not been impacted significantly by the fishery until recent years. Following a gradual increase in catch rates up until the late 2000s, handline catch rates have generally stabilised at an intermediate level since 2008/09. Catch rates for fish traps peaked in 2004/05, but declined since, largely influenced by changes in permissible bait types (Fig. 23.1C). Fish trap catch rates in 2017/18 were the highest recorded since 2007/08.

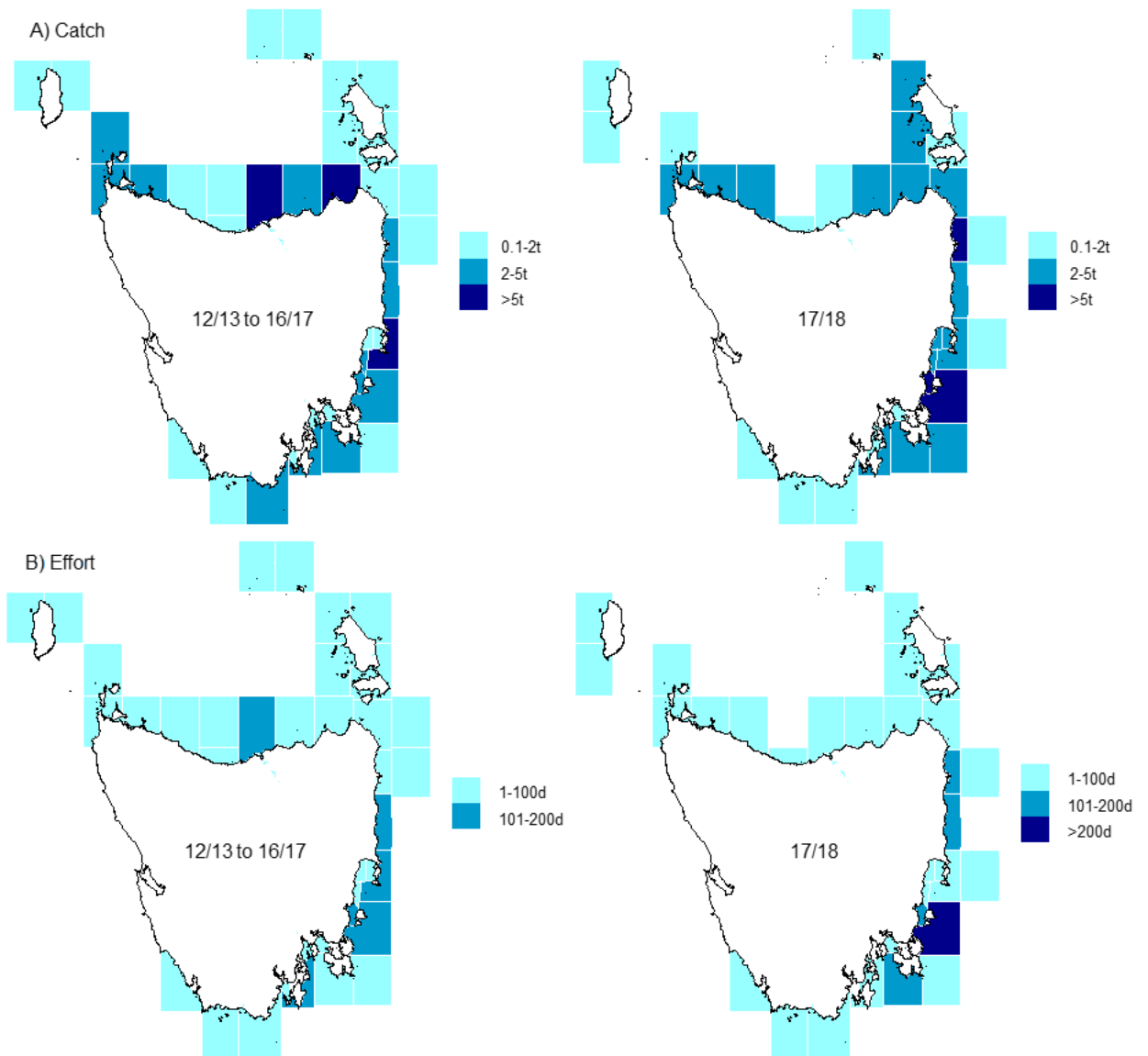
The broad-scale analyses are, however, generally insensitive to changes in abundance at the level of individual reefs at which the fishery impacts the stocks. Marked regional shifts of effort have occurred in the fishery over the years and may have masked localised depletion, with fishers moving to new or lightly fished areas to maintain catches and catch rates.

### **Ecological Risk Assessment**

In the 2012/13 ERA of the Tasmanian Scalefish Fishery, fish trapping was considered a low risk to Wrasse species, because fishing effort at the time had reduced due to the banning of the preferred bait for trap fishing (abalone guts). Risks to by-product species such as Leatherjackets were very low due to small levels of catch and the risk to non-retained species and the general ecosystem were assessed as either low or negligible (Bell et al. 2016).



**Figure 23.1** A) Annual commercial catch (t) by gear best estimates of recreational catches (blue squares). B) Commercial effort by method based on day fished relative to 1995/96. C) Commercial catch per unit effort (CPUE) based on weight per day fished relative to 1995/96. HL=hand-line, FP=fish trap.



**Figure 23.2** (A) Wrasse catches (t) and (B) effort (days) for fish trap, hand-line and by fishing blocks averaged from 2012/13 to 2016/17 (left) and during 2017/18 (right).

## Reference points

Performance indicators	Proposed reference points	Breached?	By how much?
<b>Fishing mortality</b>	<ul style="list-style-type: none"> <li>Catch &gt; 3<sup>rd</sup> highest catch value from the reference period (94.1 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Catch &lt; 3<sup>rd</sup> lowest catch value from the reference period (83.4 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Catch variation from the previous year above the greatest inter-annual increase from the reference period (26.7 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Catch variation from the previous year above the greatest inter-annual decrease from the reference period (37.9 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Latest recreational catch estimate &gt; recreational catch estimate from the reference period (13.6 t)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Proportion of recreational catch to total catch &gt; previous proportion estimate (13.1% in 2007/08)</li> </ul>	No	
<b>Biomass</b>	<ul style="list-style-type: none"> <li>CPUE &lt; 3<sup>rd</sup> lowest CPUE value from the reference period (0.0135 t/days fished)</li> </ul>	No	
	<ul style="list-style-type: none"> <li>Rate of CPUE decline over last 3 years is greater than the largest 3-year CPUE decline during the reference period (-0.0014)</li> </ul>	No	
<b>Stock stress</b>	<ul style="list-style-type: none"> <li>Significant change in the size/age composition of commercial catches</li> </ul>	Not assessed	
	<ul style="list-style-type: none"> <li>Significant numbers of unhealthy fish landed</li> </ul>	No	

## Stock status

**SUSTAINABLE**

The minimum size limit provides good protection (several years after reaching the size at maturity) for the spawning stock of Purple Wrasse and for female Bluethroat Wrasse. The limit does not, however, provide the same level of protection for male Bluethroat Wrasse which are derived through sex change from mature females, typically at sizes after they have entered the fishery. This, along with the fact that male wrasse are strongly site attached and have higher catchability (being more aggressive than females), suggests that they are vulnerable to fishing. An underwater visual census in 2008, however, showed that wrasse were more abundant in accessible sites (such as areas near boat ramps) compared to other fish and crustaceans, which may indicate a greater resilience to the impact of fishing mortality (Stuart-Smith et al. 2008), while other surveys have observed higher abundances outside of areas protected from fishing than inside (Walsh et al. 2017). Regardless, in extreme situations it is possible that localised heavy fishing pressure could result in 'sperm shortage' in Bluethroat Wrasse that could affect spawning success even though there may be a robust population of mature (sub-legal size) females present. The removal of the maximum size limit for Wrasse in 2004 may have exacerbated this potential problem. However, there are no clear indications of spawning stock shortages in Tasmania.

As highlighted in previous assessments the banning of the preferred bait for trap fishing in 2008 has dramatically changed fisher's behaviour and fish traps are now rarely used. This has resulted in Bluethroat Wrasse contributing a greater proportion of the catch than beforehand. Industry reported that there is no acceptable substitute for abalone guts (i.e. no other bait yields the same return) and that trap fishing has become less viable. This is supported by catch rate data in Figure 23.1.

Previous assessments have highlighted that increasing catches up to 2006/07 reflected a strong interest in the species, and there were concerns that these levels of fishing mortality were not sustainable, as justified by the decline in catch rates (see Fig. 23.1). In 2017/18, hand-line catches and effort increased slightly from previous years reaching record high, and will need to be closely monitored as there is the potential for localised depletions of legal-sized Wrasse, especially if effort becomes concentrated in particular regions. Catch rates have, however, remained relatively stable for almost a decade providing no evidence that the level of fishing mortality has been too high. Based on this stability, the stock is classified as sustainable.

# References

- AFMA. 2008. Blue warehou (*Seriolella brama*) stock rebuilding strategy: Southern and Eastern Scalefish and Shark Fishery (SESSF) - December 2008. Australian Fisheries Management Authority, Canberra.
- AFMA. 2011. Blue warehou (*Seriolella brama*) stock rebuilding strategy: Southern and Eastern Scalefish and Shark Fishery (SESSF) - Revised July 2011. Australian Fisheries Management Authority, Canberra.
- AFMA. 2012. Blue warehou (*Seriolella brama*) stock rebuilding strategy: Southern and Eastern Scalefish and Shark Fishery (SESSF) - Revised April 2012. Australian Fisheries Management Authority, Canberra.
- AFMA. 2014. Blue warehou (*Seriolella brama*) Stock Rebuilding Strategy. Australian Fisheries Management Authority, Canberra, ACT.
- Andrews J., J. Lyle, K. Hall, L. Maloney, and T. Emery. 2016. Eastern School Whiting *Sillago flindersi*. In: C. Stewardson, J. Andrews, C. Ashby, M. Haddon, K. Hartmann, P. Hone, P. Horvat, S. Mayfield, A. Roelofs, K. Sainsbury, T. Saunders, J. Stewart, I. Stobutzki and B. Wise (eds) 2016. Status of Australian fish stocks reports 2016, Fisheries Research and Development Corporation, Canberra.
- Annala, J. H. 1994. Report from the Fishery Assessment Plenary, May 1994: stock assessments and yield estimates (unpublished). MAF Fisheries Greta Point library, Wellington
- Ayling, G. M., K. C. Wilson, and D. A. Ratkowsky. 1975. Sand flathead (*Platycephalus bassensis*), an indicator species for mercury pollution in Tasmanian waters. Marine Pollution Bulletin 6:142–144.
- Bani, A. 2005. Temporal and spatial variability of the life history characteristics of sand flathead, *Platycephalus bassensis*. University of Tasmania, Hobart.
- Bani, A., and N. A. Moltschaniwsky. 2008. Spatio-temporal variability in reproductive ecology of sand flathead, *Platycephalus bassensis*, in three Tasmanian inshore habitats: potential implications for management. Journal of Applied Ichthyology 24:555–561.
- Barnes, L. M., C. A. Gray, and J. E. Williamson. 2011. Divergence of the growth characteristics and longevity of coexisting Platycephalidae (Pisces). Marine and Freshwater Research 62:1308–1317.
- Barrett, N. S. 1995a. Short- and long-term movement patterns of six temperate reef fishes (Families Labridae and Monacanthidae). Marine and Freshwater Research 46:853–860.
- Barrett, N. S. 1995b. Aspects of the biology and ecology of six temperate reef fishes (Families: Labridae and Monacanthidae). University of Tasmania, Hobart.
- Bell, J. D., J. Lyle, J. Andre, and K. Hartmann. 2016. Tasmanian scalefish fishery: ecological risk assessment. Institute for Marine and Antarctic Studies, Hobart, Australia.
- Bertoni, M. 1995. The reproductive biology and feeding habits of the snook, *Sphyraena novaehollandiae*, in South Australian waters. Southern Fisheries 3:34–35.
- Blackburn, M. 1960. A study of condition (weight for length) of Australian barracouta, *Thyrsites atun* (Auphrasen). Australian Journal of Marine and Freshwater Research 11:14–41.
- Blackburn, M., and P. E. Gartner. 1954. Populations of barracouta, *Thyrsites atun* (Euphrasen), in Australian waters. Australian Journal of Marine and Freshwater Research 5:411–468.
- Bruce, B. D., F. J. Neira, and R. W. Bradford. 2001a. Larval distribution and abundance of blue and spotted warehou (*Seriolella brama* and *S. punctata*: Centrolophidae) in southeastern Australia. Marine and Freshwater Research 52:631–639.
- Bruce, B. D., K. Evans, C. A. Sutton, J. W. Young, and D. M. Furlani. 2001b. Influence of mesoscale oceanographic processes on larval distribution and stock structure in jackass morwong (*Nemadactylus macropterus*: Cheilodactylidae). ICES Journal of Marine Science 58:1072–1080.
- Bulman, C., F. Althaus, X. He, N. J. Bax, and A. Williams. 2001. Diets and trophic guilds of demersal fishes of the south eastern Australian shelf. Marine and Freshwater Research 52:537–548.
- Bulman, C., S. Condie, J. Findlay, B. Ward, and J. Young. 2008. Management zones from small pelagic fish species stock structure in southern Australian waters. FRDC report 2006/076. CSIRO, Hobart.

- Burchmore, J. J., D. A. Pollard, M. J. Middleton, J. D. Bell, and B. C. Pease. 1988. Biology of four species of whiting (Pisces: *Sillaginidae*) in Botany Bay, New South Wales. *Australian Journal of Marine and Freshwater Research* 39: 709–727.
- Butcher, A. R., and A. R. Hagedoorn. 2003. Age, growth and mortality estimates of stout whiting, *Sillago robusta* Stead (Sillaginidae), from Southern Queensland, Australia. *Asian Fisheries Science* 16: 215–228.
- BWAG. 1998. Stock assessment report: Blue warehou 1998. Australian Fisheries Management Authority, Canberra.
- Chubb, C. F., I. C. Potter, C. J. Grant, R. C. J. Lenanton, and J. Wallace. 1981. Age structure, growth rates and movements of sea mullet, *Mugil cephalus* L. and yellow-eye mullet, *Aldrichetta forsteri* (Valenciennes), in the Swan-Avon River System, Western Australia. *Australian Journal of Marine and Freshwater Research* 32:605–628.
- Coleman, N., and M. Mobley. 1984. Diets of commercially exploited fish from Bass Strait and adjacent Victorian waters, South-eastern Australia. *Australian Journal of Marine and Freshwater Research* 35:549–560.
- Crawford, C. M. 1984. An ecological study of Tasmanian flounder. University of Tasmania, Hobart.
- Crawford, C. M. 1986. Development of eggs and larvae of the flounders *Rhombosolea tapirina* and *Ammotretis rostratus* (Pisces: Pleuronectidae). *Journal of Fish Biology* 29:325–334.
- Curtis, T. D., and J. S. Shima. 2005. Geographic and sex specific variation in growth of yellow eyed mullet, *Aldrichetta forsteri*, from estuaries around New Zealand. *New Zealand Journal of Marine and Freshwater Research* 39:1277–1285.
- Denny, C. M., and D. R. Schiel. 2001. Feeding ecology of the banded wrasse *Notolabrus fucicola* (Labridae) in southern New Zealand: prey items, seasonal differences, and ontogenetic variation. *New Zealand Journal of Marine and Freshwater Research* 35:925–933.
- Dixon, P. I. 1987. Stock identification and discrimination of commercially important whittings in Australian waters using genetic (FIRTA 83/16): Final Report. University of New South Wales, Centre for Marine Science.
- DPIF. 1998. Scalefish Fishery: Policy Document. Tasmanian Department of Primary Industry and Fisheries, Hobart.
- DPIPWE. 2015. Framework to support developmental fishing program for Australian sardines. Department of Primary Industries, Parkes, Water & Environment, Tasmania.
- Dunning, M. C. 1998. Zoogeography of arrow squids (Cephalopoda: Ommastrephidae) in the Coral and Tasman Seas, southwest Pacific. Pages 435-453 in N. A. Voss, M. Vecchione, R. B. Toll, and M. J. Sweeney, editors. *Systematics and Biogeography of Cephalopods*. Smithsonian Institution Press, Washington D.C.
- Dunning, M. C., and E. C. Förch. 1998. A review of the systematics, distribution and biology of arrow squids of the genus *Nototodarus* Pfeffer, 1912 (Cephalopoda, Ommastrephidae). Pages 393–404 in N. A. Voss, M. Vecchione, R. B. Toll, and M. J. Sweeney, editors. *Systematics and Biogeography of Cephalopods*. Smithsonian Institution Press, Washington D.C.
- Edgar, G. J. 1997. *Australian marine life: the plants and animals of temperate waters*. Reed Books, Melbourne.
- Edgar, G. D. 2008. *Australian marine life: the plants and animals of temperate waters*. New Holland Publishers, Sydney.
- Edgar, G. J., and N. S. Barrett. 1999. Effects of the declaration of marine reserves on Tasmanian reef fishes, invertebrates and plants. *Journal of Experimental Marine Biology and Ecology* 242:107–144.
- Edgar, G. J., N. S. Barrett, and A. J. Morton. 2004. Patterns of fish movement on eastern Tasmanian rocky reefs. *Environmental Biology of Fishes* 70:273–284.
- Elliott, N. G., and R. D. Ward. 1994. Enzyme variation in jackass morwong, *Nemadactylus macropterus* (Schneider 1801) (Teleostei: Cheilodactylidae) from Australian and New Zealand waters. *Australian Journal of Marine and Freshwater Research* 45:51–67.

- Emery, T., Lyle, J. and Hartmann, K. 2016. Tasmanian Scalefish fishery assessment 2014/15. Institute for Marine and Antarctic Studies, Hobart.
- Emery, T., Lyle, J. and Hartmann, K. 2017. Tasmanian Scalefish fishery assessment 2015/16. Institute for Marine and Antarctic Studies, Hobart.
- Ewing, G., and J. Lyle. 2015. Low-cost monitoring regime to assess relative abundance and population characteristics of sand flathead. Institute for Marine and Antarctic Studies, Hobart, Tasmania.
- Ewing, G., and J. Lyle. 2018. Low-cost monitoring regime to assess relative abundance and population characteristics of sand flathead 2018 update. Institute for Marine and Antarctic Studies, Hobart, Tasmania.
- Fairbridge, W. S. 1951. The New South Wales tiger flathead, *Neoplatycephalus macrodon* (Ogilby). I. Biology and age determination. Australian Journal of Marine and Freshwater Research 2:117–178.
- Ferguson, G. 2006. Fisheries biology of the greenback flounder *Rhombosolea tapirina* (Günther 1862) (Teleostei: Pleuronectidae) in South Australia. SARDI Aquatic Sciences Publication No. RD06/0008-1. South Australian Research and Development Institute (Aquatic Sciences), Adelaide.
- FishBase. 2013. Monacanthidae. Accessed 30/09/2013.
- Fletcher, W. J. (1995). Application of the otolith weight-age relationship for the pilchard *Sardinops sagax neopilchardus*. Canadian Journal of Fisheries and Aquatic Sciences 52:657–664.
- Flood, M., I. Stobutzki, J. Andrews, G. Begg, R. Fletcher, C. Gardner, J. Kemp, A. Moore, A. O'Brien, R. Quinn, J. Roach, K. R. Rowling, K. Sainsbury, T. Saunders, T. Ward, and M. Winning. 2012. Status of key Australian fish stocks reports 2012. Fisheries Research and Development Corporation, Canberra.
- Flood, M., I. Stobutzki, J. Andrews, C. Ashby, G. Begg, R. Fletcher, C. Gardner, L. Georgeson, S. Hansen, K. Hartmann, P. Hone, P. Horvat, L. Maloney, B. McDonald, A. Moore, A. Roelofs, K. Sainsbury, T. Saunders, T. Smith, C. Stewardson, J. Stewart, and B. Wise. 2014. Status of key Australian fish stocks reports 2014. Fisheries Research and Development Corporation, Canberra.
- Francis, M. 2001. Coastal fishes of New Zealand. Reed Publishing, Auckland.
- Frimodt, C. 1995. Multilingual illustrated guide to the world's commercial coldwater fish. Wiley- Blackwell, Oxford.
- Gavrilov, G. M., and N. P. Markina. 1979. The feeding ecology of fishes of the genus *Serirolella* (fam. Nomeidae) on the New Zealand plateau. Journal of Ichthyology 19:128–135.
- Godfriaux, B. L. 1974. Food of tarakihi in western Bay of Plenty and Tasman Bay, New Zealand. New Zealand Journal of Marine and Freshwater Research 8:111–153.
- Goldworthy, S. D., C. Bulman, X. He, J. Larcombe, and C. Littnan. 2003. Trophic interactions between marine mammals and Australian fisheries: ecosystem approach. Pages 62–99 in N. Gales, M. Hindell, and R. Kirkwood, editors. Marine mammals: fisheries, tourism and management issues. CSIRO Publishing, Collingwood, Victoria.
- Gomon, M., D. Bray, and R. Kuitert. 2008. Fishes of Australia's southern coast. Reed New Holland.
- Gorman, T. B. S. 1962. Yellow-eyed mullet *Aldrichetta forsteri* Cuvier and Valenciennes in Lake Ellesmere.
- Grant, C. J., T. R. Cowper, and D. D. Reid. 1978. Age and growth of snoek, *Leionura atun* (Euphrasen), in south-eastern Australian waters. Australian Journal of Marine and Freshwater Research 29:435–444.
- Grant, E. M. 1991. Grant's Fishes of Australia. E.M. Grant Pty. Ltd., Queensland.
- Grewe, P. M., A. J. Smolenski, and R. D. Ward. 1994. Mitochondrial DNA variation in jackass morwong, *Nemadactylus macropterus* (Teleostei: Cheilodactylidae) from Australian and New Zealand waters. Canadian Journal of Fisheries and Aquatic Sciences 51:1101–1109.
- Haddon, M., and A. Punt. 2018. simpleSA: A package containing functions to facilitate relatively simple stock assessments. R package version 0.1.10.
- Hardwood, N. J., and M. P. Lokman. 2006. Fecundity of banded wrasse (*Notolabrus fucicola*) from Otago, Southern New Zealand. New Zealand Journal of Marine and Freshwater Research 40:467–476.

- Harries, D. N., and R. L. Croome. 1989. A review of past and present inshore gill netting in Tasmania with particular reference to the bastard trumpeter, *Latridopsis forsteri* Castelnau. Papers and Proceedings of the Royal Society of Tasmania 123:97–110.
- Harries, D. N., and P. S. Lake. 1985. Aspects of the biology of inshore populations of Bastard Trumpeter, *Latridopsis forsteri* (Castleneau, 1872) in Tasmanian waters. Tasmanian Fisheries Research 27:19–43.
- Hartmann, K., and J. M. Lyle. 2011. Tasmanian Scalefish Fishery-2009/10. Tasmanian Aquaculture and Fisheries Institute, Hobart.
- Henry, G. W., and J. M. Lyle. 2003. National recreational and indigenous fishing survey. Final report to FRDC. Project No 99/158. Canberra
- Hobday, D. K., and J. W. J. Wankowski. 1987. School whiting *Sillago bassensis flindersi*: reproduction and fecundity in eastern Bass Strait, Australia. Internal Report 153. Victorian Department of Conservation, Forests and Lands. Fisheries Division.
- Hoedt, F. E., and W. F. Dimmlich. 1995. Egg and larval abundance and spawning localities of the anchovy (*Engraulis australis*) and pilchard (*Sardinops neopilchardus*) near Phillip Island, Victoria. Marine and Freshwater Research 46:735–743.
- Hoening, J. M., and W. D. Lawing. 1982. Estimating the mortality rate using the maximum order statistic for age. ICES Journal of Marine Science 7: 13.
- Hurst, R. J., S. L. Ballara, and D. MacGibbon. 2012. Fishery characterisation and standardised CPUE analyses for barracouta, *Thyrsites atun*, (Euphrasen, 1791) (Gempylidae), 1989–90 to 2007–08. New Zealand Fisheries Assessment Report 2012/12 NIWA, Wellington.
- Hutchinson, W. 1993. The reproductive biology and induced spawning of striped trumpeter, *Latris lineata*. University of Tasmania, Hobart.
- Izzo, C., T. M. Ward, A. R. Ivey, I. M. Suthers, J. Stewart, S. C. Sexton, B. M. Gillanders. (2017). Integrated approach to determining stock structure: implications for fisheries management of sardine, *Sardinops sagax*, in Australian waters. Reviews in Fish Biology and Fisheries 27: 267–284.
- Jackson, G. D., B. McGrath Steer, S. Wotherspoon, and A. J. Hobday. 2003. Variation in age, growth and maturity in the Australian arrow squid *Nototodarus gouldi* over time and space-what is the pattern? Marine Ecology Progress Series 264:57–71.
- Jackson, G. D., R. K. O'Dor, and Y. Andrade. 2005a. First tests of hybrid acoustic/archival tags on squid and cuttlefish. Marine and Freshwater Research 56:425–430.
- Jackson, G. D., S. Wotherspoon, and B. L. McGrath-Steer. 2005b. Temporal population dynamics in arrow squid *Nototodarus gouldi* in southern Australian waters. Marine Biology 146:975–983.
- Jenkins, G. P. 1986. Composition, seasonality and distribution of ichthyoplankton in Port Phillip Bay, Victoria. Australian Journal of Marine and Freshwater Research 37:507–520.
- Jones, G. K. 1990. Growth and mortality in a lightly fished population of garfish (*Hyporhamphus melanochir*), in Baird Bay, South Australia. Transactions of the Royal Society of South Australia 114:37–45.
- Jones, G. K., Q. Ye, S. Ayyazian, and P. Coutin. 2002. Fisheries ecology and habitat ecology of southern sea garfish (*Hyporhamphus melanochir*) in southern Australian waters. South Australian Research and Development Institute.
- Jordan, A. R. 2001a. Reproductive biology, early life-history and settlement distribution of sand flathead (*Platycephalus bassensis*) in Tasmania. Marine and Freshwater Research 52:589–601.
- Jordan, A. R. 2001b. Age, growth and spatial and interannual trends in age composition of jackass morwong, *Nemadactylus macropterus*, in Tasmania. Marine and Freshwater Research 52:641–660.
- Jordan, A. R., D. M. Mills, G. Ewing, and J. M. Lyle. 1998. Assessment of inshore habitats around Tasmania for life-history stages of commercial finfish species. Tasmanian Aquaculture and Fisheries Institute, Hobart.
- Kailola, P. J., M. J. Williams, P. C. Stewart, R. E. Reichelt, A. McNee, and C. Grieve. 1993. Australian fisheries resources. Bureau of Resource Sciences, Canberra.

- Klaer, N. 2010. Tiger flathead (*Neoplatycephalus richardsoni*) stock assessment based on data up to 2009. CSIRO, Hobart.
- Klumpp, D. W., and P. D. Nichols. 1983. Nutrition of the southern sea garfish (*Hyporhamphus melanochir*): gut passage rate and daily consumption of two food types and assimilation of seagrass components. Marine Ecology Progress Series 12:207–216.
- Knuckey, I. A. 2006. Southern and Eastern Scalefish and Shark Fishery – Bycatch Utilisation Scoping Study. Department of Primary Industries Agribusiness Group, Victoria.
- Knuckey, I. A., and K. P. Sivakumaran. 2001. Reproductive characteristics and per-recruit analyses of blue warehou (*Seriola lalandi*): implications for the south-east fishery of Australia. Marine and Freshwater Research 52:575–587.
- Last, P. R., E. O. G. Scott, and F. H. Talbot. 1983. Fishes of Tasmania. Tasmanian Fisheries Development Authority Hobart.
- Lo, N. C. H., B. J. Macewicz, and D. A. Griffith. 2005. Spawning biomass of Pacific sardine (*Sardinops sagax*) from 1994–2004 off California. California Cooperative Oceanic Fisheries Investigations Reports 46:93–112.
- Lyle, J. M. 1999. Licensed recreational fishing and an evaluation of recall biases in the estimation of recreational catch and effort. Final Report to the Marine Recreational Fishing Council Tasmanian Aquaculture and Fisheries Institute, Hobart.
- Lyle, J. M. 2000. Assessment of the licensed recreational fishery of Tasmania (Phase 2). Final Report to FRDC. Project 96/161. Tasmanian Aquaculture and Fisheries Institute Hobart.
- Lyle, J. M. 2005. 2000/01 survey of recreational fishing in Tasmania. Tasmanian Aquaculture and Fisheries Institute, Hobart.
- Lyle, J. M., and W. B. Ford. 1993. Review of trawl research 1979–1987, with summaries of biological information for the major species, with summaries of biological information for the major species. Technical Report 46. Department of Sea Fisheries, Tasmania, Hobart.
- Lyle, J. M., and A. R. Jordan. 1999. Tasmanian scalefish fishery assessment - 1998. Tasmanian Aquaculture and Fisheries Institute, Hobart.
- Lyle, J. M., and R. Murphy. 2001. Long distance migration of striped trumpeter. Fishing Today 14:6.
- Lyle, J. M., and S. R. Tracey. 2012. Recreational gillnetting in Tasmania – an evaluation of fishing practices and catch and effort. Report to FishWise. IMAS, Hobart.
- Lyle, J. M., K. Krusic-Golub, and A. K. Morison. 2000. Age and growth of jack mackerel and the age structure of the jack mackerel purse seine catch. FRDC project no 1995/034. TAFI, Hobart.
- Lyle, J. M., N. Moltschaniwskyj, A. J. Morton, I. W. Brown, and D. Mayer. 2007. Effects of hooking damage and hook type on post-release survival of sand flathead (*Platycephalus bassensis*). Marine and Freshwater Research 58:445–453.
- Lyle, J. M., S. R. Tracey, K. E. Stark, and S. Wotherspoon. 2009. 2007-08 survey of recreational fishing in Tasmania. Tasmanian Aquaculture and Fisheries Institute, Hobart.
- Lyle, J. M., K. E. Stark, and S. R. Tracey. 2014a. 2012-13 survey of recreational fishing in Tasmania. Institute for Marine and Antarctic Studies, Hobart, Tasmania.
- Lyle, J. M., J. D. Bell, B. M. Chuwen, N. S. Barrett, S. R. Tracey, and C. D. Buxton. 2014b. Assessing the impacts of gillnetting in Tasmania: Implications for by-catch and biodiversity. Institute for Marine and Antarctic Studies, Hobart, Tasmania.
- Machida, S. 1983. A brief review of the squid survey by Hoyo Maru No. 67 in southeast Australian waters in 1979/80. Memoirs of the National Museum Victoria 44:291–295.
- May, J. L., and J. G. H. Maxwell. 1986. Trawl fish from temperate waters of Australia. CSIRO Division of Fisheries Research, Hobart.
- Millar, R. B. (2015). A better estimator of mortality rate from age-frequency data. Canadian Journal of Fisheries and Aquatic Sciences 72: 364–375.
- Moltschaniwskyj, N. A., and G. T. Pecl. 2003. Small-scale spatial and temporal patterns of egg production by the temperate loliginid squid *Sepioteuthis australis*. Marine Biology 142:509–516.

- Moore, B. R., J. M. Lyle, and K. Hartmann. 2018. Tasmanian Banded Morwong Fishery Assessment. 2017/18. Institute for Marine and Antarctic Studies, Hobart.
- Morison, A. K., I. A. Knuckey, C. A. Simpfendorfer, and R. C. Buckworth. 2012. 2011 Stock assessment summaries for the South East Scalefish and Shark Fishery. Southern and Eastern Scalefish and Shark Fishery Assessment Group, AFMA.
- Murphy, R., and J. M. Lyle. 1999. Impact of gillnet fishing on inshore temperate reef fished, with particular reference to banded morwong. Final report to FRDC, project No. 95/145. Tasmanian Aquaculture and Fisheries Institute, Hobart.
- Nakamura, I., and N. V. Parin. 1993. FAO Species Catalogue. Vol. 15. Snake mackerels and cutlassfishes of the world (families Gempylidae and Trichiuridae). An annotated and illustrated catalogue of the snake mackerels, snoeks, escolars, gemfishes, sackfishes, domine, oilfish, cutlassfishes, scabbardfishes, hairtails, and frostfishes known to date. FAO.
- Neira, F. J., A. G. Miskiewicz, and T. Trnski. 1998. Larvae of temperate Australian fishes: laboratory guide for larval fish identification. University of Western Australia Press.
- Neira, F. J., M. I. Sporcic, and A. R. Longmore. 1999. Biology and fishery of pilchard, *Sardinops sagax* (Clupeidae), within a large south-eastern Australian bay. Marine and Freshwater Research 50:43–55.
- Nichols, D. S., D. Williams, G. A. Dunstan, P. D. Nichols, and J. K. Volkman. 1994. Fatty acid composition of Antarctic and temperate fish of commercial interest. Comparative Biochemistry and Physiology Part B: Biochemistry and Molecular Biology 107:357–363.
- Norman, M. 2000. Cephalopods, a world guide. ConchBooks, Hackenheim.
- Norman, M., and A. Reid. 2000. A guide to squid, cuttlefish and octopuses of Australasia. CSIRO Publishing/The Gould League of Australia, Collingwood/Moorabbin
- O'Shea, S., K. S. Bolstad, and P. A. Ritchie. 2004. First records of egg masses of *Nototodarus gouldi* McCoy, 1888 (Mollusca: Cephalopoda: Ommastrephidae), with comments on eggmass susceptibility to damage by fisheries trawl. New Zealand Journal of Zoology 31:161–166.
- O'Sullivan, D., and J. M. Cullen. 1983. Food of the Squid *Nototodarus gouldi* in Bass Strait. Australian Journal of Marine and Freshwater Research 34:261–285.
- Parker, S., and D. Fu. 2011. Age composition of the commercial tarakihi (*Nemadactylus macropterus*) catch in quota management area TAR 2 in fishery year 2009–2010. New Zealand Fisheries Assessment Report 2011/59.
- Patterson, H., L. Georgeson, I. Stobutzki, and R. Curtotti. 2015. Fishery status reports 2015. Australian Bureau of Agricultural and Resource Economics and Sciences, Canberra.
- Patterson, H., R. Noriega, L. Georgeson, I. Stobutzki and R. Curtotti. 2016. Fishery status reports 2016, Australian Bureau of Agricultural and Resource Economics and Sciences, Canberra.
- Patterson, H., R. Noriega, L. Georgeson, J. Larcombe and R. Curtotti. 2017. Fishery status reports 2017, Australian Bureau of Agricultural and Resource Economics and Sciences, Canberra.
- Patterson, H., J. Larcombe, S. Nicol and R. Curtotti. 2018. Fishery status reports 2018, Australian Bureau of Agricultural and Resource Economics and Sciences, Canberra.
- Paul, L. J. 2000. New Zealand fishes. Identification, natural history and fisheries. Reed Publishing, Auckland.
- Paul, L. T., and P. R. Taylor. 1998. A summary of the biology, recreational and commercial landings, and stock assessment of yellow-eyed mullet, *Aldrichetta forsteri* (Cuvier and Valenciennes, 1836) (Mugiloidei: Mugilidae). New Zealand Fisheries Assessment Research Document 98/17, NIWA, Wellington.
- Paxton, J. R., D. F. Hoese, G. R. Allen and J. E. Hanley. Pisces. Petromyzontidae to Carangidae. Zoological Catalogue of Australia, Vol. 7. Australian Government Publishing Service, Canberra, 665 p.
- Pecl, G. T. 2000. Comparative life-history of tropical and temperate Sepioteuthis squids in Australian waters. PhD Thesis, James Cook University, Townsville.

- Pecl, G. T. 2004. The *in situ* relationships between season of hatching, growth and condition in the southern calamary, *Sepioteuthis australis*. *Marine and Freshwater Research* 55:429–438.
- Pecl, G. T., and N. A. Moltschaniwsky. 2006. Life history of a short-lived squid (*Sepioteuthis australis*): resource allocation as a function of size, growth, maturation, and hatching season. *ICES Journal of Marine Science* 63:995–1004.
- Pecl, G. T., N. A. Moltschaniwskyj, S. R. Tracey, and A. R. Jordan. 2004. Inter-annual plasticity of squid life-history and population structure: ecological and management implications. *Oecologia (Berlin)* 139:515–524.
- Pecl, G. T., S. R. Tracey, J. M. Semmens, and G. D. Jackson. 2006. Use of acoustic telemetry for spatial management of *Sepioteuthis australis*, a highly mobile inshore squid species. *Marine Ecology Progress Series* 328:1–15.
- Reid J (2018). Decadal changes in the population dynamics and stock structure of southern sea garfish (*Hyporhamphus melanochir*) within Tasmania. Honours thesis, University of Tasmania.
- Richardson, B. J. 1982. Geographical distribution of electrophoretically detected protein variation in Australian commercial fishes. II: Jackass morwong, *Cheilodactylus macropterus* (Bloch and Schneider). *Australian Journal of Marine and Freshwater Research*, 33: 927–931.
- Robinson, N., A. Skinner, L. Sethuraman, H. McPartlan, N. Murray, I. Knuckey, D. C. Smith, J. Hindell, and S. Talman. 2008. Genetic stock structure of blue-eye trevalla (*Hyperoglyphe antarctica*) and warehou (*Seriola lalandi* and *Seriola punctata*) in south-eastern Australian waters. *Marine and Freshwater Research* 59:502–514.
- Rowling, K. R. 1994. Tiger flathead, *Neoplatycephalus richardsoni*. Pages 124-136 in R. D. J. Tilzey, editor. *The South East Fishery: a scientific review with reference to quota management*. Bureau of Resource Sciences, Australian Government Print Service, Canberra.
- Ruwald, F. P. 1992. Larval feeding trials with striped trumpeter, *Latris lineata*. In: D. A. Hancock, editor. *Larval biology*. Australian Society for Fish Biology Workshop, Hobart, 20 August 1991. Bureau of Rural Resources Proceedings, AGPS, Canberra.
- Ruwald, F. P., L. D. Searle, and L. A. Oates. 1991. A preliminary investigation into the spawning and larval rearing of striped trumpeter, *Latris lineata*. Technical Report 44. Department of Primary Industries, Tasmania.
- Sahlqvist, P., and M. Skirtun. 2011. Southern Squid Fishery. In: J. Woodhams, I. Stobutzki, S. Vieira, R. Curtotti, and G. Begg, editors. *Fisheries Status Reports 2010: status of fish stocks and fisheries managed by the Australian Government*. Australian Bureau of Agriculture and Resource Economics and Sciences, Canberra.
- Scott, T. D., C. J. M. Glover, and R. V. Southcott. 1974. *The marine and freshwater fishes of South Australia*, 2nd ed. Government Printer, South Australia.
- SFAC. 2015. Scalegfish fishery advisory committee (SFAC) minutes No. 59, 26 November 2015. Page 38 Department of Primary Industries, Water and the Environment. Hobart.
- Shepherd, S. A., and P. S. Clarkson. 2001. Diet, feeding behaviour, activity and predation of the temperate blue-throated wrasse, *Notolabrus tetricus*. *Marine and Freshwater Research* 52:311–322.
- Smith, D. C. 1994. Blue warehou. Page 360 in R. D. J. Tilzey, editor. *The South East Fishery: a scientific review with particular reference to quota management*. Bureau of Resource Sciences, Parkes.
- Smith, D. C., I. Montgomery, K. P. Sivakumaran, K. Krusic-Golub, K. Smith, and R. Hodge. 2003. The fisheries biology of bluethroat wrasse (*Notolabrus tetricus*) in Victorian waters. FRDC Report No. 97/128. Marine and Freshwater Institute, Victoria, Australia.
- Smith, H. K. 1983. Fishery and Biology of *Nototodarus gouldi* (McCoy, 1888) in western Bass Strait. *Memoirs of the National Museum Victoria* 44:285–290.
- Smith, A. D. M., T. M. Ward, F. Hurtado, N. Klaer, E. Fulton, and A. E. Punt. 2015a. Review and update of harvest strategy settings for the Commonwealth Small Pelagic Fishery: Single species and ecosystem considerations. FRDC Project No. 2013/028, Final report.
- Smith, T. M., C. P. Green, and C. D. H. Sherman. 2015b. Patterns of connectivity and population structure of the southern calamary *Sepioteuthis australis* in southern Australia. *Marine and Freshwater Research* 66:942–947.

- St Hill, J. L. 1996. Aspects of the biology of southern sea garfish, *Hyporhamphus malanochir*, in Tasmanian waters. University of Tasmania, Hobart.
- Stark, K. E. 2008. Ecology of the arrow squid (*Nototodarus gouldi*) in southeastern Australian waters: a multi-scale investigation of spatial and temporal variability. PhD Thesis, University of Tasmania, Hobart.
- Steer, M. A., N. A. Moltschaniwskyj, and F. C. Gowland. 2002. Temporal variability in embryonic development and mortality in the southern calamary *Sepioteuthis australis*: a field assessment. Marine Ecology Progress Series 243:143–150.
- Steer, M. A., N. A. Moltschaniwskyj, D. S. Nichols, and M. Miller. 2004. The role of temperature and maternal ration in embryo survival: using the dumpling squid *Euprymna tasmanica* as a model. Journal of Experimental Marine Biology and Ecology 307:73–89.
- Stevens, J. D., and H. F. Hansfeld. 1982. Age determination and mortality estimates on an unexploited population of Jack mackerel *Trachurus declivis* (Jenyns, 1841) from southeast Australia. Report No 148, CSIRO Marine Laboratories, Cronulla.
- Stewart, J., G. Ballinger, and D. Ferrell. 2010. Review of the biology and fishery for Australian sardines (*Sardinops sagax*) in New South Wales - 2010. Industry and Investment NSW-Fisheries Research Report Series No. 26.
- Stewart, J., J. Hughes, J. McAllister, J. M. Lyle, and M. MacDonald. 2011. Australian salmon (*Arripis trutta*): population structure, reproduction, diet and composition of commercial and recreational catches. Fisheries Final Report Series No. 129. Industry & Investment NSW.
- Stuart-Smith, R. D., N. S. Barrett, C. M. Crawford, S. D. Frusher, D. G. Stevenson, and G. J. Edgar. 2008. Spatial patterns in impacts of fishing on temperate rocky reefs: Are fish abundance and mean size related to proximity to fisher access points? Journal of Experimental Marine Biology and Ecology 365:116–125.
- Sutton, C. P., D. J. MacGibbon, and D. W. Stevens. 2010. Age and growth of greenback flounder (*Rhombosolea tapirina*) from southern New Zealand. New Zealand Fisheries Assessment Report 2010/48. NIWA, Wellington.
- Thresher, R. E., C. H. Proctor, J. S. Gunn, and I. R. Harrowfield. 1994. An evaluation of electronprobe microanalysis of otoliths for stock identification of nursery areas in a southern temperate groundfish, *Nemadactylus macropterus* (Cheilodactylidae). Fishery Bulletin 92:817–840.
- Tilzey, R. D. J. 1994. The South East fishery: a scientific review with particular reference to quota management. Bureau of Resources Sciences, Australia.
- Tilzey, R. D. J., M. Zann-Schuster, N. L. Klaer, and M. J. Williams. 1990. The South East Trawl Fishery: biological synopses and catch distributions for seven major commercial fish species.
- Tracey, S. R., and J. M. Lyle. 2005. Age validation, growth modeling and mortality estimates for striped trumpeter (*Latris lineata*) from south-eastern Australia: making the most of patchy data. Fishery Bulletin 103:169–182.
- Tracey, S. R., J. M. Lyle, G. Ewing, K. Hartmann, and A. Mapleston. 2013. Offshore recreational fishing in Tasmania 2011/12. Institute for Marine and Antarctic Studies, Hobart.
- Tracey, S. R., A. J. Smolenski, and J. M. Lyle. 2007a. Genetic structuring of *Latris lineata* at localized and transoceanic scales. Marine Biology 152:119–128.
- Tracey, S. R., J. M. Lyle, and M. Haddon. 2007b. Reproductive biology and per-recruit analyses of striped trumpeter (*Latris lineata*) from Tasmania, Australia: implications for management. Fisheries Research 84:358–367.
- Triantafillos, L. 2004. Effects of genetic and environmental factors on growth of southern calamary, *Sepioteuthis australis*, from southern Australia and northern New Zealand. Marine and Freshwater Research 55:439–446.
- Triantafillos, L., and M. Adams. 2001. Allozyme analysis reveals a complex population structure in the southern calamary *Sepioteuthis australis* from Australia and New Zealand. Marine Ecology Progress Series 212:193–209.

- Triantafillos, L., G. D. Jackson, M. Adams, and B. McGrath Steer. 2004. An allozyme investigation of the stock structure of arrow squid *Nototodarus gouldi* (Cephalopoda: Ommastrephidae) from Australia. *ICES Journal of Marine Science* 61:829–835.
- Uozumi, Y. 1998. Fishery biology of arrow squids, *Nototodarus gouldi* and *N. sloanii* in New Zealand Waters. *Bulletin of the National Research Institute of Far Seas Fisheries* 35:1–111.
- van den Enden, T., R. W. G. White, and N. G. Elliott. 2000. Genetic variation in the greenback flounder *Rhombosolea tapirina* Günther (Teleostei, Pleuronectidae) and the implications for aquaculture. *Marine and Freshwater Research* 51:23–33.
- Ward, T. M., and Grammer, G. L. 2016. Commonwealth Small Pelagic Fishery Assessment Report 2015. Report to the Australian Fisheries Management Authority. SARDI publication F2010/000270–7, SARDI Research Report Series 900, South Australian Research and Development Institute (Aquatic Sciences), Adelaide.
- Ward, T. M., O. Burnell, A. Ivey, J. Carroll, J. Keane, J. Lyle, and S. Sexton. 2015. Summer spawning patterns and preliminary Daily Egg Production Method survey of Jack Mackerel and Australian Sardine off the East Coast. South Australian Research and Development Institute (Aquatic Sciences).
- Ward, T., A. Moore, J. Andrews, J. Norriss, and J. Stewart. 2017. Australian Sardine *Sardinops sagax*. In: C. Stewardson, J. Andrews, C. Ashby, M. Haddon, K. Hartmann, P. Hone, P. Horvat, S. Mayfield, A. Roelofs, K. Sainsbury, T. Saunders, J. Stewart, I. Stobutzki and B. Wise (eds) 2016. Status of Australian fish stocks reports 2016, Fisheries Research and Development Corporation, Canberra.
- Walsh, A. T., Barrett, N., and Hill, N. 2017. Efficacy of baited remote underwater video systems and bait type in the cool-temperature zone for monitoring 'no-take' marine reserves. *Marine and Freshwater Research* 68: 568–580.
- Wayte, S. E. 2013. Management implications of including a climate-induced recruitment shift in the stock assessment for jackass morwong (*Nemadactylus macropterus*) in southeastern Australia. *Fisheries Research* 142:47–55.
- Webb, B. F. 1976. Aspects of the biology of jack mackerel *Trachurus declivis* (Jenyns) from south east Australian waters. *Tasmanian Fisheries Research* 10:1–17.
- Welsford, D. C. 2003. Early life-history, settlement dynamics and growth of the temperate wrasse, *Notolabrus fucicola* (Richardson 1840), on the east coast of Tasmania. PhD thesis. University of Tasmania, Hobart.
- Williams, H., and G. Pullen. 1986. A synopsis of biological data on the jack mackerel *Trachurus declivis* Jenyns. Tasmanian Department of Sea Fisheries Technical Report 10. 34 p.
- Woodhams, J., S. Vieira, and I. Stobutzki. 2013. Fishery status reports 2012. Australian Bureau of Agricultural and Resource Economics and Sciences, Canberra.
- Ziegler, P. E. 2012. Fishing tactics and fleet structure of the small-scale coastal scalefish fishery in Tasmania, Australia. *Fisheries Research* 134-136:52–63.
- Zwolinski, J. P., and D. A. Derner. 2013. Measurements of natural mortality for Pacific sardine (*Sardinops sagax*). *ICES Journal of Marine Science* 70:1408–1415.

# Appendix 1 - Common and scientific names of species from catch returns

Common name	Scientific name	Common name	Scientific name
Alfonsino	<i>Beryx</i> spp.	Pilchard	Fam. Clupeidae
Anchovy	Fam. Engraulidae	Rays bream	Fam. Bramidae
Atlantic salmon	<i>Salmo salar</i>	Redbait	<i>Emmelichthys nitidus</i>
Australian Salmon	<i>Arripis</i> spp.	Red fish	Fam. Berycidae
Barracouta	<i>Thyrsites atun</i>	Red Mullet	<i>Upeneichthys</i> spp.
Boarfish	Fam. Pentacerotidae	Silverfish	Fam. Atherinidae
Bream	<i>Acanthopagrus butcheri</i>	Snapper	<i>Pagrus auratus</i>
Butterfish	Spp unknown	Stargazer	Fam. Uranoscopidae
Cardinal fish	Fam Apogonidae	Sweep	<i>Scorpis</i> spp
Cod deep sea	<i>Mora moro</i>	Tailor	<i>Pomatomus saltatrix</i>
Cod, bearded rock	<i>Pseudophycis barbata</i>	Thetis fish	<i>Neosebastes thetidis</i>
Cod, red	<i>Pseudophycis bachus</i>	Trevalla, white	<i>Seriolaella caerulea</i>
Cod, unspec.	Fam. Moridae	Trevally, silver	<i>Pseudocaranx dentax</i>
Dory, john	<i>Zeus faber</i>	Trout, rainbow	<i>Oncorhynchus mykiss</i>
Dory, king	<i>Cyttus traversi</i>	Trumpeter, bastard	<i>Latridopsis forsteri</i>
Dory, mirror	<i>Zenopsis nebulosus</i>	Trumpeter, striped	<i>Latris lineata</i>
Dory, silver	<i>Cyttus australis</i>	Trumpeter, unspec.	Fam. Latridae
Dory, unspec.	Fam. Zeidae	Warehou, blue	<i>Seriolaella brama</i>
Eel	<i>Conger</i> spp.	Warehou, spotted	<i>Seriolaella punctata</i>
Flathead	Fam Plactycephalidae	Whiptail	Fam. Macrouridae
Flounder	Fam. Pleuronectidae	Whiting	Fam. Sillaginidae
Garfish	<i>Hyporhamphus melanochir</i>	Whiting, King George	<i>Sillaginoides punctata</i>
Gurnard	Fam. Triglidae & Fam. Scorpaenidae	Wrasse	<i>Notolabrus</i> spp.
Gurnard perch	<i>Neosebastes scorpaenoides</i>	<b>'Commonwealth' spp</b>	
Gurnard, red	<i>Chelidonichthys kumu</i>	Blue grenadier	<i>Macruronus novaezelandiae</i>
Hardyheads	Fam. Atherinidae	Gemfish	<i>Rexea solandri</i>
Herring cale	<i>Odax cyanomelas</i>	Hapuka	<i>Polyprion oxygeneios</i>
Kingfish, yellowtail	<i>Seriola lalandi</i>	Oreo	Fam. Oreosomatidae
Knifejaw	<i>Oplegnathus woodwardi</i>	Trevalla, blue eye	<i>Hyperoglyphe antarctica</i>
Latchet	<i>Pterygotrigla polyommata</i>	Tunas	
Leatherjacket	Fam. Monacanthidae	Albacore	<i>Thunnus alalunga</i>
Ling	<i>Genypterus</i> spp.	Skipjack	<i>Katsuwonus pelamis</i>
Luderick	<i>Girella tricuspidata</i>	Southern bluefin	<i>Thunnus maccoyii</i>
Mackerel, blue	<i>Scomber australasicus</i>	Tuna, unspec.	Fam. Scombridae
Mackerel, jack	<i>Trachurus declivis</i>	<b>Sharks</b>	
Marblefish	<i>Aplodactylus arctidens</i>	Shark, angel	<i>Squatina australis</i>
Morwong, banded	<i>Cheilodactylus spectabilis</i>	Shark, blue whaler	<i>Prionace glauca</i>
Morwong, blue	<i>Nemadactylus valenciennesi</i>	Shark, bronze whaler	<i>Carcharhinus brachyurus</i>
Morwong, dusky	Fam. Cheilodactylidae	Shark, elephant	<i>Callorhynchus milii</i>
Morwong, grey	<i>Nemadactylus douglasii</i>	Shark, gummy	<i>Mustelus antarcticus</i>
Morwong, jackass	<i>Nemadactylus macropterus</i>	Shark, saw	<i>Pristophorus</i> spp.
Morwong, red	Fam. Cheilodactylidae	Shark, school	<i>Galeorhinus galeus</i>
Morwong, unspec.	Fam. Cheilodactylidae	Shark, seven-gilled	<i>Notorynchus cepedianus</i>
Mullet	Fam. Mugilidae	Shark, spurdog	Fam. Squalidae
Nannygai	<i>Centroberyx affinis</i>	<b>Cephalopods</b>	
Perch, magpie	<i>Cheilodactylus nigripes</i>	Calamari	<i>Sepioteuthis australis</i>
Perch, ocean	<i>Helicolenus</i> spp.	Cuttlefish	<i>Sepia</i> spp.
Pike, long-finned	<i>Dinolestes lewini</i>	Octopus	<i>Octopus</i> spp.
Snook	<i>Sphyræna novaehollandiae</i>	Squid, Gould's	<i>Nototodarus gouldi</i>

## Appendix 2 - Data restrictions and quality control

There have been a number of administrative changes that have affected the collection of catch and effort data from the fishery. The following restrictions and adjustments have been applied when analysing the data as an attempt to ensure comparability between years, especially when examining trends over time.

### Tasmanian logbook data

#### i) Correction of old logbook landed catch weights

Prior to 1995, catch returns were reported as monthly summaries of landings. With the introduction of a revised logbook in 1995, catch and effort was recorded on a daily basis for each method used. Since catch data reported in the old general fishing return represent landed catch, it has been assumed to represent processed weights. For example, where a fish is gilled and gutted, the reported landed weight will be the gilled and gutted and not whole weight. By contrast, in the revised logbook all catches are reported in terms of weight and product form (whole, gilled and gutted, trunk, fillet, bait or live). If a catch of a species is reported as gilled and gutted then the equivalent whole weight can be estimated by applying a standard conversion factor<sup>3</sup>.

Without correcting for product form, old logbook and revised logbook catch weights are not strictly compatible. In an attempt to correct for this and provide a 'best estimate', a correction factor was calculated using catch data from the revised logbook and applied to catches reported in the old logbook. A species based ratio of the sum of estimated whole weights (adjusted for product form) to the sum of reported catch weights was used as the correction factor (Lennon 1998).

#### ii) Effort Problems

Records where effort (based on gear units, Table 2.1) was zero or null, or appeared to be recorded incorrectly (implausible), were flagged. The catch was included in catch summaries but the records were not included in gear unit effort and catch rate calculations. These records were, however, used in calculating days fished and daily catches.

#### iii) Vessel restrictions

In all analyses of catch and effort, catches from six vessels (four Victorian based and two Tasmanian based) have been excluded. These vessels were known to have fished consistently in Commonwealth waters and their catches of species such as Blue Warehou and ling tended to significantly distort catch trends. In fact, all four Victorian vessels and one of the Tasmanian vessels ceased reporting on the General Fishing Returns in 1994. With the introduction of the South East Fishery Non-Trawl logbook (GN01) in 1997, the remaining Tasmanian vessel ceased reporting fishing activity in the Tasmanian logbook.

### Commonwealth logbook data:

Commonwealth logbook data from Australian Fisheries Management Authority was included in the analyses so that the assessment reflected all catches from Tasmanian waters.

#### (i) Area restrictions

Commonwealth logbook records were only included if the catch was taken in fishing blocks adjacent to Tasmania and the maximum depth of the fishing operation was less than 200 m. These conditions were applied to all records except where Striped or Bastard Trumpeter were

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<sup>3</sup> Conversion factors to whole weights are 1.00 for whole, live or bait; 2.50 for fillet; 1.50 for trunk; and 1.18 for gilled and gutted.

caught. All records that included catches of these species were included for analysis, because these species are managed under Tasmanian jurisdiction in all waters adjacent to Tasmania.

Fishing blocks adjacent to land and used in the analyses (refer Fig. A1) include:

3C2, 3D1, 3F1, 3F2, 3G1, 3G2, 3C4, 3D3, 3F4, 3G3, 3G4, 3H3, 3H4, 4C2, 4D1, 4D2, 4E1, 4G2, 4H1, 4H2, 4D4, 4E3, 4E4, 4F4, 4G3, 4G4, 4H3, 4H4, 5D2, 5E2, 5F1, 5F2, 5H1, 5D4, 5E3, 5H3, 6E1, 6H1, 6E3, 6G4, 6H3, 7E1, 7E2, 7G1, 7G2, 7H1, 7E4, 7F3, 7F4, 7G3.



**Fig. A1** Numbers for fishing blocks used in calculation of catch figures.

(ii) Duplicate records

A number of records in Commonwealth logbooks had matching records (fisher, date, gear type) in the Tasmanian database. Such records were examined individually and decisions made as to whether it was more appropriate to keep the Tasmanian record, the Commonwealth record or both. In most situations the Tasmanian logbook entry was kept and the Commonwealth record excluded. The only exceptions were records with extra information in the Commonwealth record, e.g. catch of a Commonwealth species that was not recorded in the Tasmanian logbook.

## Appendix 3 - Annual Tasmanian scalefish production

**Table A.1.** Annual Tasmanian scalefish, small pelagic fish, cephalopod and shark production (whole weight in tonnes) by financial year since 1995/96 based on General Fishing Returns and Commonwealth (GN01, GN01A and SSJF) logbook returns. Note data includes Commonwealth catches of Striped Trumpeter, Bastard Trumpeter and Gould's Squid in State waters.

Species	95/96	96/97	97/98	98/99	99/00	00/01	01/02	02/03	03/04	04/05	05/06	06/07	07/08	08/09	09/10	10/11	11/12	12/13	13/14	14/15	15/16	16/17	17/18
<b>Key Scalefish species (excl. small pelagics)</b>																							
Australian Salmon	413.2	287.3	475.7	384.7	363.7	485.0	462.1	407.2	167.2	336.5	254.2	115.0	256.1	338.8	372.3	203.5	189.4	331.3	65.6	42.2	89.3	18.9	76.1
Barracouta	19.3	53.8	65.2	27.6	25.0	15.1	136.0	67.5	87.5	101.0	60.1	26.6	13.3	13.3	7.6	5.0	4.0	1.1	1.1	1.7	0.4	1.4	0.9
Boarfish	7.3	10.0	6.2	3.2	2.5	3.6	5.5	3.6	4.3	3.6	5.0	5.2	4.7	2.6	2.7	1.9	3.4	2.1	1.0	0.6	0.7	0.7	1.1
Cod	18.6	12.8	9.4	9.6	8.8	3.7	3.0	2.3	2.1	1.6	2.0	2.6	2.3	3.3	2.6	2.8	2.4	2.0	2.0	2.0	1.5	1.3	0.9
Flathead, sand	13.7	12.7	13.0	10.1	12.5	8.2	13.1	10.8	10.6	13.9	12.6	12.0	11.5	13.0	9.2	6.7	7.5	5.5	6.8	8.1	2.7	6.4	3.5
Flathead, tiger	34.1	31.3	44.5	37.1	44.4	53.0	35.9	27.2	17.9	58.8	75.7	44.8	62.0	37.8	66.3	47.6	52.7	31.2	20.2	23.5	64.4	74.0	39.4
Flounder	33.4	29.4	29.7	25.2	18.6	12.3	13.0	10.9	14.9	14.7	10.9	13.0	7.8	5.1	5.2	5.2	4.0	2.0	2.1	1.5	1.0	3.3	3.9
Garfish	56.2	91.6	83.0	101.7	91.7	81.4	87.8	92.5	66.2	85.5	89.3	50.0	31.0	63.0	49.3	43.2	53.0	51.5	37.9	33.8	21.9	16.4	8.9
Gurnard	13.5	10.4	9.1	7.0	9.6	7.4	5.3	9.7	6.8	6.1	5.1	5.7	4.7	2.6	1.5	2.1	1.2	1.1	0.6	1.9	2.1	2.7	1.8
Leatherjacket	14.5	12.6	13.3	12.9	16.6	16.7	16.6	13.7	14.8	10.4	8.5	8.8	5.3	5.5	3.0	2.9	2.2	2.4	2.9	2.1	1.3	2.6	2.6
Ling	15.0	13.3	8.3	4.3	1.8	1.2	0.9	0.4	0.8	0.7	0.4	0.4	0.4	0.1	0.1	0.1	0.1	0.1	1.2	0.0	0.1	0.1	0.1
Marblefish	3.5	5.6	3.0	2.6	4.2	4.0	4.4	3.1	0.6	1.1	0.5	2.2	2.3	1.1	0.5	0.2	0.2	0.3	0.2	0.3	0.3	0.3	0.0
Morwong, banded	85.8	78.0	72.6	42.4	34.2	39.0	53.7	56.0	46.4	45.6	54.4	50.3	52.6	37.1	44.6	40.9	40.3	37.9	34.1	30.1	32.9	34.0	30.3
Morwong, jackass	27.1	18.7	33.2	17.5	15.9	13.1	14.8	14.7	16.6	17.5	13.1	11.7	4.6	5.3	5.9	3.2	3.1	1.5	1.0	0.8	3.2	1.6	3.3
Morwong, other	5.4	7.4	7.4	6.3	1.4	0.6	1.4	1.9	1.2	1.8	1.3	1.3	2.5	1.4	1.2	0.9	0.7	0.7	0.6	0.7	0.3	0.6	0.3
Mullet	1.0	1.7	1.7	2.2	4.9	4.8	2.5	4.0	4.3	2.4	3.2	2.0	0.1	1.4	1.8	2.1	0.5	4.4	0.5	0.8	2.4	0.4	0.3
Other	26.1	24.1	24.6	17.0	13.6	10.2	10.5	29.8	20.9	24.7	26.0	13.6	9.1	31.6	4.6	1.5	2.3	1.8	1.9	1.8	0.9	2.7	3.0
Snook	13.7	15.2	17.7	3.2	4.1	5.9	6.6	6.6	3.7	2.2	2.9	6.7	7.0	8.7	7.9	7.5	6.7	6.3	9.1	9.0	2.6	9.4	5.9
Trevally	8.4	6.0	5.4	6.5	2.7	1.6	4.7	5.9	3.4	3.7	6.3	3.6	8.8	4.5	3.8	1.9	2.1	5.4	4.3	5.7	2.8	3.6	3.3
Trumpeter, bastard	60.1	51.8	40.7	47.7	36.4	26.1	23.9	21.0	23.2	18.5	23.4	21.3	19.1	16.7	10.5	9.8	9.6	9.5	8.3	6.5	8.4	6.4	4.3
Trumpeter, striped	58.3	79.4	78.1	99.0	95.0	45.5	39.9	36.6	36.9	23.9	19.0	18.7	12.2	10.7	10.8	19.7	20.9	17.3	10.5	13.0	7.1	12.1	14.1
Trumpeter, unspec.	0.0	0.1	0.6	3.3	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.1	0.0	0.1	0.1	0.0	0.1	0.1	0.3	0.2	0.0	0.0
Warehou, blue	82.3	128.4	187.6	272.2	187.1	34.2	66.4	49.3	27.6	19.1	20.0	29.3	25.3	26.8	37.5	10.7	3.8	8.5	5.8	2.8	7.4	7.6	12.6
Warehou, other	14.6	15.6	4.2	1.0	0.0	0.0	0.1	0.2	0.1	0.8	0.1	0.0	0.1	0.6	0.2	0.0	0.0	0.2	0.0	0.1	0.3	0.4	1.2
Whiting	1.4	0.1	0.0	23.3	9.6	36.5	39.6	35.9	50.9	31.6	2.3	38.1	31.4	32.5	26.7	34.2	15.5	13.8	36.6	1.9	20.7	26.0	16.1
Wrasse	83.4	110.1	100.0	90.7	85.5	88.4	92.3	72.0	75.1	100.1	92.9	112.9	87.6	68.1	72.0	72.7	68.0	64.2	65.1	81.8	72.7	79.1	83.8
<b>Total key scalefish</b>	<b>1110</b>	<b>1107</b>	<b>1334</b>	<b>1258</b>	<b>1090</b>	<b>997.3</b>	<b>1140</b>	<b>982.6</b>	<b>704.2</b>	<b>925.9</b>	<b>789.3</b>	<b>596.0</b>	<b>662.0</b>	<b>731.5</b>	<b>748.2</b>	<b>526.3</b>	<b>493.7</b>	<b>602.3</b>	<b>319.6</b>	<b>273.0</b>	<b>347.7</b>	<b>311.9</b>	<b>317.7</b>

Table A.1 Continued. Whole weight in tonnes by financial year

Species	95/96	96/97	97/98	98/99	99/00	00/01	01/02	02/03	03/04	04/05	05/06	06/07	07/08	08/09	09/10	10/11	11/12	12/13	13/14	14/15	15/16	16/17	17/18
<b>Small pelagics</b>																							
Australian sardine	6.6	4.3	15.4	2.8	1.7	3.2	0.7	0.0	0.3	0.8	0.0	0.0	13.2	14.5	0.4	0.0	0.0	0.0	0.1	0.0	0.0	33.3	0.1
Mackerel, jack	26.2	19.3	19.7	59.8	14.7	9.1	19.4	19.4	41.1	12.8	6.8	2.6	202.8	919.7	917.0	35.7	56.4	0.2	0.4	5.5	1.0	0.1	2.0
Mackerel, other	2.0	1.3	1.0	0.5	2.1	0.1	0.0	0.1	0.0	0.5	0.5	0.2	10.3	0.2	0.3	0.8	0.1	1.9	4.2	1.1	0.2	2.8	0.5
Redbait	0.1	0.0	0.0	4.0	0.0	0.0	0.0	0.0	3.4	1.0	1.4	0.3	300.1	521.4	121.6	15.0	0.1	0.1	0.0	0.0	0.0	0.0	0.0
<b>Total small pelagics</b>	<b>34.9</b>	<b>24.8</b>	<b>36.1</b>	<b>67.0</b>	<b>18.5</b>	<b>12.4</b>	<b>20.1</b>	<b>19.5</b>	<b>44.8</b>	<b>15.2</b>	<b>8.7</b>	<b>3.1</b>	<b>526.4</b>	<b>1456</b>	<b>1039</b>	<b>51.6</b>	<b>56.7</b>	<b>2.1</b>	<b>4.6</b>	<b>6.6</b>	<b>1.2</b>	<b>36.2</b>	<b>2.7</b>
<b>Cephalopods</b>																							
Calamari, southern	33.0	19.0	26.6	94.4	87.4	78.0	105.2	108.8	86.8	114.2	44.6	85.4	89.0	78.6	51.1	54.9	50.8	63.9	67.8	75.9	106.2	122.6	60.6
Cuttlefish	0.2	0.3	0.2	0.0	0.0	0.0	0.7	2.4	1.0	0.2	0.4	0.1	0.3	0.3	0.1	0.1	0.1	0.2	0.1	0.1	0.5	0.3	0.2
Octopus, pale	58.7	30.2	34.1	61.2	43.2	43.0	29.0	43.4	55.8	80.9	100.3	97.4	92.2	64.7	129.8	57.0	110.5	81.5	111.3	74.9	89.9	56.1	102.5
Octopus, other <sup>4</sup>													3.0	2.2	2.1	5.9	3.8	4.5	8.3	4.7	7.5	19.2	6.7
Squid, Gould's	5.7	7.8	12.9	79.7	481.3	39.7	2.4	1.9	2.1	2.6	1.8	687.7	45.9	45.5	121.3	131.2	516.6	1071.8	0.0	31.4	416.8	175.6	528.0
<b>Total cephalopods</b>	<b>97.7</b>	<b>57.3</b>	<b>73.8</b>	<b>235.3</b>	<b>612.0</b>	<b>160.7</b>	<b>137.2</b>	<b>156.6</b>	<b>145.8</b>	<b>198.0</b>	<b>147.1</b>	<b>870.8</b>	<b>230.3</b>	<b>191.3</b>	<b>304.3</b>	<b>249.0</b>	<b>681.7</b>	<b>1221.9</b>	<b>187.6</b>	<b>187.0</b>	<b>620.8</b>	<b>373.8</b>	<b>698.0</b>
<b>Sharks<sup>5</sup></b>																							
Elephant shark	58.0	48.9	21.4	14.7	17.0	16.7	18.4	16.5	10.2	7.6	5.7	9.0	1.9	1.5	2.4	1.3	2.7	1.9	1.4	0.6	0.2	1.8	1.2
Gummy shark	750.5	543.8	348.6	113.4	109.7	53.9	23.5	14.2	24.7	41.6	12.4	13.6	13.8	9.8	9.8	9.3	7.5	7.9	6.0	7.6	8.2	11.1	9.1
Draughtboard shark	0.0	0.0	0.0	0.0	0.0	0.0	1.7	0.7	1.0	0.8	1.3	1.2	0.4	0.3	0.2	0.3	0.0	0.1	0.0	0.0	0.0	0.0	0.0
Sawshark	127.4	74.4	29.2	6.8	3.4	12.3	21.4	20.4	20.6	23.5	5.9	3.4	0.3	0.1	0.1	0.1	0.4	0.0	0.0	0.0	0.0	0.0	0.0
School shark	252.1	171.5	71.7	31.5	11.3	1.7	2.2	1.4	7.0	2.6	0.6	1.8	1.0	0.7	1.8	1.4	1.9	2.1	1.1	1.4	1.1	1.7	2.0
Seven-gilled shark	6.1	4.9	6.1	1.9	10.3	16.3	18.8	7.4	11.5	8.4	3.8	3.9	0.5	2.3	1.1	1.4	1.1	0.8	0.7	1.0	1.1	0.4	0.1
Other shark	26.4	16.1	11.3	6.8	6.5	4.8	5.8	3.6	3.2	1.1	0.6	2.3	0.9	0.7	0.3	0.9	0.6	0.6	0.7	0.9	0.8	0.9	1.8
<b>Total sharks</b>	<b>1220</b>	<b>859.6</b>	<b>488.2</b>	<b>175.1</b>	<b>158.3</b>	<b>105.7</b>	<b>91.9</b>	<b>64.3</b>	<b>78.3</b>	<b>85.6</b>	<b>30.4</b>	<b>35.1</b>	<b>18.8</b>	<b>15.4</b>	<b>15.7</b>	<b>14.8</b>	<b>14.2</b>	<b>13.3</b>	<b>10.0</b>	<b>11.5</b>	<b>11.5</b>	<b>15.9</b>	<b>14.1</b>

<sup>4</sup> Includes Gloomy Octopus (*Octopus tetricus*), Maori octopus (*Macroctopus maorum*), Southern Octopus (*O. australis*) and unspecified octopus. Excludes bycatch in rock lobster pots.

<sup>5</sup> Since 2001/02, shark catches have been reported in Commonwealth logbooks. Tasmania has jurisdiction of all shark species inside 3 nm except gummy and school shark, and fishers are on bycatch possession limits for all species.



# Appendix 4 - Annual catch, effort and number of vessels by fishing methods

**Table A.2** Annual catch, effort and number of vessels by fishing methods. # Effort units are defined in Table 1.1. \* Catch data not shown where five or fewer vessels involved.

Gear	Year	Catch (t)	Effort #	Days fished	Vessels
Auto squid jig	95/96	*	3000	2	1
	96/97	*	14560	13	3
	97/98	*	3040	8	1
	98/99	*	133728	66	3
	99/00	387.7	899711	336	13
	00/01	28.1	280200	115	10
	01/02	*	10220	7	3
	02/03	*	9000	5	1
	03/04	*	3000	2	1
	04/05	*	500	1	1
	05/06	-	-	-	-
	06/07	681.9	395762	155	9
	07/08	*	741	12	2
	08/09	*	671	12	2
	09/10	*	3339	69	5
	10/11	130.8	422	79	14
	11/12	515.9	3060	187	13
	12/13	1069.6	11557	249	8
	13/14	*	90	3	1
14/15	*	551	33	2	
15/16	415.7	7066	146	7	
16/17	*	1913	87	5	
17/18	527.0	6089	160	6	
Beach seine	95/96	469.2	1086	559	53
	96/97	351.7	1355	685	50
	97/98	520.9	1206	582	44
	98/99	441.7	872	398	40
	99/00	422.9	901	430	33
	00/01	529.0	789	373	31
	01/02	572.2	1070	495	30
	02/03	490.7	1063	511	35
	03/04	238.1	1282	458	31
	04/05	397.0	975	368	27
	05/06	308.4	653	304	25
	06/07	140.6	528	234	25
	07/08	266.3	630	263	17
	08/09	382.6	597	282	12
	09/10	410.3	473	251	15
	10/11	225.2	359	200	13
	11/12	203.7	380	184	13
	12/13	334.3	321	163	14
	13/14	77.1	287	174	15
14/15	61.9	325	202	17	
15/16	88.5	229	150	14	
16/17	18.6	203	125	10	
17/18	82.6	212	135	8	
Drop-line	95/96	19.9	438	158	31
	96/97	30.0	433	203	27
	97/98	24.7	539	222	42
	98/99	31.8	666	309	38
	99/00	30.8	385	291	48
	00/01	15.8	382	248	36
	01/02	12.8	220	258	35
	02/03	18.8	264	350	43
	03/04	19.4	378	281	51
	04/05	14.1	351	219	31
	05/06	9.3	185	204	33
	06/07	7.1	259	137	28
	07/08	3.0	39	55	19
	08/09	3.1	45	41	18
	09/10	2.4	35	25	7
	10/11	5.9	100	58	18
	11/12	9.2	125	92	18
	12/13	8.6	159	86	17
	13/14	4.3	126	47	9
14/15	3.4	110	44	8	
15/16	2.4	38	38	10	
16/17	8.1	149	103	15	
17/18	7.1	197	62	11	

Table A.2 (continued)

Gear	Year	Catch (t)	Effort #	Days fished	Vessels
Dip-net	95/96	*	320	83	5
	96/97	24.2	1518	364	10
	97/98	37.9	1903	449	21
	98/99	43.6	2784	579	29
	99/00	29.8	2319	505	35
	00/01	22.8	1430	371	27
	01/02	24.8	1561	387	27
	02/03	18.7	1259	337	20
	03/04	25.6	1557	374	19
	04/05	27.4	1521	305	16
	05/06	39.1	2167	376	18
	06/07	22.6	1308	244	18
	07/08	15.2	1057	227	17
	08/09	16.4	1229	287	14
	09/10	18.0	1376	270	15
	10/11	16.0	958	227	11
	11/12	22.6	1192	222	12
	12/13	23.3	1209	174	10
	13/14	16.9	1115	147	10
	14/15	12.9	752	118	14
15/16	13.4	638	95	11	
16/17	8.1	425	71	8	
17/18	4.3	281	67	8	
Danish seine	95/96	*	474	163	2
	96/97	*	360	116	1
	97/98	*	456	133	1
	98/99	*	375	94	1
	99/00	*	515	139	2
	00/01	*	589	152	2
	01/02	*	491	145	2
	02/03	*	354	129	3
	03/04	*	278	127	2
	04/05	*	282	108	2
	05/06	*	418	132	3
	06/07	*	475	157	3
	07/08	*	482	162	3
	08/09	*	387	134	3
	09/10	*	557	181	4
	10/11	*	426	155	5
	11/12	*	262	94	3
	12/13	*	327	112	5
	13/14	*	382	116	3
	14/15	*	251	56	3
15/16	*	412	120	4	
16/17	*	461	132	2	
17/18	*	356	107	2	
Fish trap	95/96	41.8	8264	1401	66
	96/97	57.2	10710	1796	66
	97/98	49.9	9870	1875	71
	98/99	53.7	10657	1559	56
	99/00	56.1	11030	1637	62
	00/01	54.3	9356	1548	68
	01/02	49.0	6098	1278	62
	02/03	38.2	6177	1246	58
	03/04	48.0	6308	1414	58
	04/05	46.7	7409	1222	54
	05/06	44.6	12302	1421	54
	06/07	44.2	11001	1328	47
	07/08	27.9	9761	916	44
	08/09	13.8	5316	568	29
	09/10	9.5	2257	341	23
	10/11	9.0	2712	410	23
	11/12	8.6	2934	352	22
	12/13	6.9	1585	188	17
	13/14	8.7	1666	271	21
	14/15	6.3	2042	213	25
15/16	3.4	1451	164	14	
16/17	10.5	8886	307	20	
17/18	7.6	2327	245	19	

Table A.2 (continued)

Gear	Year	Catch (t)	Effort #	Days fished	Vessels
Gillnet	95/96	347.1	223553	5437	257
	96/97	382.3	231140	5185	232
	97/98	446.3	231412	5249	216
	98/99	493.3	166505	4689	204
	99/00	361.7	152144	4169	203
	00/01	173.5	86838	3187	186
	01/02	196.0	71109	3303	180
	02/03	231.3	85628	3395	168
	03/04	190.2	69189	2904	160
	04/05	154.4	53965	2491	137
	05/06	170.2	51591	2402	123
	06/07	170.0	56742	2543	132
	07/08	161.3	59308	2425	115
	08/09	127.1	46668	1977	95
	09/10	128.7	47049	2022	95
	10/11	96.8	45324	1775	86
	11/12	100.6	40060	1607	85
	12/13	103.4	40564	1591	72
13/14	99.3	37824	1456	66	
14/15	83.4	32352	1285	58	
15/16	92.6	31449	1315	57	
16/17	82.7	31048	1220	57	
17/18	77.1	29833	1006	49	
Hand collection	95/96	7.0	1198	433	6
	96/97	*	298	154	5
	97/98	*	261	214	4
	98/99	*	918	267	5
	99/00	*	947	228	3
	00/01	*	1008	272	3
	01/02	*	1185	317	4
	02/03	*	625	160	3
	03/04	*	678	184	3
	04/05	*	400	125	3
	05/06	*	226	97	4
	06/07	*	439	200	5
	07/08	8.4	571	201	10
	08/09	*	122	52	4
	09/10	*	124	55	4
	10/11	5.3	403	135	9
	11/12	*	177	99	4
	12/13	*	244	117	5
13/14	4.0	352	134	6	
14/15	*	62	29	5	
15/16	*	77	42	2	
16/17	2.0	129	52	6	
17/18	*	131	100	3	
Hand-line	95/96	74.3	16964	1612	147
	96/97	94.3	21542	1893	135
	97/98	97.5	21076	1702	145
	98/99	88.2	17668	1278	127
	99/00	87.9	16688	1439	134
	00/01	74.2	13585	1541	130
	01/02	87.3	15527	1603	138
	02/03	72.2	15025	1552	125
	03/04	76.4	15610	1411	127
	04/05	100.5	19953	1803	123
	05/06	82.6	20247	1884	116
	06/07	107.3	22745	2139	128
	07/08	92.0	19985	2032	119
	08/09	80.5	17814	1740	93
	09/10	84.1	17284	1603	92
	10/11	91.4	17766	1670	98
	11/12	83.8	16039	1725	98
	12/13	77.4	12601	1458	97
13/14	78.0	13285	1367	87	
14/15	105.7	18748	1805	94	
15/16	104.4	21874	1886	90	
16/17	84.5	15276	1648	90	
17/18	94.2	17440	1838	89	

Table A.2 (continued)

Gear	Year	Catch (t)	Effort #	Days fished	Vessels
Small mesh net	95/96	38.7	10971	285	19
	96/97	27.0	7965	260	14
	97/98	21.8	7875	246	17
	98/99	31.2	7772	282	14
	99/00	22.7	6232	210	15
	00/01	20.8	8170	256	14
	01/02	24.7	9863	259	11
	02/03	22.9	10297	284	11
	03/04	23.0	7254	228	11
	04/05	15.3	5982	220	13
	05/06	21.7	5890	191	11
	06/07	16.4	7144	202	11
	07/08	15.3	6447	183	10
	08/09	9.5	4817	152	7
	09/10	9.4	4089	134	7
	10/11	7.1	2281	98	8
	11/12	10.9	4127	152	9
	12/13	12.4	4784	189	11
	13/14	10.1	3342	150	11
14/15	9.7	4003	153	9	
15/16	4.8	1267	71	10	
16/17	17.1	4709	114	7	
17/18	7.0	1420	76	7	
Purse seine	95/96	35.18	418	185	11
	96/97	30.42	336	153	10
	97/98	41.79	319	154	7
	98/99	76.90	246	150	8
	99/00	37.02	244	123	10
	00/01	*	224	104	4
	01/02	*	216	91	5
	02/03	*	139	76	4
	03/04	*	68	45	3
	04/05	*	130	70	5
	05/06	*	122	60	4
	06/07	*	86	41	4
	07/08	*	117	121	5
	08/09	*	195	153	3
	09/10	*	170	99	4
	10/11	*	33	32	3
	11/12	*	63	60	3
	12/13	*	106	101	4
	13/14	*	27	27	3
14/15	15.69	51	43	6	
15/16	*	26	22	3	
16/17	*	89	70	5	
17/18	*	60	43	4	
Squid jig	95/96	10.2	5389	125	23
	96/97	5.7	640	77	14
	97/98	15.2	4381	211	18
	98/99	89.8	10200	613	53
	99/00	151.4	39240	989	64
	00/01	67.8	13173	793	53
	01/02	85.6	12544	925	65
	02/03	91.8	19220	1228	68
	03/04	69.8	15554	1223	73
	04/05	104.8	22362	1424	79
	05/06	35.4	11223	767	59
	06/07	74.4	14114	1204	67
	07/08	77.7	16220	1406	52
	08/09	71.0	15127	1079	35
	09/10	45.5	14505	926	37
	10/11	49.3	12932	874	28
	11/12	42.1	11378	772	28
	12/13	51.1	11465	865	32
	13/14	52.2	11160	799	31
14/15	60.4	14532	905	36	
15/16	73.6	16809	1093	44	
16/17	112.6	25100	1433	50	
17/18	49.8	18099	1153	36	

Table A.2 (continued)

Gear	Year	Catch (t)	Effort #	Days fished	Vessels
Spear	95/96	14.1	1403	368	21
	96/97	19.3	1853	464	27
	97/98	16.8	1981	483	40
	98/99	19.8	1812	452	38
	99/00	19.3	2233	475	25
	00/01	14.4	1586	355	22
	01/02	13.1	1296	279	19
	02/03	10.3	1366	247	22
	03/04	10.5	1446	289	22
	04/05	13.5	1609	357	24
	05/06	7.9	1009	271	22
	06/07	15.4	1414	362	20
	07/08	9.8	957	255	21
	08/09	5.6	886	181	14
	09/10	5.0	663	170	15
	10/11	4.0	626	179	16
	11/12	5.0	622	176	16
	12/13	3.3	489	128	17
13/14	3.7	339	109	12	
14/15	2.1	266	88	13	
15/16	1.4	188	49	8	
16/17	3.5	407	101	8	
17/18	5.7	565	149	9	
Troll	95/96	19.6	3497	352	58
	96/97	62.1	9755	600	90
	97/98	76.2	13318	680	83
	98/99	46.1	9307	464	73
	99/00	39.7	6184	421	51
	00/01	36.5	7913	440	49
	01/02	145.1	18669	1007	94
	02/03	84.8	8510	669	74
	03/04	89.9	8995	789	89
	04/05	91.5	6797	758	77
	05/06	63.4	5019	491	54
	06/07	30.3	4664	350	52
	07/08	13.3	2648	172	30
	08/09	20.0	3023	181	27
	09/10	12.3	2542	137	22
	10/11	10.9	2501	110	19
	11/12	10.4	2153	93	12
	12/13	5.1	1622	72	10
13/14	5.3	2078	79	13	
14/15	5.9	2894	113	12	
15/16	2.0	1448	52	7	
16/17	6.0	2456	122	11	
17/18	5.2	2026	100	8	